

1 Drought early warning based on optimal risk forecasts in
2 regulated river systems: application to the Jucar River Basin
3 (Spain)

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10 **Abstract**

11 Droughts are a major threat to water resources systems management. Timely anticipation
12 results crucial to defining strategies and measures to minimise their effects. Water managers
13 make use of monitoring systems in order to characterise and assess drought risk by means
14 of indices and indicators. However, there are few systems currently in operation that are
15 capable of providing early warning with regard to the occurrence of a drought episode. This
16 paper proposes a novel methodology to support and complement drought monitoring and
17 early warning in regulated water resources systems. It is based in the combined use of two
18 models, a water resources optimization model and a stochastic streamflow generation
19 model, to generate a series of results that allow evaluating the future state of the system.
20 The results for the period 1998-2009 in the Jucar River Basin (Spain) show that accounting
21 for scenario change risk can be beneficial for basin managers by providing them with
22 information on the current and future drought situation at any given moment. Our results
23 show that the combination of scenario change probabilities with the current drought
24 monitoring system can represent a major advance towards improved drought management
25 in the future, and add a significant value to the existing national State Index (SI) approach for
26 early warning purposes.

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29 **Highlights**

- 30 • Modelling the past to anticipate future drought is an ineffective and risky approach
- 31 • A new method for continuous drought monitoring and early warning in regulated
32 catchments is proposed
- 33 • Reservoir storage probability is a reliable indicator for drought status in regulated
34 catchments
- 35 • New approach adds value to existing monitoring and early warning methods

36 **Keywords**

37 Monitoring; Early Warning System; Optimisation Modelling; Water Resources Systems
38 Analysis; Aquatool

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40 **1. Introduction**

41 Droughts are a major threat to the sound operation and management of water resources
42 systems. Developing new approaches to anticipate them will help in defining strategies and
43 measures to minimise their effects. The use of monitoring systems to calculate drought
44 indices and indicators can help water managers characterize droughts and define risk
45 scenarios. The activation of a drought scenario in a system will trigger a number of
46 measures addressed to minimise the possibilities of developing into a worse scenario and
47 minimizing the possible effects of the current situation.

48 The assessment of drought severity requires the use of an index which fulfils well-known
49 criteria (Tsakiris et al. 2013): operational usefulness, physical meaning, sensitivity to a wide
50 range of drought conditions, applicability in all parts of the globe, quick response to changes
51 due to drought and high availability of required data. Commonly, such an index is a prime
52 variable for assessing the effect of a drought and defining different drought parameters,
53 which include intensity, duration, severity and spatial extent as defined by Yevjevich (1967)
54 in his theory of runs. A time series of drought indices provides a framework for evaluating
55 drought parameters of interest. Generally, drought indices are categorized as
56 meteorological, hydrological, agricultural or remote sensing-based (Rossi and Cancelliere
57 2013). Mishra and Singh (2010) and Pedro-Monzonis et al. (2015) made an extensive review
58 of existing univariate drought indices both concluding that each index performance is region
59 specific mostly due to the characteristics of the variables used for their calculation and the
60 purpose of the analysis. In addition, in recent time some authors have also attempted to
61 combine all the variables (e.g. precipitation, soil, water content) that lead to different
62 physical forms of drought in so-called multivariate drought indices (Rajsekhar et al. 2015). In
63 some cases, the index is built as an aggregation of variables selected according to their
64 relation each drought type (Keyantash and Dracup 2004; Rajsekhar et al. 2015). In other, the
65 index is constructed using copulas to derive the joint distribution of two or more variables
66 (Kao and Govindaraju 2010; Hao and AghaKouchak 2013).

67 An indicator system is a drought monitoring system that allows the anticipation in the
68 application of mitigation measures for the reduction of socio-economic and environmental
69 impacts of droughts (Estrela and Vargas 2012). Such systems can also be considered early
70 warning systems for their capacity to anticipate the effects that drought may have on the
71 system in order to trigger necessary mitigation measures (Rossi et al. 2008). In most cases,
72 these systems are normally formed by basic variables selected at different points in a river

73 basin that are capable of defining the current drought status. Their reliability will depend on
74 their capacity to represent, using real-time data: 1) the relationship between significant
75 reductions of water availability with deviations of meteorological and hydrological
76 components from their average; 2) detecting early stages of drought development; 3)
77 provide results that allow comparison between events both in time and space; and 4)
78 assessing the severity of the ongoing situation in order to support decision making for
79 triggering drought mitigation actions. Additionally, in the case of regulated water resources
80 systems, it would be desirable that the indicator is capable of showing the evolution of
81 management and how this would change the drought status of the system if new operation
82 rules are envisaged.

83 Different drought early warning systems have been developed at different spatial scales, but
84 a very small number of such systems are actually in operation (Rossi and Cancelliere 2013).
85 This is mainly due to the low density of meteorological and hydrological gauging networks,
86 the sharing of the data among different agencies with different objectives, and to the lack of
87 universal standards in computing drought indices (Rossi 2003). In addition, the development
88 of indicator systems based on observational frameworks cannot provide sufficient
89 anticipation with regard to the event in progress in order to activate the necessary measures
90 to mitigate its effects (Haro et al. 2014). Efforts have been made to correlate drought indices
91 to impacts (Stagge et al. 2015), but these relationships only provide insight after the event
92 has finished and the impacts reported. Mishra and Singh (2011) acknowledged that to
93 develop suitable techniques for forecasting the onset and termination of droughts is still a
94 major research challenge due to the inability to predict drought conditions accurately for
95 months or years in advance. Due to these inaccuracies and uncertainties, drought
96 management relies nowadays mainly on risk assessment. Risk assessment during the
97 operation phase of a system is often referred as conditioned risk assessment. With this

98 procedure, the state of the system is usually evaluated for the short-term to explore
99 alternative mitigation measures and policies for an ongoing drought episode. This same
100 assessment approach can be adopted for early warning purposes (Cancelliere et al 2009).

