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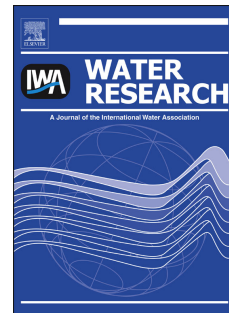
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1 ***Low energy ballasted flotation***

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9

10 ***Abstract***

11 A novel process which involves the replacement or supplementation of bubbles in the dissolved
12 air flotation process with low density beads is presented. The work comprised a series of bench
13 scale flotation trials treating three commonly encountered algal species (*Microcystis*, *Melosira* and
14 *Chlorella*) that were removed in a flotation cell configured as either: conventional dissolved air
15 flotation (DAF); ballasted flotation using low density 70 micron glass beads with a density of 100
16 kg.m⁻³; or a hybrid process of ballasted flotation combined with conventional DAF. Results
17 indicated that the bead only system was capable of achieving better residual turbidity than
18 standard DAF at bead concentrations of 500 mg.L⁻¹. Addition of beads in combination with
19 standard DAF reduced turbidity further to even lower residual turbidity levels. Algae removal was
20 improved when glass beads were dosed, but removal was dependent on algal species. *Microcystis*
21 was removed by 97% for bead only systems and this removal did not change significantly with the
22 addition of air bubbles. *Melosira* was the next best removed algae with bead only dosed systems
23 giving similar removals to that achieved by standard DAF using a 10% air recycle ratio (81 and
24 76% removal respectively). *Chlorella* was the least well removed algae by bead only systems
25 (63% removal). However, removal was rapidly improved to 86% by the addition of air bubbles
26 using only a 2% recycle ratio. Energy estimations suggested that at least a 50% energy reduction
27 could be achieved using the process offering a potential route for future development of low
28 energy separation processes for algae removal.

29

30 **Keywords:** Algae, Bubbleless, Dissolved air flotation, Energy,

31

32 **Introduction**

33 Dissolved air flotation (DAF) is an established solid-liquid separation technology process
34 in water treatment for removal of low density floc including those containing algae or
35 dominated by natural organic matter (NOM) and in low temperature countries (Schofield,
36 2001). In the DAF process, floc formed in preceding coagulation and flocculation stages
37 are separated from water by the attachment of bubbles onto the floc. The bubble-floc
38 aggregate becomes less dense than water and therefore floats to the top in a flotation tank
39 forming a sludge blanket. Clarified water exits the tank from beneath the float, whilst the
40 sludge blanket is periodically removed from the top. A key component in any DAF
41 system is the generation of micro-bubbles by saturating air with water. During saturation,
42 between 5-15% of the clarified flow is recycled and mixed with air supplied by a
43 compressor. The air-water mixture is then pressurised to between 400-650 kPa to
44 dissolve the air into the water. The pressurised air-water mix is then introduced into the
45 flotation tank at atmospheric pressure through nozzles. As a result of the release of the
46 pressure drop, the excess air precipitates out in the form of bubbles that are typically
47 between 40-100 μm (AWWA, 1997). A benefit of the system is in its ability to adjust to
48 water quality and solid concentration changes by altering the number of bubbles released
49 by changing the recycle ratio enabling changes in the particle loading to be effectively
50 matched by addition of more or less air bubbles.

51 As well as being a large capital cost, the saturator and recycling systems account for
52 approximately 50% of the operating costs of a DAF system (Haarhoff and Rykaart,
53 1995). This is principally from the electrical energy consumption of around $0.3 \text{ kWh}\cdot\text{m}^{-3}$
54 of treated water for the operation of the compressor of the saturator and the pumping of
55 the recycling system (Viitasaari *et al.*, 1995). Consequently significant saving could be

56 made if the need for bubble generation could be removed. A bubbleless system may be
57 achieved using the concept of ballasted flotation. In ballasted flotation, a low density
58 material is incorporated into the floc to give the aggregate an overall density less than
59 that of water so that the particle floats without the need for bubbles to be attached. Low
60 density materials that could be used include a range of commercially available hollow
61 spheres composed of latex or glass or solid particles that float in water (composed of a
62 material such as polystyrene). This concept is described in two patents: WO/2006/008474
63 and US Patent 6890431 but there is no other published research on the process. Analogy
64 of the ballasted flotation concept can be made with sedimentation systems where floc
65 densities are increased by adding ballasting agents of high density. Examples of
66 ballasting agents include activated carbon, recycled sludge (Landon *et al.*, 2006),
67 magnetic particles (Booker *et al.*, 1996) and sand (Plum *et al.* 1998). The latter is perhaps
68 the most commonly adopted version under the trade name Actiflo[®] and is used for a
69 range of applications including tertiary treatment of sewage, intermittent discharges and
70 potable water treatment (Guibelin *et al.*, 1994; Imasuen *et al.*, 2004). Similarly, the
71 advantage of using a low density ballasting agent could have the equivalent effect of
72 enhancing flotation (in combination with bubbles) or replacing the need for bubbles
73 entirely resulting in a significant energy reduction for the flotation process. Ballasted
74 flotation could be used in all applications where standard DAF is currently used, such as
75 treatment of waters that are dominated by algae or NOM. The work presented here
76 investigates the practical feasibility of ballasted flotation by examining the efficacy of
77 implementing recyclable low density beads to replace the bubbles used in DAF in bench
78 scale jar test trials for removal of particles from water spiked with algae.

79

80 ***Materials and Methods***

81 A series of bench scale jar tests was carried out to determine the feasibility of using low
82 density glass beads for flotation of floc dominated by algae. Tests were carried out in one
83 of two formats: 1) **Traditional DAF** and 2) **Ballasted flotation**.

84 *Traditional DAF*: Jar tests were carried out using a model DBT6 DAF batch jar tester
85 (EC Engineering, Canada). The DAF jar tester operates in a similar way to a standard jar
86 tester during the floc formation stage using 1 L samples of water contained in 1 L square
87 beakers. Water was rapid mixed for 1 minute at 200 rpm followed by a slow stir period at
88 30 rpm for 15 minutes. For flotation of floc, the DAF jar tester adds pressurised water
89 saturated with air into the jar through diffusers enabling bubbles to form that can attach to
90 the flocs and float them to the surface of the jar. The amount of air saturated water added
91 into the jar was varied from 0-10% of the 1 L sample in the jar (referred to as the recycle
92 ratio). The 0% recycle ratio represented a sedimentation system because no air bubbles
93 were dosed into the system to enable flotation to take place. Water was sampled from a
94 sampling tap a third of the way up the jar after 10 minutes of flotation. For each jar test,
95 samples were analysed for turbidity using a Hach 2100 turbidimeter after 10 minutes of
96 flotation following the addition of air bubbles into the jar tester.

