

1 **ASSESSING THE PERCEPTION AND REALITY OF ARGUMENTS**
2 **AGAINST THERMAL WASTE TREATMENT PLANTS IN TERMS OF**
3 **PROPERTY PRICES**

4

5 **K.J.O. Phillips, P.J. Longhurst, S.T. Wagland***

6

7 Centre for Energy and Resource Technology, School of Applied Sciences, Cranfield

8 University, Cranfield, Bedfordshire, MK43 0AL, UK

9 _____

10 * Corresponding author. Tel.: +44 (0) 1234 750111 extn 2404; Fax: +44 (0) 1234

11 751671

12 *E – mail address:* s.t.wagland@cranfield.ac.uk

1 **Abstract**

2 The thermal processing of waste materials, although considered to be an essential part
3 of waste management, is often sharply contested in the UK. Arguments such as
4 health, depletion of resources, cost, noise, odours, traffic movement and house prices
5 are often cited as reasons against the development of such facilities. This study aims
6 to review the arguments and identify any effect on property prices due to the public
7 perception of the plant. A selection of existing energy from waste (EfW) facilities in
8 the UK, operational for at least 7 years, was selected and property sales data, within 5
9 km of the sites, was acquired and analysed in detail. The locations of the properties
10 were calculated in relation to the plant using GIS software (ArcGIS) and the distances
11 split into 5 zones ranging from 0-5 km from the site. The local property sale prices,
12 normalised against the local house price index, were compared in two time periods,
13 before and after the facility became operational, across each of the 5 zones. In all
14 cases analysed no significant negative effect was observed on property prices at any
15 distance within 5 km from a modern operational incinerator. This indicated that the
16 perceived negative effect of the thermal processing of waste on local property values
17 is negligible.

18

19 **Keywords:** energy from waste; thermal waste treatment; public perception; property
20 prices

21

22

23

24

25

1 **1. Introduction**

2 Thermal waste processing with energy generation (electrical and heat) is seen
3 as a sustainable and effective solution to both waste management and energy
4 generation (Cheng and Hu, 2010; Institution of Mechanical Engineers, 2007; Jamasb
5 and Nepal, 2010; Michaels, 2009; Murphy and McKeogh, 2004; Papageorgiou et al.,
6 2009; Porteous, 2005; Porteous, 1998; Burnley et al., 2011; Haley, 1990). However,
7 these plans are often fiercely contested by anti-incinerator groups and local residents
8 (Jamasb and Nepal, 2010; Porteous, 1998; Achillas et al., 2011; DEFRA, 2007;
9 Furuseth and O'Callaghan, 1991). The main protests against these facilities are
10 related to perceived health risks, the effects on house prices and noise, smell and
11 increased traffic (Achillas et al., 2011; Friends of the Earth, 2000; British Society for
12 Ecological Medicine, 2008). The effect on recycling rates, depletion of resources and
13 the effectiveness of alternative waste management solutions are also cited as reasons
14 against such treatment options (Friends of the Earth, 2000).

15 Previous studies and reports largely agree that the health risks from modern
16 thermal waste processing plants, especially increased cancer incidence related to
17 dioxin emissions, are very small (Institution of Mechanical Engineers, 2007; Porteous,
18 2005; Porteous, 1998; HPA, 2009; Health Protection Scotland, 2009; DEFRA, 2004).
19 Indeed, the reported emissions from modern facilities are significantly lower than the
20 limits prescribed in the Waste Incineration Directive [2000/76/EC], and is considered
21 to be “the most strictly controlled combustion process in the UK” (Porteous, 1998).

22 Energy from Waste (EfW), specifically Mass Burn Incineration (MBI), is
23 widely seen as an essential part of an integrated waste management solution but the
24 primary constraint on the widespread uptake is the negative perceptions of politicians
25 and the public. The justification is put down to the historic memory of “dirty”

1 incinerators of the past, but this study will demonstrate that this is no longer the case
2 with modern incinerators “being the natural companion to practicable recycling”
3 (Porteous, 1998).

4 Incinerators have a chequered history with a large negative hang over from the
5 dirty polluting plant of the 1970s. Opposition groups range from International, such
6 as the Global Anti Incinerator Alliance (GAIA); National, such as the United
7 Kingdom Without Incineration Network (UKWIN) and local groups, such as the
8 Cheshire Anti Incinerator Network (CHAIN), Guildford Anti Incinerator Network
9 (GAIN) and Hatfield Anti Incineration (HAI). Friends of the Earth and Greenpeace
10 are vocal opponents of incineration and actively support local groups through
11 publications such as “How to Win: Campaign Against Incinerators” (Friends of the
12 Earth, 2000). These groups often vary between a genuine concern for the local area
13 and a visceral, often illogical, abhorrence of incineration on ideological grounds.

14

15 **2. Arguments against incineration**

16 There are a number of offered disadvantages of EfW facilities, which
17 include air pollution and health, effects on climate change and the destruction of
18 valuable resources, toxic waste generation, noise and traffic, and poor public
19 perception.

20

21 **2.1. Air pollution and health**

22 Health is a much cited argument against incineration. The public are very
23 conscious of the perceived health risks, for example a study in Greece found that 43%
24 of respondents cited health issues when asked for reasons to protest against a new
25 incinerator (Achillas et al., 2011).

1 Anti-incinerator groups widely cite a report by the British Society for
2 Ecological Medicine (2008), a charity aimed at promoting public health, as reference
3 to the health impacts, which discusses several negative health effects of such facilities
4 through the release of dioxins and particulates. However this report was widely
5 rebuked by the Health Protection Agency (2005) and Enviro (2006) for using
6 “inaccurate and outdated material” (Enviro, 2006) and confusing the issues of health
7 impact and risk, especially with relation to alternatives to incineration.

8 The main thrust of opposition is related to the emissions of dioxins, a family of
9 cancer causing compounds. Many processes produce dioxins and it is important to
10 put the emissions from a modern incinerator in context. Table 1 shows the
11 comparative amounts of dioxins produced by other processes from sinter plants to
12 domestic coal combustion. As can be seen a modern EfW plant produces around half
13 the dioxins as a coal power plant and less than 4 times that of domestic wood
14 combustion. .