101 Alecci et al. (1986) considered that the risk assessment of a water supply system is a
102 problem that is better approached through a set of several indices and analysing the
103 probability of suffering shortages of different entities. This is due to the many complexities
104 existing within a water resources system such as the stochastic nature of inflows, the high
105 interconnection that exists between different components of the system, the competition for
106 water by conflicting demands, the definition of what elements are at risk, and the uncertain
107 character of the impacts in different drought episodes. Traditionally, reliability, resiliency and
108 vulnerability have been the indices used to capture the different performance aspects of
109 water supply systems (Hashimoto et al. 1982). However, these indices are normally
110 representative of just one particular use, defining the state of the system with regard to the
111 probability of a failure for such index. Since all drought events are unique, so too are their
112 effects both temporally and spatially. Therefore, it is necessary to have an indicator that is
113 capable of summarising the state of the system for any given situation. In regulated systems,
114 it will be the volume stored in reservoirs since it provides an overview of the previous
115 management of the system and is the basis for future resources allocation.

116 This paper proposes a novel methodology to support drought monitoring and scenario
117 definition in regulated water resources systems. It is based on the results of two models, an
118 optimisation model and a stochastic streamflow generation model, both of which have been
119 calibrated and validated in previous research (Haro et al. 2012a, 2012b, and 2014b; Ochoa-
120 Rivera 2002). Using storage in reservoirs as a summary indicator of the future system
121 status, we propose a combined use of the two models to generate a series of results that
122 can support and complement drought monitoring and early warning systems currently in

123 place in a river basin. The methodology is applied to the Jucar River Basin in Spain to
124 evaluate the probability of a scenario change several years in advance. The proposed
125 method has the potential to enhance decision making under highly uncertain hydrological
126 situations, and provide water resource planners and managers with new insights both
127 regarding the behavior of the system and the development of drought episodes.

128 **2. Case study description**

129 The Jucar River Basin is located in the eastern part of the Iberian Peninsula in Spain (Figure
130 1). This basin is the most important of the 9 water exploitation systems in the Jucar River
131 Basin Demarcation (Demarcacion Hidrografica del Jucar – DHJ in Spanish). In the Valencia
132 coastal plain, where the Jucar River has its mouth, there is a shallow lake called Albufera,
133 with an associated wetland. Both, the lake and the wetland depend on return flows from
134 irrigated areas in the basin, and also on groundwater flows from the coastal aquifer beneath
135 the plain (Andreu et al. 2009). It is the largest system of the DHJ both in surface (22,261
136 km²) and in volume of resources (1,548 hm³/year).

137 The river is an example of a typical Mediterranean river, characterized by a semi-arid climate
138 in most of the basin territory consisting of low precipitation rates (475mm/year) during the
139 year combined with exceptional convective storms that can lead to flooding and seasonal
140 summer scarcity that occurs when irrigation requirements are at their highest. Urban
141 demand accounts for circa 143.3 hm³/year and the water demand for irrigated agriculture
142 reaches 1034.3 hm³/year. Water supply to small urban areas comes mainly from wells and
143 springs, but large metropolitan areas such as Albacete, Sagunto and Valencia rely on
144 surface water (Andreu et al. 2009). According to the White Book of Groundwater (CEDEX
145 1995), nearly three quarters (73%) of the resources in the territory of the DHJ have
146 subterranean origin. This highlights the major importance that groundwater resources have

147 in the management of these basins. The total amount of available groundwater resources in
148 the basin is 1,225 hm³/year. However, this only represents the estimated volume in all the
149 groundwater bodies without accounting for their sharing between other basins or the
150 relationship these bodies have with the surface water system.

151 With regard to droughts, the Jucar River Basin can be considered to be one of the most
152 vulnerable areas in the western Mediterranean region, due to high water exploitation
153 indexes, and the environmental and water quality problems that arise when droughts occur.
154 This situation has triggered increased use of non-conventional resources in recent years,
155 such as reuse of wastewater and drought emergency wells. Also, conjunctive use of surface-
156 ground waters has historically been a very important option in the region to provide
157 robustness against droughts. The integrated use of these three resource options was
158 considered a major success in adapting to the latest drought episode between 2005 and
159 2008 (Ortega-Reig et al 2014).

160 The operation of the system is mainly multi-year. The Alarcon and Contreras reservoirs, at
161 the headwaters of the system, are capable of storing the highly variable streamflow coming
162 from their upstream sub-basins. The third most important reservoir in the system, the Tous,
163 is operated on an annual basis. Before the summer season it stores incoming mid-basin
164 streamflow and upstream reservoirs releases to supply the different demands within the
165 Valencia Plain. By the end of the summer, the reservoir is emptied in order to prevent floods
166 originated from often intense autumn rainfall events.

167 **3. Methodology**

168 In this section, we present the indicator system currently in use in the Jucar River basin as
169 well as in most of Spanish river basins. Despite being a useful methodology to evaluate the
170 actual drought conditions in the basin, it has low forecasting capacity; making preventive

171 management of droughts inefficient and/or very difficult. To complement the information
172 provided by the indicator, we developed a methodology to derive the probability of drought
173 scenario change for a four year planning horizon. It is based on the Monte Carlo evaluation
174 of the results of multiple runs of an optimization model of the system. Based on this analysis,
175 we derive distribution functions on the future state of the basin and combine them with
176 trigger values for each drought scenario.

177 3.1. Current drought indicator system for Spanish river basins

178 One of the objectives of Spanish Drought Plans is providing means for anticipating drought
179 events. To do this, it is necessary to establish an early warning system that allows
180 forecasting drought characteristics and assessing their effects on the system. Spanish basin
181 operators have adopted a method of drought indicators based on the analysis of historic
182 data that reflect the availability of water in the system. This indicator is known as State Index
183 (SI) and it is the result of combining several hydro-meteorological variables obtained from a
184 monitoring system. The SI has a hydrologic character since its practical interest lays on its
185 ability to serve as decision-making instrument regarding water resources management in the
186 basin. For each catchment, managers select a set of variables that best represent the water
187 resources for different demand units in the basin using values of reservoirs storage,
188 piezometric levels, natural streamflow and areal precipitation. In the case of the Jucar River,
189 the selected variables are detailed in CHJ (2007)¹.