97

98 The water tested was from a lowland reservoir from the east of the UK. The water had a
99 turbidity of 6.5 ± 1.7 NTU. Water was coagulated using ferric sulphate (Ferrisol XL, EA
100 West) at a dose of 3.5 mg.L^{-1} as Fe at pH 5.5 (a pre-determined optimum for this water).

101 Initial testing was carried out on the raw water. Subsequent tests to determine the
102 effectiveness of low density beads on algae removal were carried out by separately
103 spiking raw waters with three different algae species: *Microcystis* (cyanobacteria);
104 *Melosira* (diatomaceous algae); *Chlorella* (green algae). Algae were cultured in nutrient
105 rich Jaworski medium in sterile beakers at 15°C in a continuous light environment. The

106 water was spiked with algae to simulate bloom concentrations at between $0.5-1.0 \times 10^6$
107 cells.L⁻¹. Algae were enumerated using a Neubauer hemocytometer before and after
108 flotation. The number of fields of view required to count 100 individual algal cells for a
109 specific magnification was measured and equated to the volume of water contained in the
110 hemocytometer for each field of view.

111 *Ballasted flotation:* Low density glass beads from Trelleborg, Emerson and Cuming Inc
112 (Mansfield, USA) were used in flotation tests as provided by the manufacturer.
113 Manufacturer information reported the beads having a median size of 70 μm and a
114 density of 100 kg.m^{-3} . The beads were dosed into the water before the coagulant was
115 added and mixed briefly to disperse in the jar at concentrations between $100-900 \text{ mg.L}^{-1}$.
116 The jar test was then carried out as described before for recycle ratios between 0-10%. In
117 this case the 0% recycle ratio was a flotation test because the beads in the floc reduced
118 the density of the aggregate to below that of the water. To determine whether the beads
119 could be effectively re-used after coagulation, the bead-floc float was broken up by
120 rapidly mixing on the jar tester to separate the two at 200 rpm for 1 minute. The mixing
121 was stopped and the beads that floated to the top of the jar after 10 minutes were
122 collected and re-used in a subsequent jar test using the previously described coagulation
123 procedure. This was repeated five times.

124 The particle size distribution (PSD) of the beads used in this work was validated using a
125 Malvern Mastersizer (Malvern Instruments, UK). Beads were added into 1 L of de-
126 ionised (DI) water in a 1 L square beaker at a concentration of 300 mg.L^{-1} . The beads
127 were mixed on a jar tester at 200 rpm and pumped through the optical unit of the
128 Mastersizer and back into the jar. An average of three measurements was used to provide
129 the final PSD. The size of the flocs formed on the jar tester with and without glass bead
130 addition were also measured using the Mastersizer instrument. The suspension was

131 monitored by drawing water through the optical unit of the Mastersizer and back into the
132 jar by a peristaltic pump on the return tube using 5 mm internal diameter peristaltic pump
133 tubing at a flow rate of 1.5 L.hr⁻¹. Size measurements were taken every minute for the
134 duration of the jar test and logged onto a PC.

135 Modelling floc sedimentation and rise rates was carried out using Stokes' law. There is
136 some uncertainty in using Stokes' law for flocs due to their porous and irregular structure
137 but the application provides a useful relative comparison and is widespread in floc
138 analysis (Bache *et al.*, 1991; Gregory, 1997; Tang *et al.*, 2002). In this analysis, floc were
139 assumed to be spheres consisting of i) flocculated matter (algae and coagulant
140 precipitates) and ii) glass beads with a diameter of 70 µm. The density of the flocculated
141 matter was modelled between 1010-1060 kg.m⁻³. These density ranges were selected
142 based on literature values for different types of floc (1038-1065 kg.m³ for activated
143 sludge flocs (Sears *et al.*, 2006); ferric hydroxide floc density estimated 1050 kg.m⁻³
144 (Bastamante *et al.*, 2001); algae floc modelled as 1020 kg.m⁻³ (Haarhoff and Edzwald,
145 2001)). Glass bead density was taken as 100 kg.m⁻³ from manufacturer data.

146 **Results**

147 The performance of the system was dependent on both the bead concentration and the
148 equivalent recycle ratio applied (Figure 1). In the case of traditional DAF, the residual
149 turbidity ranged from 1.7 to 4.2 NTU as the recycle ratio decreased from 10 to 2% (0
150 mg/L bead concentration, Figure 1). Addition of beads to the system resulted in either no
151 change or a slight decrease in turbidity except at high bead doses and low recycle ratios
152 (Figure 1). For instance, at a recycle ratio of 6% the residual turbidity with no beads was
153 2.6 NTU and ranged between 1.4 and 2.9 NTU for bead concentration between 100 and
154 900 mg.L⁻¹. For ballasted flotation, the application of beads without the use of any
155 bubbles (0% recycle ratio) resulted in residual turbidities between 2.4 NTU at 600 mg.L⁻¹

156 and 5 NTU at 200 mg.L⁻¹ indicating that the use of beads and no bubbles approached the
157 performance of traditional DAF (Figure 1). Closer inspection of the residual water
158 revealed beads remained in the water which reflected a distribution of properties
159 observed in the beads and the fact that no pre-conditioning was conducted. This
160 observation was confirmed by recovering the floated beads and reusing them on
161 consecutive application (Figure 2). After the first use of the beads at 500 mg.L⁻¹, the
162 residual turbidity was 3.5 NTU, this was reduced to 1 NTU after the fifth use of the same
163 beads. This was below that achieved for a system at 10% recycle ratio without any beads
164 added (1.7 NTU) showing that the beads could be effectively recycled and that, in fact,
165 the bead system was capable of working better than traditional DAF after the beads had
166 been used three times or more. A 41 % improvement in residual turbidity was observed
167 using the bubbleless bead system (ballasted flotation, 0% recycle) compared to traditional
168 DAF after five uses of the bead (Figure 2). The observed improvement with multiple uses
169 reflects the removal of non floating beads due to imperfections in manufacture that lead
170 to thicker walls of the spheres than intended, increasing the density of the beads. In
171 addition, breakage of the spheres can also occur. Manufacturer data indicated that 1% of
172 the beads by volume may be expected to be failures that do not float. Given that glass
173 typically has a density of 2,200 kg.m⁻³ or above (Koike and Tomozawa, 2007), non-
174 floating beads will have a significant impact on overall floc density. However, removal of
175 such beads during a pre-conditioning process effectively negates the problem. In this
176 case, pre-conditioning was achieved through the multiple re-use of the same beads and
177 resulted in a system that generated a lower residual turbidity than traditional DAF.