15

16 >>>>>Insert Table 1<<<<<<<

17

18 The Institution of Mechanical Engineers describes the dioxin emissions limit
19 for a modern EfW plant as “an equivalent concentration to one third of a sugar lump
20 dissolved in Loch Ness” (Institution of Mechanical Engineers, 2007). Elliot et al.
21 (2000) places the increased risk of cancer within 1 km of an EfW facility to be
22 between 0.53 and 0.78 cases per million. Roberts and Chen (2006) state the overall
23 risk of dying due to emissions within 5.5 km of a facility to be 2.49×10^{-7} or 1 in 4
24 million. This is compared to 10 cases of melanoma per 100,000 (from
25 sunbathing/sunbed use) and 15,000 people a year which die from bowel cancer (which

1 is mainly diet related) (Porteous, 1998) .

2

3 **2.2. Effects on climate change and the destruction of valuable resources.**

4 Energy from Waste proponents often report that incineration of MSW is
5 carbon neutral (Jamasb and Nepal, 2010; Porteous, 1998). This is due to the mainly
6 biodegradable component of the waste and the offsetting of green house gas (GHG)
7 emissions from landfill and fossil fuel based electricity generation. For example, one
8 tonne of waste incinerated instead of sent to landfill reduces emissions of carbon
9 dioxide by 1.2 tonnes (Jamasb and Nepal, 2010) and electricity generated can offset
10 686g CO₂ kWh⁻¹ for coal and 261g CO₂ kWh⁻¹ for natural gas (Porteous, 1998).

11 Papageorgiou et al. (2009) modelled the greenhouse gas (GHG) emissions
12 from the incineration of waste, critically including emissions from transport. The
13 study concluded that without combined heat and power (CHP) the plant can
14 contribute 3.93kg CO₂ per tonne of waste, although with CHP, coupled with district
15 heating, a reduction of 148.02 kg CO₂ per tonne of waste can be obtained. Cleary
16 (2009) reviewed a range of published studies and concluded that recycling will often
17 be more favourable, in terms of Carbon footprint, where the processing of raw
18 materials is avoided. As such it is important to consider the boundaries used within a
19 lifecycle assessment.

20 In the case of resource depletion, this is discussed with respect to recycling
21 and resource recovery. However the case that incineration undermines recycling rates
22 is widely disproved in the United States and across Europe, with countries and states
23 that accept incineration also having higher recycling rates (Michaels, 2009; Porteous,
24 2005; Berenyi, 2008; Kiser, 2003).

25 Whereas recycling saves considerable energy for aluminium, steel and glass

1 this cannot be said of plastics, as the energy inherent in the material can be better
2 exploited through energy recovery, especially when transport and sorting are taken
3 into account (Lea, 1996). The case for reducing waste is accepted across the board
4 however the economic and practical issues of such widespread change is a barrier to
5 implementation.

6 Therefore it is fair to say that recycling should not be automatically chosen
7 above incineration if the energy used and carbon emitted to recycle the waste is more
8 than could be generated and offset by incineration (Institution of Mechanical
9 Engineers, 2007; Porteous, 1998). One could conclude that neither recycling nor
10 incineration represent the complete solution to sustainable waste management, and
11 should perhaps be considered together as part of a joint solution.

12 The full lifecycle of materials contained within residual waste streams needs to
13 be considered along with the Carbon costs of recovering value (recycled material or
14 energy) before a decision can be made on the most favourable option. This needs to
15 take into account the biogenic fraction of the residual waste along with the fossil-
16 derived fraction, such as plastics.

17

18 **2.3. Toxic waste**

19 The use of the term “toxic waste” from incinerators is an attempt to associate
20 the process with the negative stigma of chemical and nuclear plants. The waste
21 products from an incinerator are in two classes, bottom ash and fly ash. Both of
22 which require careful management to ensure minimal impacts on the environment.
23 Bottom ash is the larger of the two waste streams with 1 tonne of MSW producing
24 typically 0.25 tonnes of bottom ash (Porteous, 2005). This bottom ash is chemically
25 inert and is commonly used as a road building substrate (Porteous, 1998). In the case

1 where stable bottom ash is used as a road aggregate, the leaching properties have been
2 reported to be environmentally acceptable for use (Torraldo *et al.* 2013). The fly ash is
3 a small proportion, typically 40 kg per tonne of MSW, and is termed hazardous waste,
4 due to the levels of volatile heavy metals including cadmium and mercury. The ash is
5 currently disposed of in dedicated hazardous landfill; the fly ash is often treated
6 through solidification/stabilisation processes prior to disposal and so leaching from
7 the landfill into ground water is minimal (Lui *et al.* 2013).

8

9 **2.4. Noise, traffic and visual impact**

10 The noise, traffic and visual impact are usually cited with respect to some
11 health concerns, i.e. particulates from lorry traffic, but also with the effects on local
12 house prices. The most widely cited studies of the economic impacts on house prices
13 are by Kiel and McClain (1995a; 1995b). They have published two studies on the
14 effect of an EfW plant, built in 1985 in North Andover, Massachusetts, USA, on local
15 house prices.

16 Keil and McClain (1995a) attempted to model the impact of the EfW plant
17 using the hedonic pricing method. The study was concerned primarily with assigning
18 a cost associated with distance to the incinerator over 5 time periods from 1974 to
19 1992.

20 They found that each time period had a different impact on the local house
21 prices, with the construction phase amounting to a cost of \$2283 within the first mile
22 (distance from the incinerator), the on-going operation at \$6607, rising to a peak of
23 \$8100 during the online, early operation, stage.

24 Kiel and McClain (1995a) found that the negative effect of the incinerator was
25 only found within three-quarters of a mile (1.2 km) from the incinerator. However

1 properties further from the incinerator benefited from an increase in the house prices
2 up to a maximum distance of approximately 3.5 miles (5.6 km). This premium was
3 found to persist at least 7 years after the facility started operating (Kiel and McClain,
4 1995a) .

5 Kiel and McClain followed up the study with another on appreciation rates
6 (Kiel and McClain, 1995b). In this study they used the same area of North Andover,
7 Massachusetts, USA and applied two different approaches, an income capitalisation
8 model and repeat-sales technique. In this study they found similar results to the
9 previous work with reducing distance to the incinerator leading to a decrease in house
10 value.