190 For each selected variable, the value of the SI has the following expression (CHJ 2007):

$$\text{If } V_i \geq V_{av} \rightarrow SI = \frac{1}{2} \cdot \left[1 + \frac{V_i - V_{av}}{V_{max} - V_{av}} \right] \quad \text{Eq. 1}$$

$$\text{If } V_i < V_{av} \rightarrow SI = \frac{1}{2} \cdot \frac{V_i - V_{min}}{V_{av} - V_{min}} \quad \text{Eq. 2}$$

¹ A partial translation of the contents in CHJ(2007) is provided in Acacio et al. (2013)

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192 Where V_i is the value of the variable in month i ; V_{av} is the average monthly value of the
193 variable in the historic series considered; and V_{max} and V_{min} are the maximum and minimum
194 monthly values of the variable in the historic series considered respectively. The main
195 reason to follow this calculation approach is that the arithmetic average is a robust statistic,
196 as well as simple; so a comparison of the current variable value with the average of the
197 historic series considered will adjust better to the real situation of the studied region.
198 Additionally, taking into account the maximum and the minimum historic values allows
199 homogenising the different variables into a dimensionless numeric value capable of
200 quantifying the current situation with regard to the historic. This also permits to quantitatively
201 compare the different variables selected between them. Finally, the overall SI of the basin
202 and hence its drought level is defined as the weighted sum of the SI values of each of the
203 selected hydro-meteorological variables. The weight assigned to each variable depends on
204 the level of demand served. For the Jucar River, the SI consists of a combination of 12
205 different variables including precipitation, streamflow, piezometric levels and storage in
206 reservoirs at different strategic points within the basin (CHJ 2007).

207 Spanish Drought Plans establish four different levels of drought, or scenarios, namely:
208 normality, pre-alert, alert and emergency (CHJ 2007). These levels are determined
209 according to the values of the SI with the following thresholds: Normality ($SI \geq 0.5$); Pre-alert
210 ($0.5 > SI \geq 0.3$); Alert ($0.3 > SI \geq 0.15$); and Emergency ($0.15 > SI$). Figure 2 shows the evolution of
211 the SI in the Jucar River Basin between October 1998 and September 2010. Between the
212 end of the XX century and the beginning of the XXI century the basin experienced a short
213 but intense period of drought that made the SI oscillate between the pre-alert and the alert
214 levels until 2002 when the situation returned to normality after a period of intense
215 precipitation. Between 2005 and 2008, the system suffered the worst drought event on

216 record with SI reaching emergency levels several times during that period. After that, the
217 system gradually recovered to pre-alert in 2009 to finally reach the normality level in 2010.

218 Haro et al. (2014) showed the possibility that an indicator such as the SI might be insufficient
219 in order to set and trigger the most appropriate drought mitigation measures early enough to
220 be efficient. This method is limited to determine the current drought situation based on the
221 comparison of present variables values with the variables occurred in the past; making its
222 forecasting capability low, or even non-existent. Moreover, drought episodes vary between
223 one and another. Hence, it is very unlikely that the SI is capable of working as an early
224 warning system for droughts, advancing the real consequences of an upcoming event.

225 In addition, as commented above, it is important that the effects of management decisions
226 and mitigation measures are included in the monitoring process and that their modifications
227 are reflected in order to advance their efficacy and to better support decision-making. For
228 this reason, the use of risk assessment methodologies in combination with indicator systems
229 provides an interesting and novel framework to support decision making during drought
230 situations in regulated systems.

231 3.2. Drought scenario definition based on the risk assessment of the system's optimal
232 operation

233 The methodology developed is based on previous research by Sanchez-Quispe (1999),
234 Andreu and Solera (2006), Andreu et al (2007, and 2013) and Cancelliere et al (2009). Their
235 findings were successfully used in the management of previous drought episodes of the
236 Jucar River Basin. Here we present a further development of existing approaches by
237 introducing an optimisation approach that allows one to obtain the best results achievable in
238 the system and better rules for the application of mitigation and prevention measures. This
239 work further develops that presented by Haro et al. (2014a) by extending its application to a

240 multi-year regulated basin. In addition, we show how the risk assessment methodology
241 presented here is applicable to forecast drought scenarios. Figure 3 provides a schematic
242 summary of the methodology, which is briefly described below.

243 We applied a monthly Monte Carlo optimisation process to a catchment management model
244 of the Jucar River Basin previously developed in the GUI of Aquatool DSS (Andreu et al
245 1996) for the implementation of the European Water Framework Directive (CHJ 2004) and
246 the development of its latest basin plan (CHJ 2015), and shown in Figure 4. The model
247 includes the main surface storage facilities ('Alarcon', 'Contreras', and 'Tous' reservoirs) as
248 well as the main aquifers in the basin that have a crucial role in the management of the
249 system ('Mancha Oriental' and 'Plana de Valencia'). The most important demands are also
250 represented, namely: traditional irrigation in 'Plana de Valencia'; groundwater irrigation from
251 'La Mancha Oriental' aquifer; conjunctive irrigation from the newer developments along the
252 'Jucar-Turia' canal; and the urban demands of Valencia, Sagunto and Albacete, which is
253 minor in quantity but more sensitive to failures in the supply. Haro et al. (2012a and 2012b)
254 and Haro Monteagudo (2014) provide a detailed description of the optimization technique,
255 equations and constraints utilised by the model, as well as the input data it needs. A
256 previous application can also be found in Haro et al (2014b). The model runs on a monthly
257 time step fed by synthetic streamflow series generated stochastically from historically
258 observed monthly values between 1980 to 2012. There are 16 streamflow input nodes along
259 the model network, represented as thick red arrows in Figure 4. The synthetic series were
260 generated with the stochastic analysis and modelling module in Aquatool (Ochoa-Rivera
261 2002). The 16 observed streamflow time series were normalised and standardised to
262 calibrate the autoregressive model, AR(1), shown in equation 3:

$$X_t = \boldsymbol{\varphi}_1 \cdot X_{t-1} + \boldsymbol{\theta}_0 \cdot \varepsilon \quad \text{Eq. 3}$$