178 Comparison of the efficacy of the ballasted flotation in relation to differing algal species
179 indicated a small difference in performance depending on the specific species tested. The
180 bubbleless bead process (0% recycle) was seen to be most effective for flotation of the

181 algae *Microcystis* resulting in 97% removal. Removal did not change significantly with
182 the addition of bubbles, fluctuating between 92 and 96%. Conversely, for systems with
183 no beads added, removal increased from 16 to 78% removal with increasing recycle ratio
184 from 0 to 10 %. *Melosira* was the next easiest algae to remove increasing from 81%
185 removal to 96% with increasing recycle ratio for systems with bead dosing. Of note, it
186 was evident that removal of *Melosira* for a bead dosed system and no air bubbles resulted
187 in slightly better removal than for no beads at 10% recycle ratio with values of 81 and
188 76% respectively. *Chlorella* was the most poorly removed algae when no bubbles were
189 added for bead dosed systems at 63% removal, however the addition of a small number
190 of bubbles (2% recycle ratio) increased removal up to 86%. This removal was
191 significantly above the level seen for non-bead dosed systems at the highest recycle ratio
192 of 10% which produced 70% removal.

193 The range of algae removal observed during traditional DAF operation was in a similar
194 range to that seen previously in operational DAF systems of between 80-98% (Markham
195 *et al.*, 1997). The differences in removal for different algae reflects the differences in
196 structure between species. All of the algae floc showed poor removal when clarification
197 was by sedimentation. This was particularly the case for *Chlorella* and *Microcystis* which
198 were only removed by <20% in a sedimentation system. Both of these algae exist as
199 small single celled spheres between 2-10 μm (Henderson *et al.*, 2008). *Melosira* is a
200 diatom that forms much larger long chain colonies. Diatoms also contain silica in their
201 cell walls which has a high density ($2200 \text{ kg}\cdot\text{m}^{-3}$). The combined effect of increased size
202 and density explains why *Melosira* was the best removed algae by sedimentation.
203 Regardless of this, algae flocs were much better removed by flotation processes, a
204 conclusion reached by other researchers (Teixeira and Rosa, 2006). *Microcystis*, a
205 cyanobacteria, is a very low density algae because it has a gas vacuole within the cell

206 structure which enables the algae to control its buoyancy in the water column. This
207 makes removal of floc containing *Microcystis* particularly amenable to removal by
208 flotation. However, for conventional DAF, these algae floc were poorly removed until 6-
209 10% recycle ratios. With glass beads, *Microcystis* floc were very well removed by
210 flotation without the need for any bubbles (0% recycle). For the algae without a vacuole
211 (*Melosira* and *Chlorella*), the very highest removals were seen involving a combination
212 of low density beads and air bubbles. This indicates that a combined effect of algae
213 structure, morphology and density has a significant impact on removal efficiency by
214 coagulation and clarification, a conclusion that is in agreement with numerous other
215 studies on particle flotation (Valade *et al.*, 1996; Henderson *et al.*, 2008).

216 The presence of beads in the algae coagulation systems aided the removal of algae for all
217 of the recycle ratios investigated and the different algae species. In addition to improved
218 flotation, the presence of small spheres may have increased the incorporation of algae
219 into the floc that resulted in fewer non-flocculated algae in the jar test. A high
220 concentration of small particles provides nucleation points for coagulant precipitates to
221 form around and encourage floc development and can promote enmeshment of algae
222 within the floc matrix. The addition of kaolinite and activated silica has been added for
223 this purpose to improve natural organic matter removal (Gregor *et al.*, 1997).

224 The average size of the floc for systems dosed with and without glass beads was
225 significantly different for the two systems (Figure 4a and b). Non-bead dosed systems
226 grew to a median floc size of 600 μm , reaching this size after 7 minutes of the jar test.
227 For systems dosed with beads, the flocs grew to a size that reached a maximum of 260
228 μm after 4 minutes of the jar test, but stabilised at 185 μm . As can be seen in the inset
229 image in Figure 4, numerous beads were observed to be incorporated into the algae-
230 coagulant floc with over 25 beads in the floc with a diameter of 500 μm . Given that the

231 maximum floc size was reached significantly before the end of the 15 minute flocculation
232 time for both systems in the jar test experiments, shorter flocculation times are advocated.
233 This is in agreement with other research suggesting that flocculation periods of 5-10
234 minutes are recommended for DAF (Edzwald, 1995).

235 The reduced floc size observed was an indication of reduced floc strength for floc
236 containing beads given that the steady state floc size has been shown to be an indicator of
237 floc strength (Yukselen and Gregory, 2004; Jarvis *et al.*, 2006). However, although there
238 was a difference in the average floc size for systems with and without beads added, it
239 should be noted that in conventional DAF, floc are exposed to high energy when they are
240 mixed with bubbles which breaks up the floc. The shear rates in DAF have been
241 estimated to be between 1000-7600 s^{-1} (Masschelein, 1992; Fukishi *et al.*, 1995). It has
242 been shown that the maximum floc size at shear rates of 1000 s^{-1} were between 30-281
243 μm which were formed from floc sizes of 600-1200 μm at 10 s^{-1} showing that floc size
244 was significantly reduced under the conditions prevalent in DAF (Bache and Rasool,
245 2001). Flocs formed in a bead dosed system and separated with no air bubbles added
246 would not be broken up because they would not be exposed to these high shear rates,
247 enabling floc to maintain their size as formed in the flocculator. The importance of this
248 relates to the breakage products formed, which includes the formation of floc around 1
249 μm in diameter. These sized particles cause significant operational problems because
250 they are poorly removed in downstream filtration processes. Limiting exposure of flocs to
251 high shear rates in flotation, as well as in the preceding coagulation and flocculation
252 stage, is particularly important for systems containing algae that may release toxins (such
253 as *Microcystis*) under high shear stresses (Edzwald and Wingler, 1990). The proposed
254 ballasted flocculation process would eliminate the need for the high shear rates used
255 today in most operational DAF plants.