11 The Centre for Economics and Business Research (CEBR) (2003), produced a
12 report into the economic impact of an EfW in Newhaven, UK. Along with using the
13 Kiel and McClain (1995a) model and a DEFRA landfill study, a separate analysis was
14 carried out on 10 incinerators in England and Wales. The results showed that house
15 prices were around 18% lower closer to the incinerators and increased with distance.
16 These studies demonstrate properties close to an Energy from Waste plant are
17 subjected to a negative cost due to a local disamenity,

18

19 **2.5. Public perception**

20 The review of anti-incineration groups and the arguments used would suggest
21 that the public is widely against the use of incineration as part of a waste management
22 strategy. However, surveys indicate that only 13.3% of US residents (Furuseth and
23 O'Callaghan, 1991) and 6% of UK residents (Bredee and Georgeson, 2011) are
24 generally opposed to the use of waste incineration. “The public is not as opposed to
25 incineration as people might think” (Achillas et al., 2011).

1 Although this is not the whole story, as when respondents were asked about
2 siting an incinerator in their local area the survey results highlighted a different effect,
3 with 68.9% (US) (Furuseth and O'Callaghan, 1991) and 23% (UK) (Bredee and
4 Georgeson, 2011) opposing such schemes. These results demonstrate that the public's
5 perception of the technology is independent of the location of such facilities. This is
6 commonly described as the NIMBY (Not In My Back Yard) effect and is generally
7 cited with reference to any large industrial construction in local areas.

8 Public engagement is of paramount importance when encountering opposition
9 to planning. Bull et al. (2010) produced a case study of two incinerators built by the
10 same company in different areas and with different council cultures with respect to the
11 engagement process. Their results clearly demonstrate that fairness, compatibility,
12 awareness and impact on the decision process produced a more positive outcome.

13 Compensation schemes are used by waste companies to help with public
14 engagement. A recent survey into the attitudes to community buy-in by SITA UK
15 demonstrated that 71% of people surveyed would be happy to accept a new facility if
16 there was a compensation scheme, which is higher than the 58% that would support a
17 waste treatment facility in their local area without compensation (Bredee and
18 Georgeson, 2011). Although without a universal measure into the disamenity the
19 level compensation would be subjective and is often seen as a bargaining tool by local
20 residents (Snell, 2012).

21 It can be summarised that there is a widespread campaign against incineration
22 based on health, economic and ideological grounds that can be seen to influence the
23 public perception of such facilities.

24 This study aims at building on the existing knowledge by assessing the effect
25 of the public perception of thermal waste treatment plants on local neighbourhoods in

1 terms of property prices.

2

3 **3. Method**

4 A list of all operational EfW plants in the UK was compiled and parameters
5 such as location, size, local area type, operation date, planning date and distance to the
6 closest dwelling were collated. Data was to be sourced for at least 7 years after the
7 plant became operational, as Kiel and McClain (1995a) suggested that an impact was
8 seen at least to this time frame, and for a couple of years before planning was
9 announced as a baseline for comparison.

10 The data was sourced from the Land Registry price paid dataset as their data
11 was considered to be the most comprehensive, reliable and commercially available in
12 the UK. A constraint of using this data were the options available to obtain data by a
13 geographical area and the absence of detailed property information, for example size
14 and number of bedrooms. It was not possible to directly obtain sales records within a
15 set distance of the location of an incinerator, as data could only be searched based on
16 Local Authority, District, Postcode District or Region. Therefore, using GIS software
17 (ArcGIS), the postcode districts surrounding the short listed plants were identified.

18 The postal districts were analysed by the proximity to the plant, surrounding
19 geographical features and number of postcodes within the districts in a 5 km radius of
20 the plant. For example SO40 was chosen as the only relevant postal district in the
21 area around Marchwood EfW as it was the only district to have properties within 1km
22 of the plant. The postal districts on the opposite shore of Southampton Water were
23 discarded as the sea was considered a buffer to the effects of the incinerator. A
24 preliminary search was then carried out to find the total number of property sales
25 records in the selected postal districts between 1st January 1998 and 30th June 2012

1 (Table 2).

2 With respect to practicality of the data sets, three incinerators, Marchwood,
3 Chineham and Kirklees, were chosen out of the short list of six (Table 2) which
4 resulted in 49,299 records. The Newlincs facility was discounted due to the relatively
5 large distance from the nearest dwelling. Marchwood and Chineham were chosen
6 over the Sheffield and Portsmouth facilities as these allowed an insight into two
7 significantly different locations (coastal industry and rural respectively), Additionally
8 Marchwood is closer to the nearest houses development than the Portsmouth EfW
9 (also coastal industry) and so was regarded as more appropriate for this study. The
10 Kirklees facility, along with the Sheffield EfW is based in a city, however Kirklees is
11 significantly closer to the nearest domestic dwelling over a wider area, and has been
12 in operation for longer.

13 Once the property sales data had been obtained it was necessary to find the
14 distance from each property to the incinerator. This was done using GIS software
15 (ArcGIS) and the Code-Point® Open postcode data. The distance from each postcode
16 to the incinerator was then converted into 1 km zones, for example properties within 1
17 km of the incinerator were assigned to zone 1, between 1-2 km to zone 2 etc.

18

19 >>>>>Insert Table 2<<<<<<<

20

21 To compensate for the fluctuations of property prices and inflation over the
22 period, each sale price was divided by the local house price index, also obtained from
23 the Land Registry, for the month the sale took place.

24 As detailed information on the property sales, such as number of bedrooms,
25 house and garden size etc. was not available with the Land Registry dataset, it was not

1 feasible to construct a model using the hedonic pricing method as with the previous
2 Kiel and McClain (1995a; 1995b) studies. Therefore a statistical analysis of the
3 effects of distance, sale price and operation of the incinerator was designed.

4 To provide a robust comparison it was necessary to determine properties that
5 had been sold prior to the operation of the facility and also after the plant became
6 operational to ensure that the analysis was based on a like-for-like approach. The
7 postcode and property name/number were used to produce a unique property
8 identifier. This identifier was then used to match sales for each property before and
9 after the operation of the incinerator in each location. Where a property had sold
10 more than once during a single period the earliest sale was used.