263 where X_t and X_{t-1} are n variables vectors; φ_1 is an $n \times n$ autocorrelation matrix; θ_0 is an
264 $n \times n$ matrix of coefficients that multiplies the random $N(0,1)$ values vector represented by ε .
265 For this case, n has a value of 16. For the stochastic generation of synthetic streamflow
266 series from observed values, the last monthly observed value is used as a seed after
267 normalisation and standardisation. The generated time series of standardised values are
268 converted to streamflow values following the inverse path. The validation of the model
269 against the long term characteristics of the historic series (average, standard deviation,
270 number of dry years), makes it suitable to explore a large range of events.

271 The results of each optimisation run in the Monte Carlo process are the time series of
272 reservoirs storage and releases, surface and groundwater supply to the different demands,
273 aquifers relative storage and recharge, and flows in river streams. The statistical analysis of
274 all runs yields a number of indicators to assess risk.

275 When confronting an ongoing drought situation from a risk minimisation approach and a high
276 level of uncertainty, it is more useful to rely on an index that summarizes the status of the
277 basin considering all the possible events. In the case of regulated river basins, this index is
278 the state of the reservoirs. The evolution of storage in reservoirs clearly reflects the
279 operation of a system during previous periods of time, and their present status defines the
280 future use possibilities. Hence, reservoir level state probability and storage probability are
281 useful indicators with regard to drought in a regulated catchment and may support the
282 decision making process with information about what can be expected in the future.

283 Based on the previous consideration, we use the storage probability in the different
284 reservoirs in the basin as the basis to determine the risk level and the change of scenario
285 probability at the end of a number of campaigns for each month. It must be noted that
286 reservoirs levels is an important element in the Jucar River Basin drought indicator system,
287 representing almost 50% of the indicators value. We transform the reservoir levels

288 probability distribution into state index distributions following the calculation method above
289 by comparing the results to the historic series of observed levels. Afterwards, we determine
290 the probability of scenario change for each month by crossing each state index distribution
291 by the threshold levels defined by the state index methodology.

292 We applied this methodology in the Jucar River Basin for the period between hydrologic
293 years 1998-1999 and 2008-2009. During these 10 years, two of the most important drought
294 episodes for the Jucar River Basin in history took place (CHJ 2007; van Lanen et al. 2013):
295 the short but intense drought of 1999-2000 and the long drought episode between 2005 and
296 2008.

297 The optimisation process tends to empty the reservoirs by the end of the optimisation period.
298 Thus, setting the multiple risk assessment runs for just one year would not provide adequate
299 results since we want to make use of the perfect forecast principle of optimisation.
300 Therefore, optimisation periods of four years were used for each run extracting the results of
301 the first year. Three hundred series of 48 months generated with the autoregressive model
302 from equation 3 proved sufficient to yield representative results in the Monte Carlo
303 optimisation process for each monthly run.

304 **4. Results**

305 4.1. State Index complementation with scenario change probability

306 Figure 5 shows the result of applying the proposed methodology together with the evolution
307 of the Jucar River observed state index for the three first years of the optimisation period
308 considered in each run. The fourth year is disregarded because it coincides with the end of
309 the optimisation period, when the algorithm uses all the available water. For each month, we
310 have the actual drought scenario as defined by the thresholds and the probability of each

311 scenario occurring one to three years later corresponding to Figures 5a to 5c, respectively.
312 In Figure 5a, the probability of a scenario change in the next year is low, with a general
313 tendency to remain at the same level. In Figures 5b and 5c, the probabilities of a scenario
314 change increase after two and three years and how this provides a better insight of what can
315 be expected in the system. With these results, the methodology proposed adds value to the
316 actual State Index by showing the probability that the current situation might change in the
317 future, hence providing additional support for decision makers in terms of activating
318 mitigation measures, which normally require some time to start operating appropriately.

319 The probability of scenario change with one year anticipation (Figure 5a) is useful for the
320 middle and end of drought episodes as well as for annually operated systems. For example,
321 soft preventive measures could have been maintained in February 2001 despite the
322 entrance in the normality scenario in order to prevent the posterior quick fall to almost
323 emergency one year later. Conversely, the two and three year anticipation probabilities
324 (Figures 5b and 5c) are useful in detecting the possible start of a drought situation,
325 especially in multi-year systems. Between 2004 and 2008, the State Index dropped from the
326 normality scenario to emergency in about one year (June 2004 to June 2005) and then
327 remained in that situation for two years. This situation is captured in Figures 5b and 5c,
328 where the probabilities of being in a scenario worse than normality two and three years after
329 June 2004 exceeded 50%.

330 4.2. Approximation of SI values with risk results

331 Previous stakeholder participation experiences in the Jucar River with risk assessment tools
332 have shown that, in general, risk results obtained for an 80% probability of exceedance level
333 and one year in advance are trusted as good approximations of the future state of the
334 system. These results can be easily extracted from the tools used to perform the proposed

335 methodology, as well as any other risk level results. Hence, we explored the ability of the
336 proposed methodology to approximate SI from a probabilistic perspective.