256 One consequence of dosing glass beads into the system will be an initial increase in
257 sludge volume. However, because the beads will be removed from the sludge and re-
258 used, the volumes of sludge to be treated and disposed of will be the same as that for a
259 conventional DAF system.

260 The use of rise velocity modelling to establish the sensitivity with which bead properties
261 influenced performance indicated that the density of the coagulated material had little
262 impact on settling and rise rates at the mean floc size observed in the current study of
263 around 200 μm (Figure 5a). The theoretical settling rate of the flocs with no beads varied
264 from 0.08 $\text{m}\cdot\text{h}^{-1}$ and 0.47 $\text{m}\cdot\text{h}^{-1}$ for the lowest and highest floc densities used. A floc
265 containing 10 beads had a theoretical rise velocity of 2.8-3.0 $\text{m}\cdot\text{h}^{-1}$ with around 43% of
266 the total floc volume contributed from the bead. A floc containing 20 beads had a rise
267 velocity of 6.0-6.1 $\text{m}\cdot\text{h}^{-1}$ but would only contain 15% floc matter whilst above 23 beads,
268 the volume of the beads would exceed the volume of the complete 200 μm floc. As a
269 comparison to these modelled values, rise velocities for bubble-floc aggregates have been
270 measured as 3 $\text{m}\cdot\text{h}^{-1}$ for ferric hydroxide-algae floc (Vlaski *et al.*, 1997) for floc with an
271 average size of 15-20 μm . The rise velocities of activated sludge flocs were captured
272 between 1.8 and 37.8 $\text{m}\cdot\text{h}^{-1}$ with two thirds of the flocs measured having rise rates
273 between 5 and 15 $\text{m}\cdot\text{h}^{-1}$ (Ljunggren *et al.*, 2004).

274 The simple calculations have demonstrated that it is possible for floc containing beads to
275 have rise velocities similar to the range observed in other studies. Given the similar or
276 better turbidity removals observed for ballasted flotation (with no bubbles) when
277 compared with conventional DAF, it would be expected that the performance observed in
278 jar tests would be translated to continuous systems. The key is to ensure that enough
279 beads are incorporated into the floc to enable high rates of flotation and promote the
280 formation of large floc. For a 200 μm floc, the average floc size seen in this work, this

281 would require between 10-20 beads to be contained in the floc structure. If larger floc can
282 be formed and maintained it would be possible to generate flocs with theoretical rise rates
283 of $>40 \text{ m.h}^{-1}$ for floc $>500 \mu\text{m}$ containing over 300 glass beads (Figure 5b).

284

285 ***Discussion***

286 This bench-scale study has shown that using floating beads potentially offers an
287 alternative means of separating floc from treated water giving similar levels of residual
288 turbidity to conventional flotation systems with air bubbles. In principle, any coagulated
289 material (algae, activated sludge, NOM or minerals) could be floated from the system so
290 long as enough beads are incorporated into the floc aggregate to significantly reduce the
291 density of the floc below that of the water. A conceptual flow diagram of a how a
292 ballasted flotation system may be implemented at full scale shows the replacement of the
293 saturator with a hydrocyclone to recover beads and two additional pumps to transport
294 either recycled or fresh beads into the flocculation tanks (Figure 6). The reduction in
295 energy usage by removing the need for the saturator has two benefits: a direct saving in
296 money and a reduction in carbon footprint. Evaluation of the impact of such a system
297 requires accurate information about the energy usage of individual components within
298 water works which is currently not commonly available. Estimates for the energy used
299 for the saturation system of a typical DAF plant range between 0.1 and 0.3 kWh.m^{-3}
300 (Viitasari *et al.*, 1995) and this compares to around 0.003 - 0.02 kWh.m^{-3} for a typical
301 hydrocyclone (Vion, 2000). Even after the inclusion of pumps, the ballasted flotation
302 process should still enable at least a 50% reduction in energy to be generated when
303 compared with traditional DAF. To illustrate the potential impact of this, the energy
304 saving at a standard water treatment works operating at 50 ML.d^{-1} would be $1,825,000$
305 kWh.year^{-1} if it switched from traditional DAF to the ballasted flotation process

306 (assuming a saturator operating at $0.2 \text{ kWh}\cdot\text{m}^{-3}$ and a 50% energy reduction when using
307 floating beads with hydrocyclones and additional pumping). This equates to 196
308 $\text{tCO}_2\text{e}\cdot\text{year}$ or an annual cost saving of £127,750.

309 The current study was focussed on evaluating the potential of utilising beads to ballast a
310 flotation process at bench scale. The positive results presented then raise questions about
311 its implementation, most importantly: (1) what is the risk of beads entering the final
312 water and (2) how effectively can the beads be recycled and at what loss rate. The
313 presented work provides some evidence towards the first question: First use of the beads
314 resulted in high numbers of residual beads but subsequent use reduced this number
315 significantly demonstrating that appropriate pre-conditioning is essential and effectively
316 removes the problem. Further, given the bead size of $100 \mu\text{m}$, any beads carried over
317 with the clarified water will be captured within the downstream filtration processes
318 (Henderson *et al.*, 2008). Consequently, the possibility of bead carryover into the product
319 water is very low. The second question remains crucial. Whilst batch recovery of the
320 beads through high speed mixing within a jar tester worked effectively, translation into a
321 continuous process is important as the energy required to operate the plant and the bead
322 loss rate will define the overall economics of the process. In addition, whilst it is not
323 expected, further work is required to clearly demonstrate that ballasted flotation will not
324 increase cell lysis and increase the release of algogenic organic material, particularly in
325 relation to toxic compounds from *Cyanobacteria*. However, these results have
326 demonstrated that the ballasted flotation process appears to be very effective technology
327 for algae removal and could have much wider application in water, wastewater and
328 industrial solid-liquid separation processes.