11 Properties that were sold as leasehold were excluded from the analysis as the
12 effect of lease duration was not possible to be identified. Due to the large dataset and
13 complex matching, the data was manipulated using database software (MySQL).

14 A total of 3,458 matching pairs, indicating the same property sold before and
15 after the incinerator was operational, were found across the three locations (Table 3).
16 To reduce the residuals the natural log of the prices sold in each pair, before and after,
17 divided by the index, was calculated. A repeated measures factorial ANOVA
18 statistical analysis was carried out on the data, with zone as an independent variable,
19 using statistics software (Statistica 11).

20

21 >>>>>Insert Table 3<<<<<<<<

22

23 **4. Results & Discussion**

24 The results of the statistical analysis are shown in Figures 1-3. To provide an
25 equal comparison, only results up to 5 km are shown as this was the maximum

1 distance of properties found in the Kirklees analysis. The primary hypothesis, as
2 detailed previously, suggested the operation of the plant will reduce the sale prices in
3 the local area closest to the facility. However none of the incinerator locations
4 demonstrated a statistically significant ($p < 0.05$) decrease in property prices after
5 operation of the plant with regard to distance up to 5 km.

6

7 >>>>Insert Fig 1.<<<<<<<

8 >>>>Insert Fig 2.<<<<<<<

9 >>>>Insert Fig 3.<<<<<<<

10

11 Under close inspection, the results of Marchwood (Fig. 1) revealed a small
12 increase in sale prices between zone 2 and zone 1 before operation of the plant, which
13 was reversed after the plant became operational. This effect was in line with the
14 primary hypothesis but was not statistically significant ($p < 0.05$). The sales prices in
15 zone 3 were much higher than in all other zones, both before and after the plant was
16 operational. Although the error bars were also very large and the results for this zone
17 are therefore not considered to be reliable. zones 4 and 5 indicate an overall lower
18 property value compared to zone 1 and 2 over both time periods. The results in zone
19 4 and 5 show that the property values after the facility become operation are
20 significantly higher than before, with the smallest error bars for the result set
21 demonstrated in zone 5. The price rises across the zones, excluding zone 3, in the
22 period after operation are not statistically significant. Although there is a significant
23 indication, before operation, of lower prices between zones 1 and 4.

24 Chineham (Fig. 2) demonstrated a statistically significant effect with distance,
25 although this was contrary to the hypothesis and other studies, as closer to the

1 incinerator resulted in higher prices. Kirklees (Fig. 3) illustrated a statistically
2 significant effect with distance that was in agreement with the hypothesis, however no
3 effect due to the operation of the incinerator could be found. The results of Chineham
4 (Fig. 2) and Kirkleess (Fig. 3) indicated property prices were higher in the period after
5 the plant started operation; however this was not statistically significant.

6 In this study there were some limitations with the data. Firstly the postcode
7 data used to find the distance from the incinerators was the most up to date in October
8 2012. The postcodes and boundaries are subject to change and therefore some of the
9 older sales no longer had valid postcodes and were therefore discarded. It could be
10 possible to obtain historic postcodes however this was not possible due to the
11 constraints of this project.

12 The unique property identifier used for matching property sales before and
13 after the plant became operational was based on the postcode and property
14 name/number, as previously described. Due to the changing postcode boundaries,
15 some properties may have had different postcodes in each time period and therefore
16 would have been ignored during the matching process. A unique property identifier
17 based on the address or geographic location may lead to more matching pairs.

18 The results displayed wide error bars across certain zones. These
19 discrepancies can be explained by the number of data points for each location,
20 especially in zone 3 at Marchwood and in zone 5 at Kirklees (Table 4). These zones
21 have a very low concentration of properties at both sites as they are primarily
22 agricultural with very few properties. This may explain the large sale price increase
23 in zone 3 of Marchwood (Fig. 1) which only consisted of 17 data points, primarily
24 higher value detached properties. Whilst the house price data was normalised against
25 the local house price index, these larger rural properties are likely to be more

1 desirable, resulting in higher prices, and may not follow the same trend as an average
2 property in the area. The relatively few data points is a likely source of error within
3 the presented data, as demonstrated by the large error bars in the results for this zone
4 (Fig. 1). Kirklees (Fig. 3) demonstrated a similar increase in zone 5 with only a single
5 data point recorded.

6

7 >>>>>Insert Table 4<<<<<<<<<

8

9 The analysis of Chineham (Fig. 2) indicated a significant increase in sale
10 prices close to the site of the incinerator. This was contrary to the expected results.
11 This could be explained by the location of these properties and the site of the
12 incinerator itself. The Chineham incinerator is in a rural location, surrounded by
13 agricultural land with the closest properties 600 m from the plant. These properties
14 are on an estate which is separated from the site by a main road (A33), open
15 countryside and is shielded by mature trees. It is believed that the amenity of this
16 open countryside to the properties could be masking any small effect that the
17 incinerator may have. This may be demonstrated as greater distances from the
18 incinerator mirror greater distances from the surrounding countryside. The results
19 exhibited an inversely proportional relationship of sale price to distance of both
20 factors.

21 At Marchwood (Fig. 1) a similar effect may be at play before the incinerator
22 became operational. The incinerator at Marchwood is in the middle of a large
23 industrial estate, but the closest properties are surrounded by open greenfield space on
24 the opposite sides, with one estate also bordering on Southampton Water. The
25 proximity to the sea was considered to be the primary benefit to the area, as is

1 demonstrated in zone 1 before operation of the incinerator (Fig. 1). However once the
2 incinerator began operation a reversal in property prices was observed. This effect
3 was in agreement with Kiel and McClain's (1995b) results that indicate house prices
4 are lower within three-quarters of a mile (1.2 km) of an operating facility than before
5 it was operational.

6 The Marchwood incinerator is the largest of those analysed and has a very
7 distinctive design. It can be seen from a number of locations in the local area and
8 would be immediately recognisable to local residents. This visual impact could be a
9 cause of the decrease in sales prices in this area. However, it must be noted that the
10 reduction in sales price was not found to be statistically significant in the analysis.

11 Kirklees is an inner city incinerator and is located in a heavy industrial zone
12 adjacent to a railway line. Properties in Kirklees have the lowest average value of the
13 areas studied. These effects in themselves may explain the reduced sale prices closer
14 to the plant that are demonstrated in the results (Fig. 3). This general disamenity can
15 be identified as prices were also lower before the operation of the incinerator,
16 therefore the effect of the operation of the incinerator cannot solely explain the
17 reductions.