337 Figure 6 shows the evolution of SI approximated as the 80% risk level one year in advance
338 versus the actually observed SI in the Jucar River for the period October 1998 through
339 September 2009. Both indices reflect accurately the drought events occurred in the Jucar
340 River basin for the period of study. However, while the risk based SI follows the observed
341 one during the first part of the period, there is a six months delay disconnection right before
342 the beginning of the 2004-2008 drought episode. This is due to the operation of the
343 optimisation process. The objective function in the optimisation model works tries to
344 maximise the stored volume in reservoirs while meeting all the demands and environmental
345 flows, minimising water loses from the system. First, during the wet period prior to the 2004-
346 2008 event, the optimisation model achieves better storage levels before the episode starts
347 because all the demands are met and there is water that would be lost instead at a high cost
348 for the objective function. Since the optimisation process implies perfect forecast, the model
349 is capable of storing that water. Second, when reservoirs are near to empty, like during the
350 drought period, the objective function benefits more from supplying the demands than from
351 storing water. Hence, despite the risk based SI drops below the observed one, the demands
352 still have a better level of supply than in the real situation. Therefore, the risk based results
353 offer an envelope of the actual situation, providing managers with an idea of how the system
354 can be expected to respond at different levels of risk.

355 **5. Discussion**

356 The predictions of the methodology presented improve with respect to the combined use of
357 storage, streamflow and precipitation to define a drought state index because they include
358 both previous precipitation and storage data, as well as information regarding the physical

359 system what allows obtaining its best management options. It also includes up to date
360 information of the human influence on the system by means of water demands for the
361 different sectors, and allows considering the environmental needs of the riverine ecosystems
362 in the form of environmental flows definition. In addition, the presented methodology can be
363 used afterwards to assess the risk level with the existing management rules to evaluate the
364 changes introduced by the mitigation measures. Since the methodology is meant to be used
365 every month to monitor the state of the system, any new measures could be implemented in
366 the model in real time. In this way, it is possible to select the best measures for each case
367 and their optimal application.

368 5.1. Methodological limitations

369 The methodology has a number of inherent limitations. Firstly, it was limited by the quality of
370 the stochastic streamflow series used to drive the whole process. The definition of a good
371 stochastic model requires an amount of previously observed data that is not always going to
372 be available. In addition, depending on the stochastic model used, the generated streamflow
373 series will have a different capacity of capturing the dynamics of hydrology in the system.
374 This, together with the tendency of stochastic series to reach values around the historic
375 average after a number of generations, will limit the risk forecasting ability of the method. In
376 this paper, an autoregressive AR(1) stochastic model was used. Despite being capable of
377 capturing the basic statistical parameters of the observed series, Ochoa-Rivera et al. (2007)
378 showed that the approach to streamflow modelling has a significant influence in the final
379 results. Hence, different modelling methodologies should be explored before implementing
380 the proposed methodology.

381 Secondly, optimisation is a highly resources consuming process. This means that complex
382 models of the system under study will require longer calculation periods than more simple
383 ones. The creation of models capable of representing the reality of the system while

384 maintaining a low degree of computational complexity requires a high level of knowledge
385 and understanding about the system. The Jucar River Basin has been extensively studied by
386 researchers for many years, and the methodology presented here was relatively easily
387 applicable. However, it will not be of immediate use in river basins where water level is
388 scarce and/or the relationships between the individual hydrological processes are not clear.

389 Finally, in order to be effective, the methodology and its results must be trusted, but also
390 understood, by those that will be later affected by the decisions derived from its use. The
391 model used in this study was developed conjunctively with the managers and water users of
392 the basin within a participatory process that required reaching agreements for everyone. In
393 the same way, the triggers that define each drought situation and the corresponding
394 measures are the results of negotiations between the different actors in the system. This
395 trust building process is achieved over time and thus, methods such as the one presented
396 here are unlikely to be successful at the beginning of participative management processes.

397 Anyway, as observed in Andreu et al. (2009) and Andreu et al. (2013), the very process of
398 implementing similar methodologies finally resulted in better knowledge of the system and
399 understanding of stakeholders needs with an overall improvement of management.

400 5.2. Implications for drought management

401 Existing drought monitoring systems are normally limited to measure a series of climatic and
402 hydrologic variables and calculating various indices that allow determining what is the state
403 of the system compared to the past. Such is the case of the state index used in Spanish
404 drought management plans shown above. This approach may be useful, if not the only one
405 possible in some cases, but has been revealed insufficient for its use in some systems,
406 especially regulated water resources systems (Haro et al. 2014). Using indicators based on
407 observation of hydrologic variables, and comparison with past data in systems where human
408 activities take place, are unable to represent the changes occurring in the system along time.

409 Anthropogenic actions influence not only river flows themselves with extractions and returns
410 but also runoff production and groundwater recharge, delaying or preventing water from
411 reaching the streams. Accounting for all of this and translating observed flows in one point to
412 natural regime is often an arduous task that is not always rewarded with appropriate results.
413 In addition, the parameters used for drought indices calculation are variable with time. This
414 causes that new maximum and minimum observed values have the chance to change
415 dramatically the shape of the indicator evolution. For example, if an exceptionally wet, or dry,
416 period occurred, several hydrological variables (precipitation, streamflow, reservoir storage
417 levels, etc.) could reach unprecedented levels that might change the values of the state
418 index resulting in completely erroneous impressions regarding past drought events, as well
419 as influencing the perception of future ones.

420 In regulated systems, the volume stored in the different reservoirs of the system, especially
421 the regulation reservoirs, is normally regarded as a good approximation of the actual status
422 of the whole system. Moreover, the comparison between the storage levels at the beginning
423 and the end of the hydrologic year are commonly accepted as a summary of how the
424 management of the system has been. However, the volumes stored nowadays are not
425 comparable with the volumes stored, for example, ten years ago since water uses in the
426 system change over time. This makes that the behaviour of the system, and thus the storage
427 in reservoirs is different should the new demands were considered and indicators such as
428 the one used by river basin districts in Spain cannot reflect that. In addition, the existence of
429 high risk levels of developing drought scenarios during normality situations raise concern
430 about the need for a more appropriate definition of what is considered to be normality in a
431 water resources system. For this, it is undoubtedly necessary to have a deep knowledge
432 about the system. The use of both simulation and optimisation models allow enhancing the

433 knowledge that managers and users have of the system as well as building common
434 understanding on the needs and concerns of the different actors involved.