329

330 **Conclusions**

331 Application of low density glass beads as a flotation ballasting agent effectively removes
332 the need for dissolved air in the flotation process. In the case of algae the efficacy of the
333 ballasting agent was related to the characteristics of the algae and was most effective for
334 *Microcystis* species. Floc diagnostics revealed that ballasted flocs were smaller than those
335 formed during the coagulation of algae. However, in practice these floc will not be
336 exposed to the higher shear rates of traditional DAF because of the removal of the
337 dissolved air injection stage. Floc breakage is therefore minimised, ensuring that the
338 concentration of residual turbidity in the clarified water is low and composed of larger
339 floc that will be more amenable to removal by filtration. Overall the use of beads
340 provides a low energy alternative to traditional DAF which can meet or exceed
341 performance and provide in-process and downstream benefits through extended filter run
342 times.

343

344 **References**

- 345 Bache, D. H., Hossain, M. D., Al-Ani, S. H., Jackson, P. J. (1991) Optimum coagulation
346 conditions for a coloured water in terms of floc size, density and strength. *Journal of*
347 *Water Supply: Research Technology- AQUA*, 9, 93-102.
- 348 Bache, D. H., Rasool, E. R. (2001) Characteristics of alumino-humic flocs in relation to
349 DAF performance. *Water Science and Technology*, 43 (8), 203-208.
- 350 Bustamante, H. A., Raj Shanker, S., Pashley, R. M., Karaman, M. E. (2001) Interaction
351 between cryptosporidium oocysts and water treatment coagulants. *Water Research*, 35
352 (13), 3179-3189.
- 353 Chowdhury, Z. K., Amy, G. L. (1991) Coagulation of submicron colloids in water
354 treatment by incorporation into aluminum hydroxide floc. *Environmental Science and*
355 *Technology*, 25, 1766-1773.

- 356 Degremont (2003) Integrated sludge thickening and lamellar separation performance in
357 Scottish water applications. *Filtration and Separation*, 40 (9), 22-23.
- 358 Edzwald, J. K. (1995) Principles and applications of dissolved air flotation. *Water*
359 *Science and Technology*, 31 (3-4), 1-23.
- 360 Edzwald, J. K. (2007) Developments of high rate dissolved air flotation for drinking
361 water treatment. *Journal of Water Supply: Research and Technology – AQUA*, 56 (6-7),
362 399-409.
- 363 Edzwald, J. K., Wingler, B. M (1990) Chemical and physical aspects of dissolved air
364 flotation for the removal of algae. *Journal of Water Supply: Research and technology –*
365 *AQUA*, 39, 24-34.
- 366 Feris, L. A., Rubio, J. (1999) Dissolved air flotation (DAF) performance at low saturation
367 pressures. *Filtration and Separation*, 31 (3-4), 61-65.
- 368 Fukishi, K., Tambo, N., Matsui, Y. (1995) A kinetic model for dissolved air flotation in
369 water and wastewater treatment. *Water Science and Technology*, 31 (3-4), 37-47.
- 370 Gregor, J. E., Nokes, C. J., Fenton, E. (1997) Optimising natural organic matter removal
371 from low turbidity waters by controlled pH adjustment of aluminium coagulation. *Water*
372 *Research*, 31 (12), 2949-2958.
- 373 Gregory, J. (1997) The density of particle aggregates. *Water Science and Technology*, 36
374 (4), 1-13.
- 375 Haarhoff, J., Edzwald, J. K. (2001) Modelling of floc-bubble aggregate rise rates in
376 dissolved air flotation. *Water Science and Technology*, 43 (8), 175-184.
- 377 Haarhoff, J., Rykaart, E. M. (1995) Rational design of packed saturators. *Water Science*
378 *and Technology*, 31 (3-4), 179-190.
- 379 Henderson, R., Parsons, S. A., Jefferson, B. (2008). The impact of algal properties and
380 pre-oxidation on solid-liquid separation of algae. *Water Research*, 42 (8-9), 1827-1845.

- 381 Jarvis, P., Jefferson, B., Parsons, S. A. (2006). Flocc structural characteristics using
382 conventional coagulation for a high DOC, low alkalinity ground water source. *Water*
383 *Research*, 40 (14), 2727-2737.
- 384 Koike, A., Tomozawa, M. (2007) IR investigation of density changes of silica glass and
385 soda-lime silicate glass caused by microhardness indentation. *Journal of Non-Crystalline*
386 *Solids*, 353 (24-25), 2318-2327.
- 387 Ljunggren, M., Jönsson, L., la Cour Jansen, J. (2004) Particle visualisation- A tool for
388 determination of rise velocities. *Water Science and Technology*, 50 (12), 229-236.
- 389 Markham, L., Porter, M., Schofield, T. (1997) Algae and zooplankton removal by
390 dissolved air flotation at Severn Trent Ltd surface water treatment works. In: Proceedings
391 of the CIWEM Dissolved Air Flotation International Conference, London, UK, April
392 1997.
- 393 Masschelein, W. J. (1992) Unit processes in drinking water treatment, Marcel Dekker,
394 New York.
- 395 Plum, V., Dahl, C. P., Bentsen, L., Petersen, C. R., Napstjert, L., Thomsen, N.B. (1998)
396 The Actiflo method. *Water Science and Technology*, 37 (1), 269-275.
- 397 Schofield, T. (2001). Dissolved air flotation in drinking water production. *Water Science*
398 *and Technology*, 43 (8), 9-18.
- 399 Sears, K., Alleman, J. E., Barnard, J. L., Oleszkiewicz, J. A. (2006) Density and activity
400 characterisation of activated sludge flocs. *Journal of Environmental Engineering*, 132
401 (10), 1235-1242.
- 402 Tang, P., Greenwood, J., Raper, J. A. (2002) A model to describe the settling behaviour
403 of fractal aggregates. *Journal of Colloid and Interface Science*, 247, 210-219.
- 404 Teixeira, M. R. and Rosa, M. J. (2006) Comparing dissolved air flotation and
405 conventional sedimentation to remove cyanobacterial cells of *Microcystis aeruginosa*.