18 All the locations analysed were on the sites of previous incinerators. However,
19 each had been non-operational before the planning and construction of the new plant.
20 This may have had a lingering effect on the surrounding properties, which may have
21 already been discounted, and could explain why no significant effect was discovered
22 between the periods before and after operation of the new facilities.

23 Many studies have been carried out into the economic impacts of hazardous
24 waste sites and nuclear power plants (Farber, 1998). These studies agree that house
25 values are reduced close to the site and increase with distance. However these studies

1 are not considered to be comparable to EfW facilities due to the large differences in
2 perceived risk factors.

3 The results obtained from this study are not in agreement with the previous
4 studies by Kiel and McClain (1995a; 1995b) or the CEBR (2003). The negative costs
5 associated with distance calculated by Kiel and McClain (Kiel and McClain, 1995a)
6 are not compared to other costs or benefits that may be contributing to the final house
7 price and therefore the final proportion of the effect of the incinerator cannot be
8 quantified. Although it is considered that the Kiel and McClain studies were carried
9 out at a time when incinerators had a worse perception than today. It is posited that
10 the negative perception of incineration is less pronounced now than in the 1980s and
11 1990s, due to the increased emission controls of the Waste Incineration Directive in
12 2000 and increased public engagement by waste companies. It was acknowledged
13 that a weakness in the CEBR work was that no assertion of causality could be proved.
14 This was due to using a single point in time, Q4 2002; this criticism was not valid for
15 the North Andover studies (Kiel and McClain 1995a; 1995b) which used a continuous
16 time series.

17 The effect found in the CEBR study could be attributed to the general
18 industrial nature of the locations in which many of the incinerators studied were sited
19 and that other factors in these zones, such as industrial noise, pollution or increased
20 heavy goods traffic, could be compounding the results. The general disamenity of
21 industrial estates is demonstrated by the results found at Kirklees in this study.

22 Kiel and McClain (1995a; 1995b) briefly discuss compensation programs for
23 the local residents and conclude that current schemes are inadequate as they do not
24 take into account the changing costs over time as discovered in their study. However
25 the results of this study call into question the validity of compensation schemes that

1 are intended to reimburse the perceived cost of the facility to the local residents. If no
2 negative effect on property prices can be quantified due to the operation of a new
3 facility, it would render the argument for compensation moot. Although it is more
4 likely that such schemes are offered to facilitate the planning process.

5 This study is a starting point for more research into the public perception of
6 modern incinerators and their effect on local economic variables. Recommendations
7 for further research are detailed below.

8 As in the Kiel and McClain (1995b) study more phases of the development
9 and operation of the incinerator could be considered. a long period after the operation
10 that could be broken down into further phases to identify any effects of long term
11 operation. The period before operation could be split into a baseline, before any
12 planning notifications, a planning and a construction phase. This could be used to
13 identify any effects from the construction that may have been included as the baseline
14 used in this study.

15 The number of matched price pairs for Kirklees was smaller than both other
16 facilities (Table 3), mainly due to the short time period between the operation of the
17 plant (2000) and the start date of the Land Registry price dataset (1998). To provide a
18 clearer baseline an equal time period before operation of each facility could be
19 acquired, however the analysis in this study is considered to be robust.

20 The zones chosen in this study were in 1 km increments. Smaller increments
21 could be analysed to identify if at very close range there is a statistically significant
22 effect of the operation of the incinerator. However it should be noted that the records
23 of matched pairs would be significantly fewer and could affect the robustness of the
24 study.

25 The data used in this study was obtained from the Land Registry and did not

1 include any details apart from the type and tenure of the property. Further data could
2 be obtained to provide a more thorough analysis of the other local effects contributing
3 to the sale price. A full hedonic pricing model would enable a cost to be attributed to
4 the incinerator and comparative local benefits.

5 The distances calculated in this study were from the postcode for each
6 property. A postcode may have multiple properties associated with it and therefore
7 the distances for each property were approximate. Individual positions for each
8 property may help to eliminate this issue and provide more accurate distance
9 calculations.

10 The action of supply and demand is generally agreed to be the primary effect
11 on the setting of market prices. The analysis of the sales frequency at the sites in this
12 study could provide more insight into the perception of the facilities and possibly
13 identify any change in demand across the time periods.

14 Large opposition groups can sway the perception of the public and a
15 qualification of the strength of opposition to the incinerator should be carried out and
16 quantified to identify any effect. This could include a review of local and national
17 newspaper articles about the planned incinerator to deduce the sentiment, identify the
18 number of official oppositions to the planning applications and the arguments
19 provided, conduct interviews with company representatives to gauge the level of
20 opposition they experienced and interview local people to identify their perception of
21 the plant before and after operation.

22 Compensation schemes could have been used to improve the perception of the
23 incinerator and operating companies detailed in this study. Identification of the use of
24 these schemes, if any, and the quantification of any effect that this may have had on
25 the perceived disamenity could be carried out.

1 The Marchwood incinerator is a large dome-shaped facility that is visually
2 distinctive from the standard “grey box” of traditional incinerators. Additional
3 research into the effect of the design of the plant on local residents and whether the
4 design was chosen to help with public engagement could be undertaken.

6 **5. Conclusion**

7 The results indicate that if there is a perceived cost of living close to an
8 operational incinerator, then this value would be greatly outweighed by local benefits
9 such as proximity to open countryside, access to the sea, transport links and possibly
10 catchment areas for local schools.

11 In conclusion this study did not find any significant negative effect on property
12 prices at any distance within 5 km from a modern operational incinerator. However,
13 more research is suggested which could include:

- 14 • Additional time periods, such as the different phases of development and
15 operation.
- 16 • Longer time periods before facility operation;
- 17 • Additional zones, allowing for smaller increments between zones (i.e. zones
18 smaller than 1 km);
- 19 • Acquisition of additional housing data, such as the type of house and the
20 number of residents;
- 21 • Identify the effects of facility design including whether the design was chosen
22 to help with public engagement.