435 Finally, following a drought preventive strategy in a water resources system needs
436 maintaining a continuous state of vigilance. Hence, drought monitoring systems should warn
437 of the risk that a certain situation, that is considered to involve risk, develops into a worse
438 scenario instead of just informing about the current state of the system. In this way, the
439 measures addressed to minimise the risk or mitigating the effects of a fully developed
440 drought episode would have enough time to operate and be efficient, and they could even
441 be less severe than when applied with urgency. Water resources systems management
442 involves some bureaucracy and it is necessary to take into account that the activation of
443 measures normally will take some time after the declaration of a new drought scenario.
444 Thus, being able to anticipate the state of the system in a way like the one presented in this
445 work can definitely help improving the performance of drought plans.

446 **6. Conclusions**

447 This paper has proposed a new methodology to support drought monitoring and scenario
448 definition in regulated water resources systems. It allows approaching droughts risk
449 assessment and early warning from a new perspective with regard to previous approaches,
450 adding value to the existing monitoring methods currently in use. The use of optimisation
451 modelling to obtain the best management of the system during uncertain hydrologic periods
452 such as droughts permits anticipating the possible outcomes of these situations without the
453 need of considering the operation rules in place that might result ineffective in these cases.
454 An important advantage of the method developed is its capacity for dealing with complex
455 systems, providing a general picture of the situation in the basin while most of the previously
456 developed indices are applicable only to a demand or to a group of demands. Thus, the

457 proposed method constitutes a step forward in the definition of drought early warning
458 systems in regulated basins. The application of the methodology in the Jucar River shows its
459 potential for supporting the definition of drought scenarios and hence improving the overall
460 drought management process in the basin. Furthermore, the methodology proposed is easily
461 exportable to other cases of study since it makes use of generalized modelling tools freely
462 available online, although it is important to keep in mind that it is necessary a good
463 knowledge of the system in order it to be effective.

464 Since no drought is identical to another, especially given a changing climate, modelling the
465 past to anticipate future drought is an ineffective and risky approach. Including future
466 changes in climate and hydrology is essential, but also future water demands and operation
467 policies must be considered in order to attain useful and reliable results for an efficient
468 anticipation to future drought events. Different operation policies may also require different
469 approaches with regard to drought management, both in the definition of scenario thresholds
470 for measures activation and the variables monitored, and the tools necessary to support
471 decision making.

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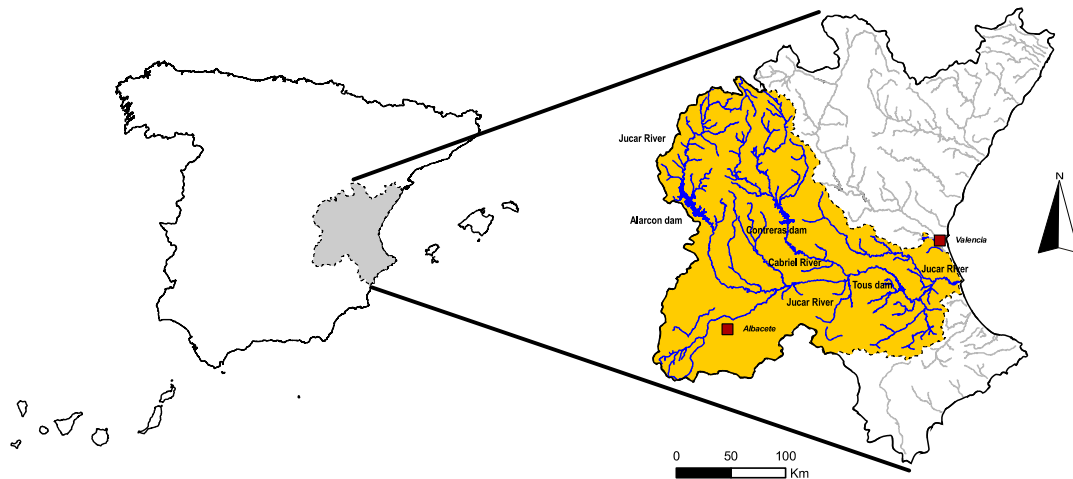


Figure 1. Location of the Jucar River Basin in Iberian Peninsula and within the other systems in CHJ

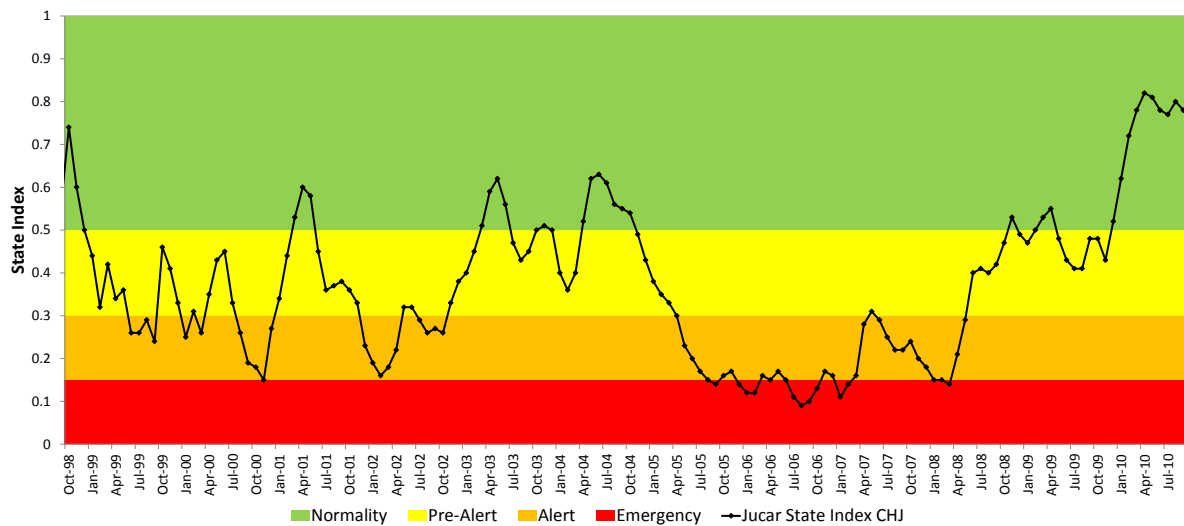


Figure 2. Evolution of the Jucar River Drought State Index with drought scenario thresholds between years 1998 to 2010