406 Part I: The key operating conditions. *Separation and Purification Technology*, 52 (1), 84-
407 94.

408 Valade, M. T., Edzwald, J. K., Tobiason, J. E., Dahlquist, J., Hedberg, T., Amato, T.
409 (1996) Particle removal by flotation and filtration: Pretreatment effects. Consistent
410 performance of DAF and the quality of DAF effluent - Despite considerable variation in
411 flocculation characteristics and flocculated water quality conditions - Demonstrate the
412 robust nature of this process. *Journal of the American Water Works Association*, 88 (12),
413 35-47.

414 Viitasaari, M., Jokela, P., Heinanen, J. (1995) Dissolved air flotation in the treatment of
415 industrial wastewaters with a special emphasis on forest and foodstuff industries. *Water*
416 *Science and Technology*, 31 (3-4), 299-313.

417 Vion, P. (2000) US Patent 6277285 - Process for the clarification of liquids and
418 suspensions.

419 Vlaški, A., Van Breemen, A. N., Alaerts, G. J. (1997) The role of particle size and
420 density in dissolved air flotation and sedimentation. *Water Science and Technology*, 36
421 (4), 177-189.

422 *Water Treatment Plant Design*, Third Edition, 1997 American Water Works Association
423 and American Society of Civil Engineers. McGraw-Hill, New York.

424 Yukselen, M., Gregory, J. (2004) The reversibility of floc breakage. *International*
425 *Journal of Mineral Processing*, 73, 251-259.

426 Zakkour, P. D., Gaterell, M. R., Griffin, P., Gochin, R. J., Lester, J.N. (2002) Developing
427 a sustainable energy strategy for a water utility. Part I: A review of the UK legislative
428 framework. *Journal of Environmental Management*, 66 (2), 105-114.

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Figure 1. Residual turbidity for increasing bead concentration at different DAF recycle ratios after 10 minutes flotation. The coagulation conditions were 3.5 mg.L^{-1} Fe at pH 5.5. Raw water turbidity 6.5 ± 1.7 NTU spiked with algae at concentrations between $0.5\text{-}1.0 \times 10^6 \text{ cells.L}^{-1}$.

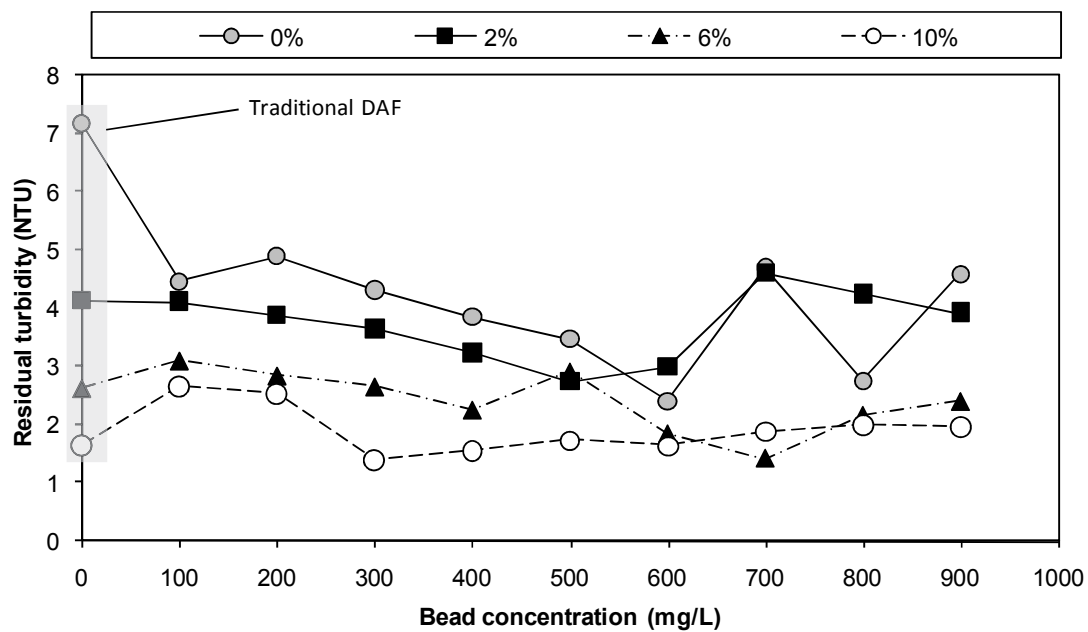
Figure 2. The residual turbidity of treated reservoir water after treatment with beads. The beads were dosed at a concentration of 500 mg.L^{-1} . No air bubbles were added into the system (0% recycle ratio). Coagulation conditions were 3.5 mg.L^{-1} Fe at pH 5.5. Raw water turbidity was 6.5 ± 1.7 NTU spiked with algae at concentrations between $0.5\text{-}1.0 \times 10^6 \text{ cells.L}^{-1}$.

Figure 3. Percentage removal of algae (from microscope counting) for *Microcystis*, *Melosira* and *Chlorella* algae species for increasing recycle ratios for systems with and without beads. Beads were dosed at a concentration of 300 mg.L^{-1} . The coagulation conditions were 3.5 mg.L^{-1} Fe at pH 5.5. Raw water turbidity was 6.5 ± 1.7 NTU spiked with algae at concentrations between $0.5\text{-}1.0 \times 10^6 \text{ cells.L}^{-1}$.