24 **Acknowledgements**

25 The authors gratefully acknowledge the input from experts in the waste sector

1 and to Dr Monica Rivas Casado and Mr Ian Truckell for their valuable financial
2 modelling and mapping expertise. The opinions expressed are the author's alone.

3

4 **References**

5

6 Achillas, C., Vlachokostas, C., Moussiopoulos, N., Banias, G., Kafetzopoulos, G. and
7 Karagiannidis, A. (2011), "Social acceptance for the development of a waste-to-
8 energy plant in an urban area", *Resour., Conserv. Recycling*, vol. 55, no. 9–10, pp.
9 857-863.

10 Berenyi, E., (2008), *Recycling and Waste-to-Energy: Are They Compatible?*,
11 Governmental Advisory Associates, Inc., Westport, CT.

12 Bredee, H. and Georgeson, R. (2011), *Public attitudes to community buy-in for waste
13 and resource infrastructure*, SITA UK, Maidenhead, Berkshire.

14 British Society for Ecological Medicine (2008), *The Health Effects of Waste
15 Incinerators - 4th Report of the British Society for Ecological Medicine*, 2nd edition
16 ed., British Society for Ecological Medicine.

17 Bull, R., Petts, J. and Evans, J. (2010), "The importance of context for effective public
18 engagement: Learning from the governance of waste", *J. Environ. Plan. Manag.*, vol.
19 53, no. 8, pp. 991-1009.

20 Burnley, S., Phillips, R., Coleman, T. and Rampling, T. (2011), "Energy implications
21 of the thermal recovery of biodegradable municipal waste materials in the United
22 Kingdom", *Waste Manag.*, vol. 31, no. 9–10, pp. 1949-1959.

23 CEBR (2003), *The Economic Impact of an EfW Incinerator in Newhaven*, Centre for
24 Economics and Business Research Ltd.

25 Chang, N. and Wang, S. F. (1995), "The development of material recovery facilities in

1 the United States: status and cost structure analysis", *Resour., Conserv. Recycling*, vol.
2 13, no. 2, pp. 115-128.

3 Cheng, H. and Hu, Y. (2010), "Municipal solid waste (MSW) as a renewable source of
4 energy: Current and future practices in China", *Bioresour. technol.*, vol. 101, no. 11,
5 pp. 3816-3824.

6 Cleary, J. (2009), "Life cycle assessments of municipal solid waste management
7 systems: A comparative analysis of selected peer-reviewed literature" *Environ. Int.*
8 vol.35, no.8, pp. 1256-1266

9 DEFRA (2004), *Review of Environmental and Health Effects of Waste Management:*
10 *Municipal Solid Waste and Similar Wastes*, Department for Environment, Food and
11 Rural Affairs, London.

12 DEFRA (2007), *Incineration of Municipal Solid Waste*, Department for Environment,
13 Food and Rural Affairs, London.

14 DEFRA (2011), *Local authority collected waste management statistics for england –*
15 *final release of quarters 1, 2, 3 and 4 2010/11.*, Department for Environment, Food
16 and Rural Affairs, London.

17 Enviros (2006), *Evaluation of the 4th Report of the British Society for Ecological*
18 *Medicine: "The Health Effects of Waste Incinerators"*, Enviros Consulting Ltd.

19 Council of the European Union (1999), *Landfill Directive - Council Directive*
20 *1999/31/EC*, Directive ed., European Union.

21 Elliott, P., Eaton, N., Shaddick, G. & Carter, R. (2000), "Cancer incidence near
22 municipal solid waste incinerators in Great Britain. Part 2: Histopathological and
23 case-note review of primary liver cancer cases", *Brit. J. Cancer*, vol. 82, no. 5, pp.
24 1103-1106.

25 Farber, S. (1998), "Undesirable facilities and property values: a summary of empirical

1 studies", *Ecol. Econ.*, vol. 24, no. 1, pp. 1-14.

2 Friends of the Earth (2000), *How to Win: Campaign against incinerators*, Friends of
3 the Earth, London.

4 Furuseth, O. J. and O'Callaghan, J. (1991), "Community response to a municipal
5 waste incinerator: NIMBY or neighbor?", *Landsc. Urban Plan.*, vol. 21, no. 3, pp.
6 163-171.

7 Haley, C. A. C. (1990), "Energy recovery from burning municipal solid wastes: a
8 review", *Resour., Conserv. Recycling*, vol. 4, no. 1-2, pp. 77-103.

9 Health Protection Scotland (2009), *Incineration of Waste and Reported Human Health*
10 *Effects*, Health Protection Scotland, Glasgow.

11 HMRC (2011), *Overview of Tax Legislation and Rates*, HM Revenue and Customs,
12 London.

13 HPA (2005), *HPA response to the British Society for Ecological Medicine report*,
14 Health Protection Agency, London.

15 HPA (2009), *The Impact on Health of Emissions to Air from Municipal Waste*
16 *Incinerators*, Health Protection Agency, London.

17 Institution of Mechanical Engineers (2007), *Energy from Waste: A Wasted*
18 *Opportunity?*, Institution of Mechanical Engineers.

19 Jamasb, T. and Nepal, R. (2010), "Issues and options in waste management: A social
20 cost-benefit analysis of waste-to-energy in the UK", *Resour., Conserv. Recycling*, vol.
21 54, no. 12, pp. 1341-1352.

22 Kiel, K. A. and McClain, K. T. (1995a), "House Prices during Siting Decision Stages:
23 The Case of an Incinerator from Rumor through Operation", *J. Environ. Econ. and*
24 *Manag.*, vol. 28, no. 2, pp. 241-255.

25 Kiel, K. A. and McClain, K. T. (1995b), "The Effect of an Incinerator Siting on

1 Housing Appreciation Rates", *J. Urban Econ.*, vol. 37, no. 3, pp. 311-323.

2 Kiser, J.V.L. (2003), *Recycling and Waste-to-Energy: The Ongoing Compatibility*
3 *Success Story*, MSW Management.

4 Lea, W. R. (1996), "Plastic incineration versus recycling: a comparison of energy and
5 landfill cost savings", *J. Hazard. Mat.*, vol. 47, no. 1-3, pp. 295-302.