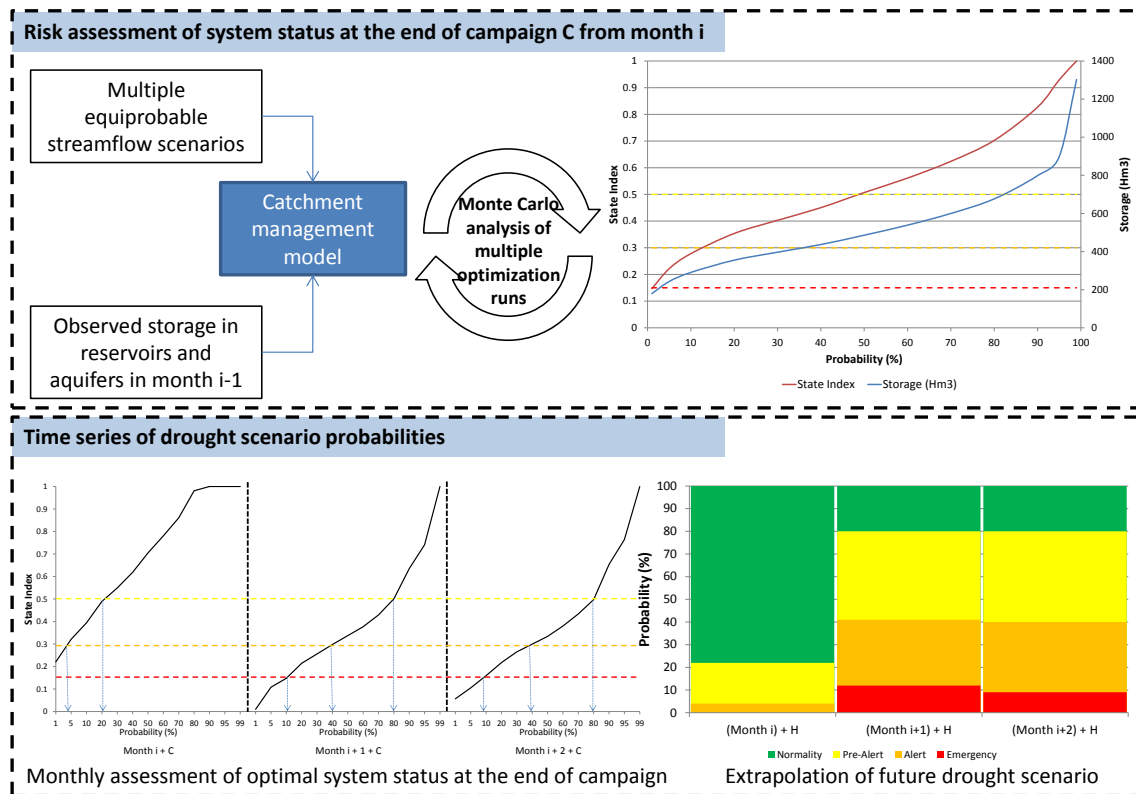


Figure 3. Schematic of the methodology for the definition of future drought scenarios