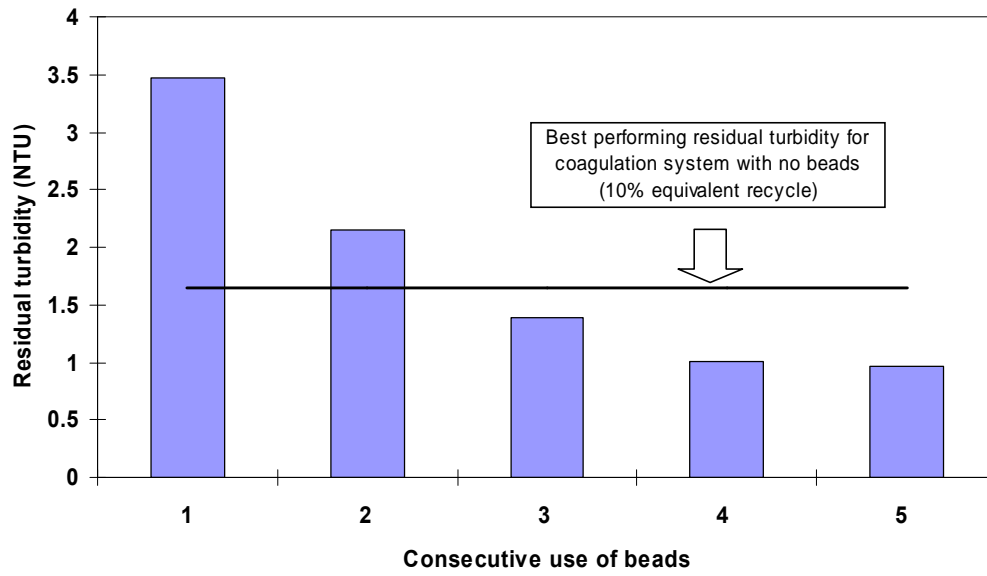
Figure 4a and b. Floc growth & PSD for coagulated systems with and without beads for water spiked with *Microcystis*. Bead concentration was 500 mg.L^{-1} and the coagulation conditions were 3.5 mg.L^{-1} as Fe at pH 5.5. Raw water turbidity was 6.5 ± 1.7 NTU spiked with algae at concentrations between $0.5\text{-}1.0 \times 10^6 \text{ cells.L}^{-1}$.

Figure 5. The change in floc settling/rise rates dependent on the number of beads in the floc and variable density (a) and floc size (b). a) Impact of the density of coagulated material (kg.m^{-3}) on settling/rise rates (SI 100 beads, floc size $200 \mu\text{m}$), b) Impact of floc size on settling/rise rates (floc size $200 \mu\text{m}$, density of coagulated matter 1020 kg.m^{-3}).

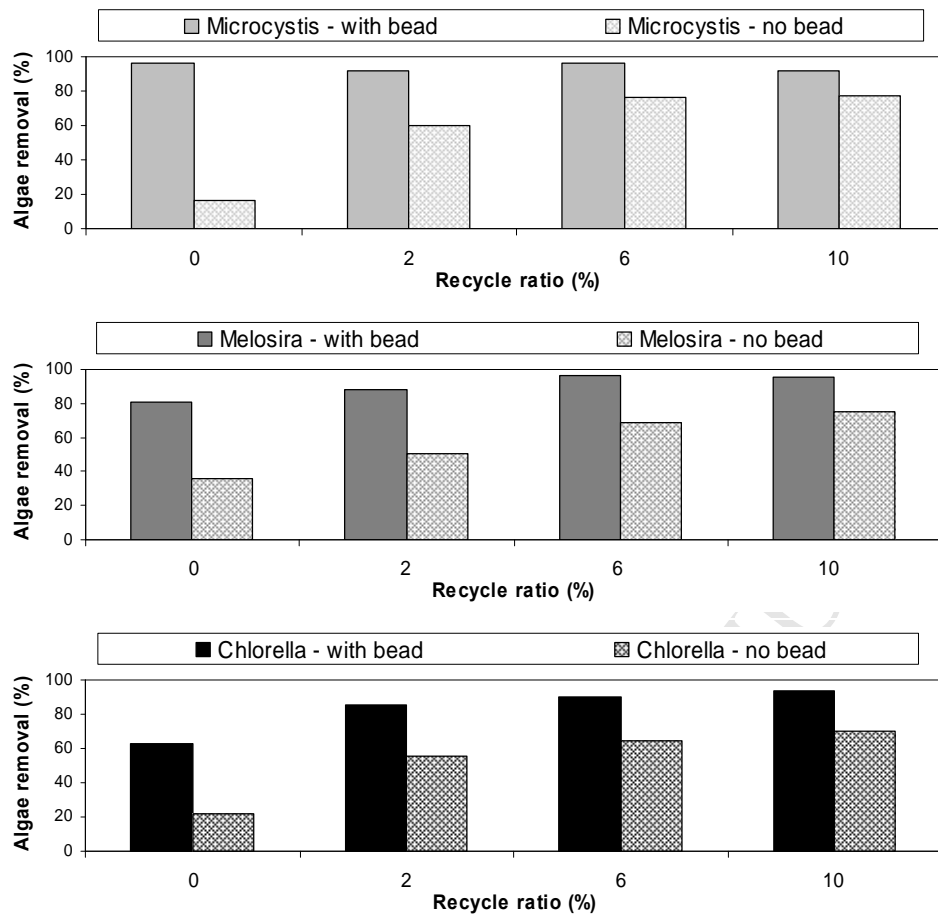
Figure 6. Conceptual schematic of the bubbleless flotation system.