6 Michaels, T. (2009), *Waste Not, Want Not: The Facts Behind Waste-to-Energy*, Energy
7 Recovery Council.

8 Liu, J., Nie, X., Zeng, X., and Su, Z., "Long-term leaching behavior of phenol in
9 cement/activated-carbon solidified/stabilized hazardous waste" *J. Environ. Manag.*,
10 vol. 115, 265-269.

11 Murphy, J. D. and McKeogh, E. (2004), "Technical, economic and environmental
12 analysis of energy production from municipal solid waste", *Renewable Energy*, vol.
13 29, no. 7, pp. 1043-1057.

14 National Atmospheric Emissions Inventory, (2013), "UK emissions data selector – Air
15 pollutants", <http://naei.defra.gov.uk/data/data-selector?view=air-pollutants> (accessed
16 22nd June 2013).

17 Papageorgiou, A., Barton, J. R. and Karagiannidis, A. (2009), "Assessment of the
18 greenhouse effect impact of technologies used for energy recovery from municipal
19 waste: A case for England", *J. of environ. manag.*, vol. 90, no. 10, pp. 2999-3012.

20 Porteous, A. (1998), "Energy from Waste: A Wholly Acceptable Waste-mangement
21 Solution", *Applied Energy*, vol. 58, no. 4, pp. 177-208.

22 Porteous, A. (2005), "Why energy from waste incineration is an essential component
23 of environmentally responsible waste management", *Waste Manag.*, vol. 25, no. 4, pp.
24 451-459.

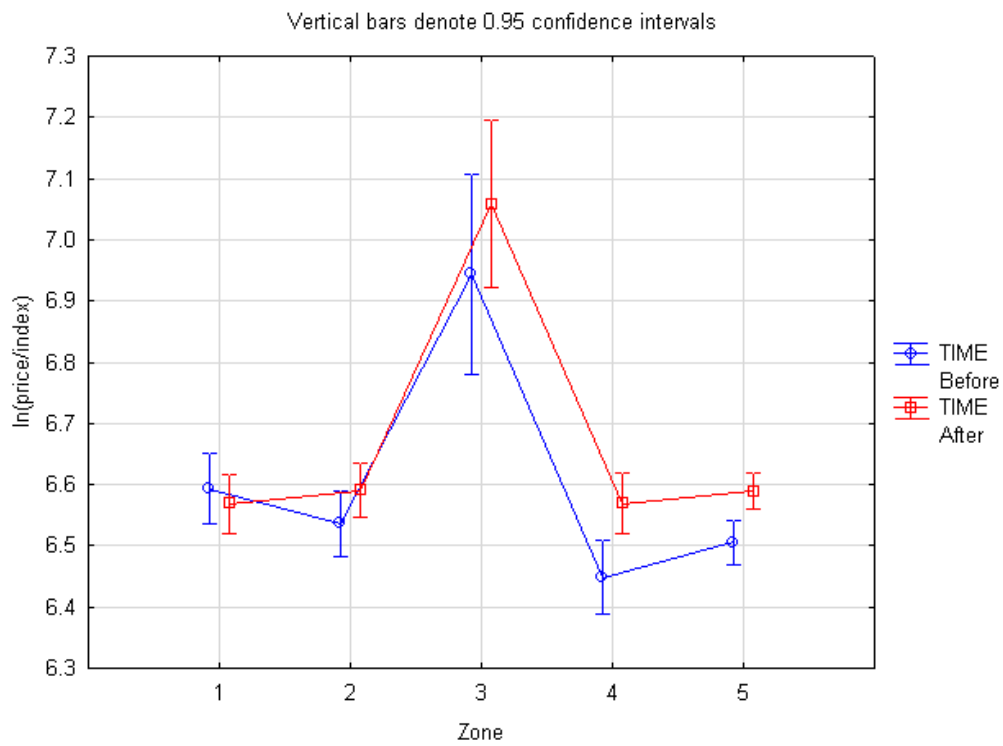
25 Roberts, R.J. and Chen, M., (2006), "Waste incineration - How big is the health risk?"

1 A quantitative method to allow comparison with other health risks”. J. Public Health,
2 vol. 28 no. 3, pp. 261-266.

3 Snell, M. (2012), Personal communication, FCC Environment.

4 Toraldo, E., Saponaro, S., Careghini, A., and Mariani, E. (2013), “Use of stabilized
5 bottom ash for bound layers of road pavements”, J. Environ. Manag. vol. 121, pp 117-
6 123.

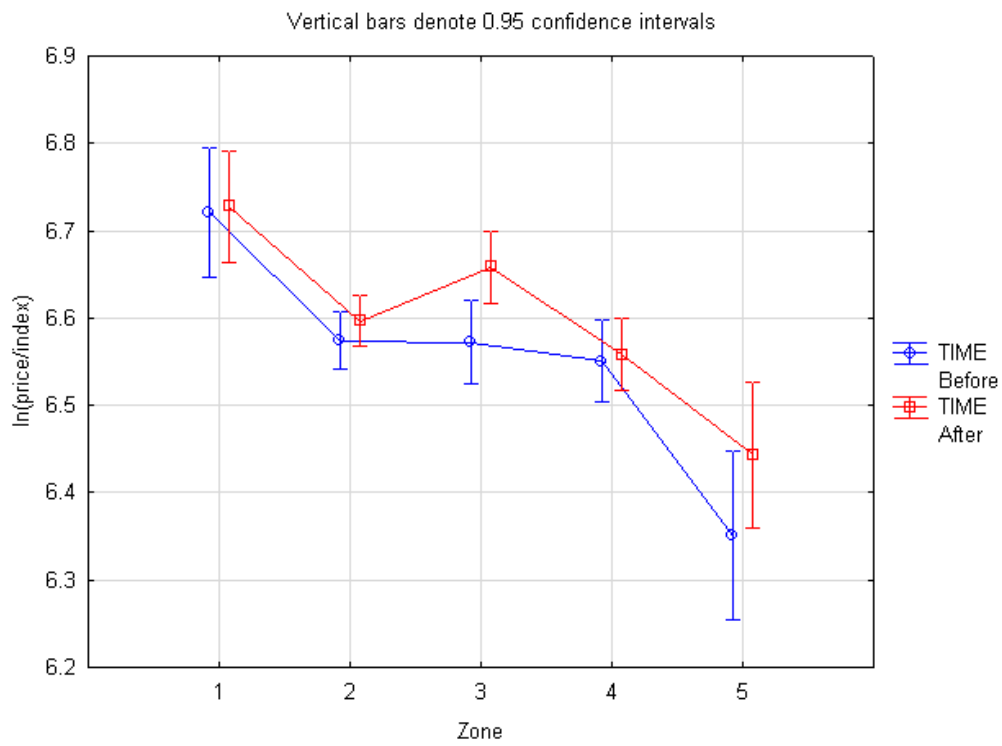
7



1

2 **Figure 1.** Value of properties sold, $\ln(\text{price}/\text{index})$, in zones 1-5 before (1998-2004)
 3 and after (2005-2012) operation of Marchwood EfW. Error bars denote 0.95
 4 confidence intervals