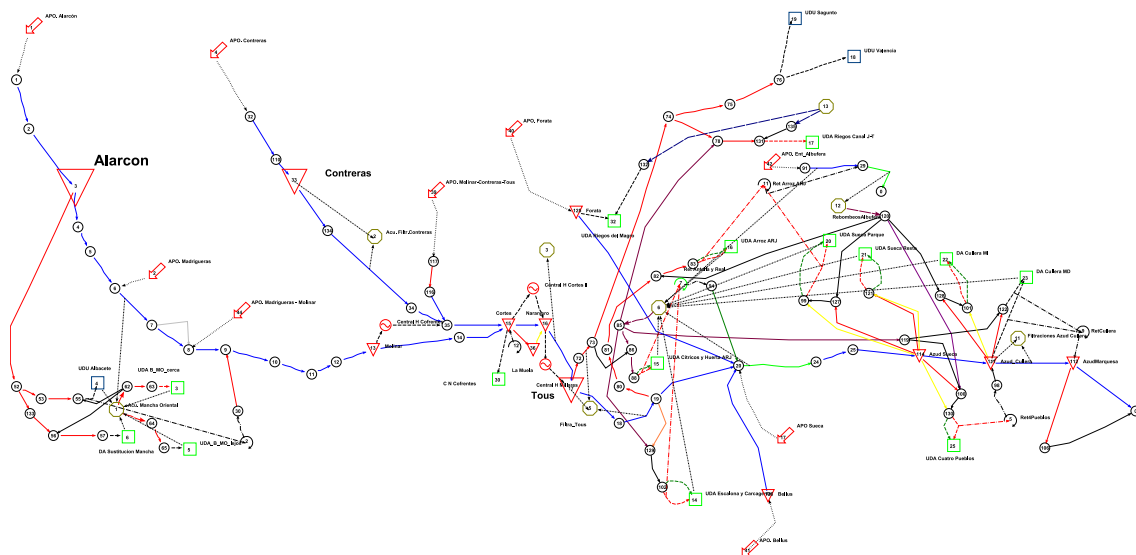


Figure 4. Scheme of the Jucar River Management Model

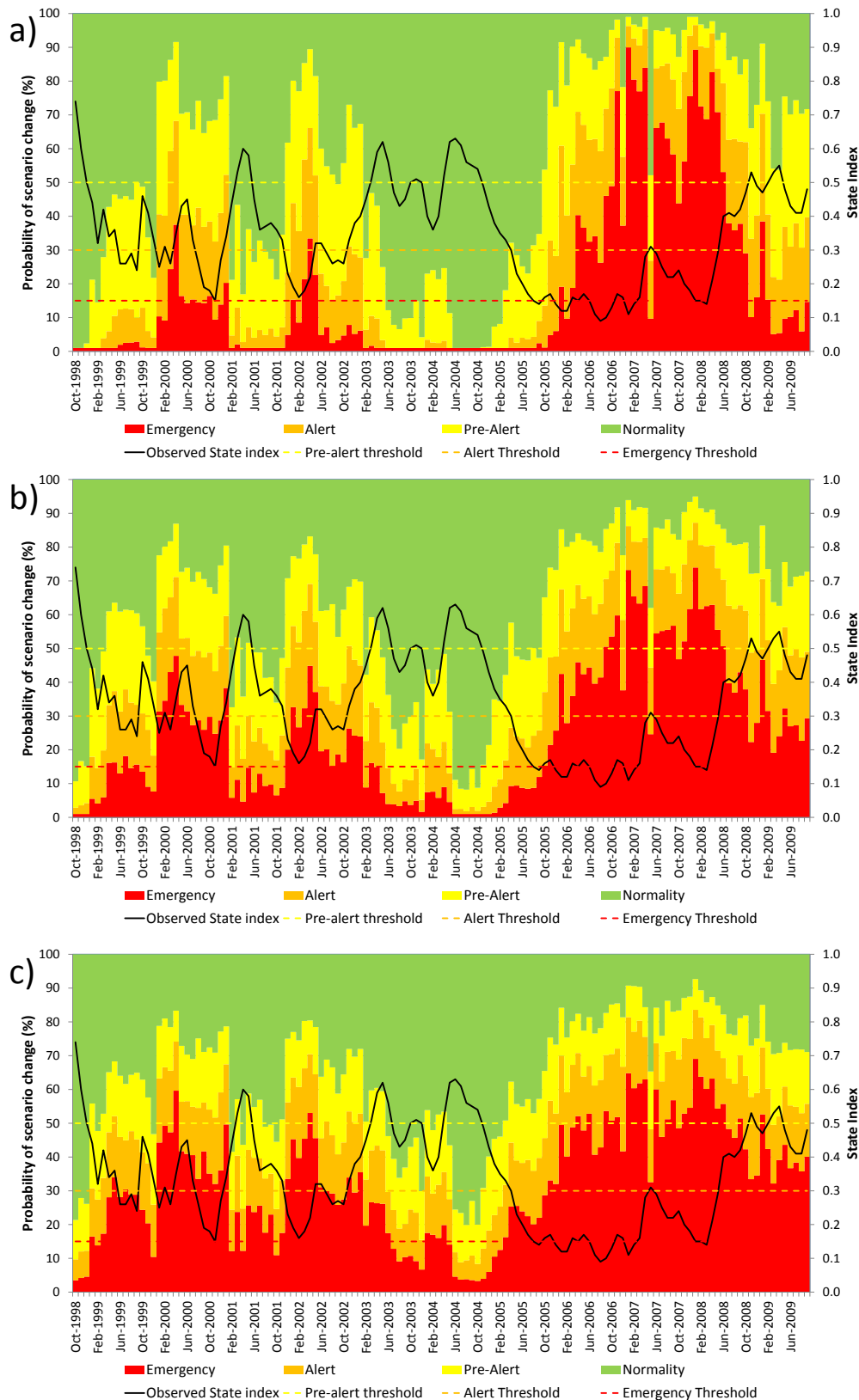


Figure 5. Probabilistic risk scenario definition vs Observed State Index in the Jucar River Basin at the end of (a) one, (b) two, and (c) three campaigns

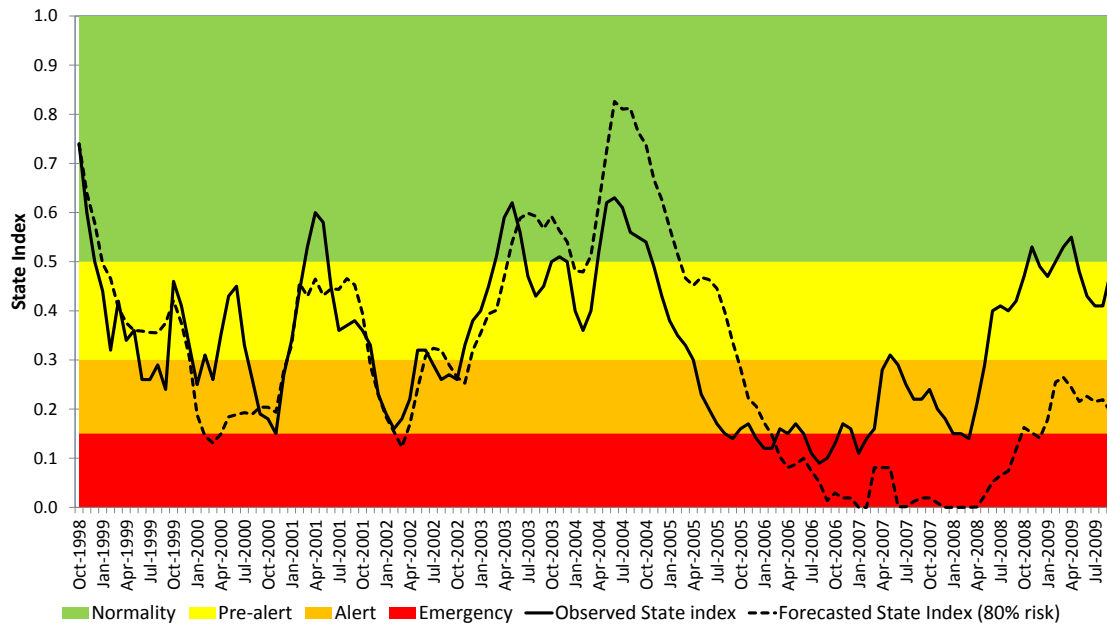


Figure 6. Observed state index values at the Jucar River versus calculated state index at 80% risk level

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