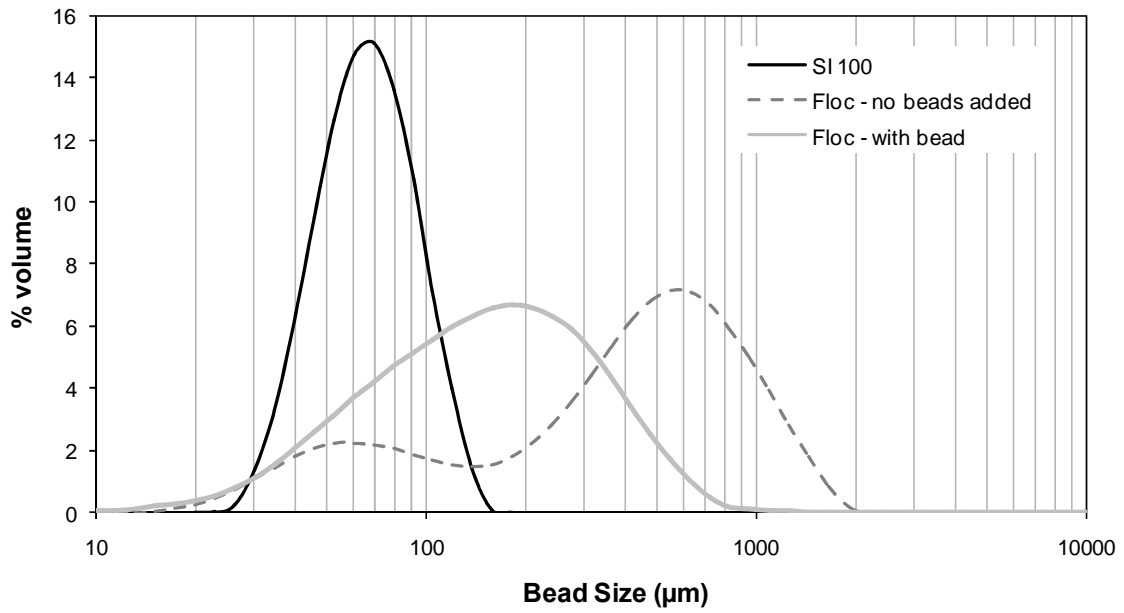
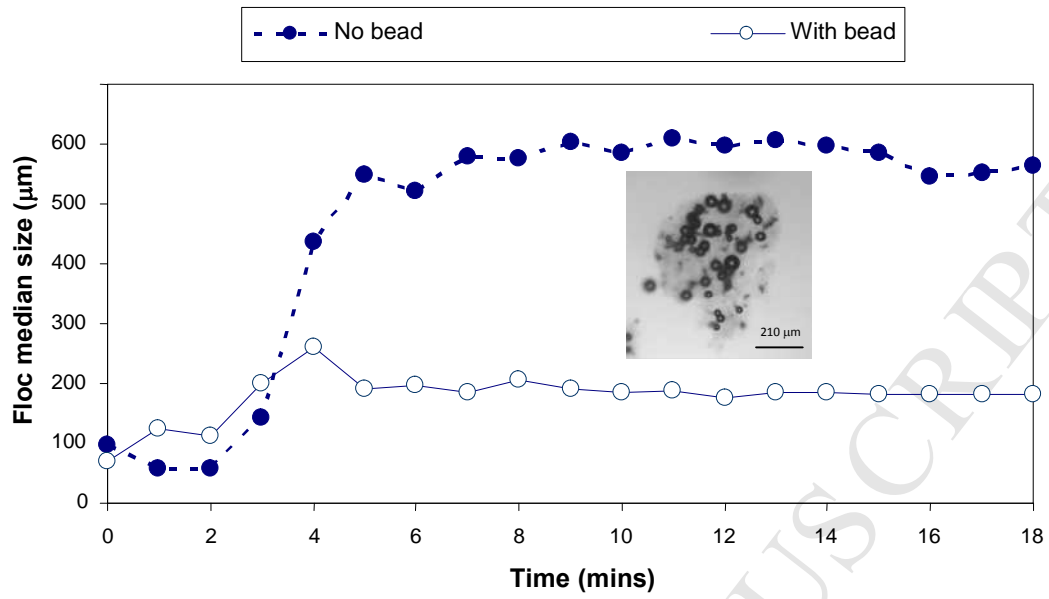


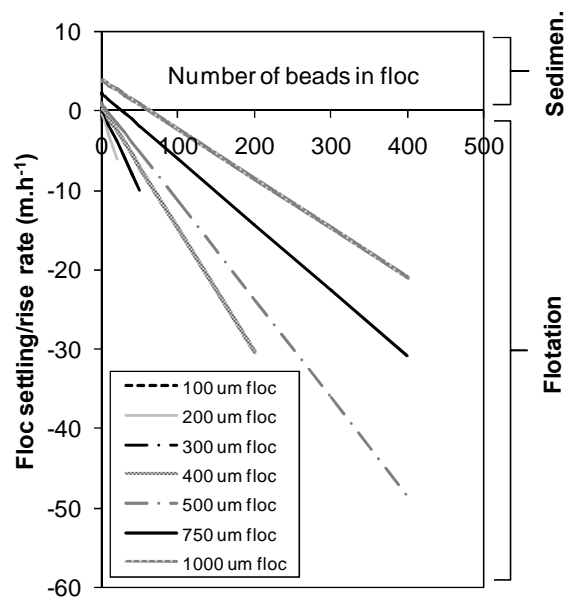
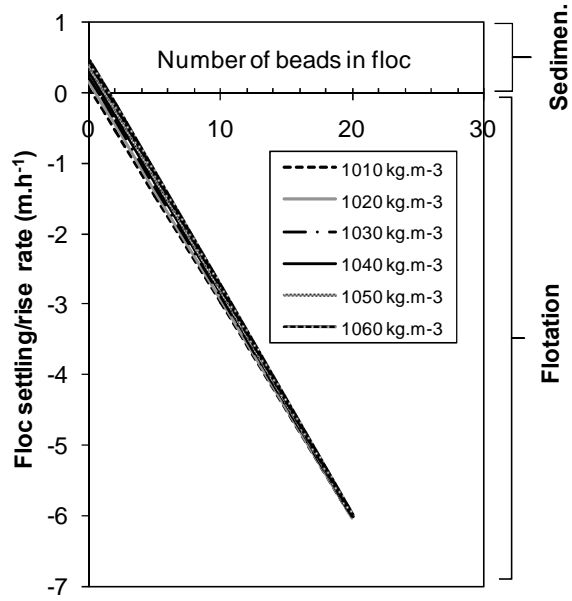
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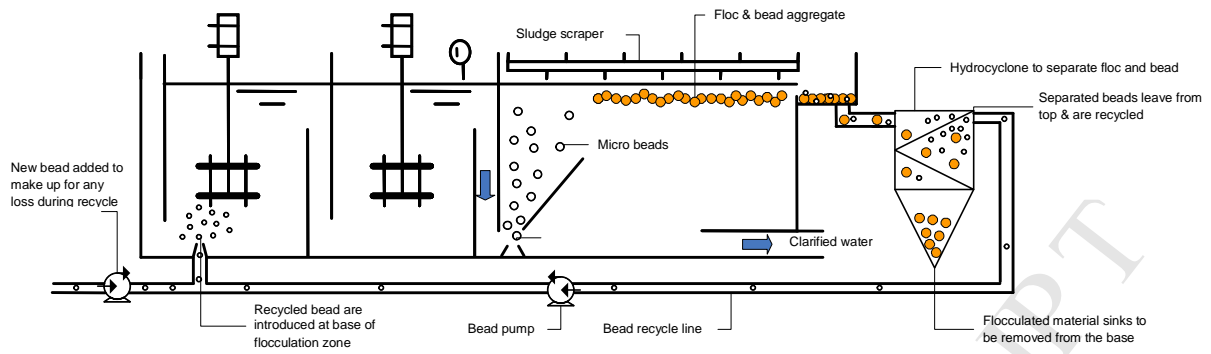
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