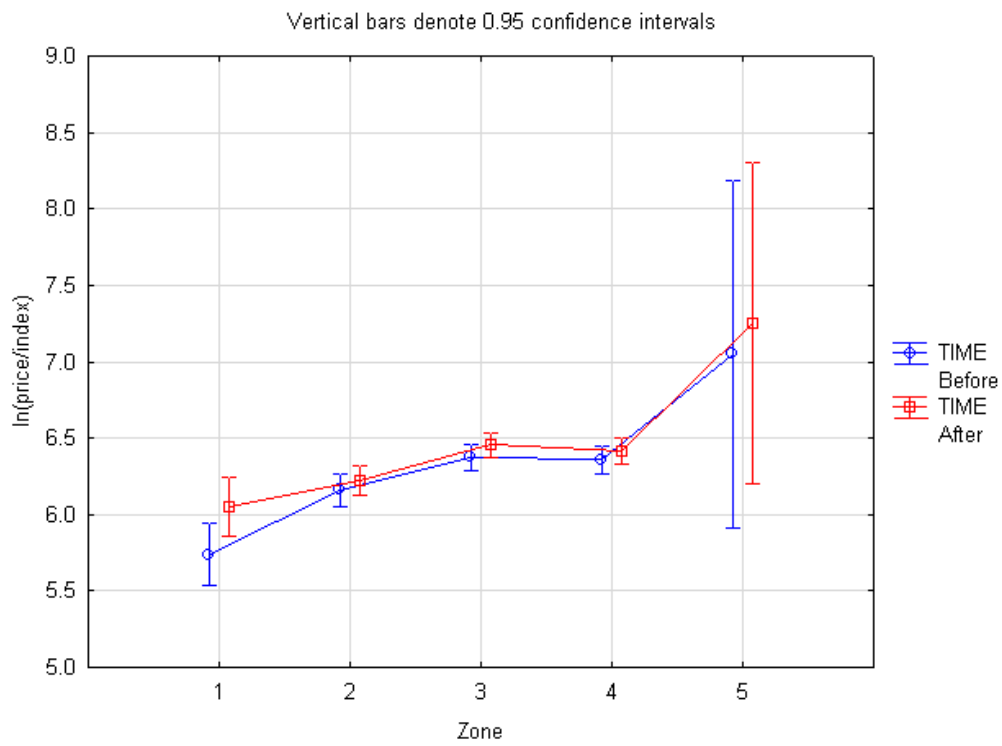
5



1

Figure 2. Value of properties sold, $\ln(\text{price}/\text{index})$, in zones 1-5 before (1998-2002) and after (2003-2012) operation of Chineham ERF. Error bars denote 0.95 confidence intervals.

2



1

Figure 3. Value of properties sold, $\ln(\text{price}/\text{index})$, in zones 1-5 before (1998-1999) and after (2000-2012) operation of Kirklees EfW. Error bars denote 0.95 confidence intervals.

Table 1. Comparison of dioxin levels from different processes 2011

Process	Amount of dioxins produced (grams International Toxic Equivalent)
Power Stations – MSW	0.66g
Power Stations – Coal	1.26g
Road transport – cars (petrol)	1.27g
Sinter plants – iron production	20.73g
Electric arc furnaces – steel production	10.94g
Domestic combustion - wood	3.07g
Domestic combustion - coal	1.45g
Landfill – escaping methane	0.63g
Crematoria	10.23g

1 Source: adapted from National Atmospheric Emissions Inventory (2013).

2

3

1 **Table 2.** Incinerator short list
 2

Incinerator	Operating company	Plant Capacity (ktpa)	Plant Location	Distance to closest dwelling (m)	Planning year	Operational from	House price data points
Marchwood	Veolia	210	Industrial	300	2001	01/12/04	SO40 12,877
Chineham ERF	Veolia	102	Rural (w/water treatment)	600	2000	01/01/03	RG24 11,107 RG27 6,158
Kirklees	SITA	150	Industrial	150	1999	2000	HD1 4,829 HD2 6,840 HD5 7,488
Sheffield	Veolia	225	Industrial	450	2002	2006	S4 43,216 S2 31,889
Portsmouth	Veolia	210	Industrial	450	2001	2005	PO2 47,991 PO3 17,932
Newlines	Cyclerval UK & TIRU Group / Newlines Development Ltd	56	Industrial	2,700	2001	2004	DN41 1,845 DN40 2,789 DN31 3,496 DN37 5,429

3

4

1 **Table 3.** Number of data points for each incinerator

2

Incinerator	Operational year	Before operation		After operation		Matched pairs
		Number of sales	Properties sold	Number of sales	Properties sold	
Chineham	2003	3,773	3,067	5,499	4,465	1,547
Marchwood	2005	6,745	5,095	4,617	3,948	1,426
Kirklees	2000	1,057	1,028	7,614	5,635	485
Totals		11,575	9,190	17,730	14,048	3,458

3

4

5

6

1 **Table 4.** Property type breakdown per zone for each incinerator
 2

Zone	Chineham				Marchwood				Kirklees			
	Detached	Semi	Terrace	Flat	Detached	Semi	Terrace	Flat	Detached	Semi	Terrace	Flat
1	57	26	36	1	29	23	83	3	3	0	6	21
2	195	92	292	1	49	47	68	0	23	1	35	59
3	100	61	131	1	13	3	1	0	64	1	46	66
4	94	29	166	0	21	46	59	0	53	1	42	63
5	10	9	52	0	111	145	102	0	0	0	1	0

3

4

5