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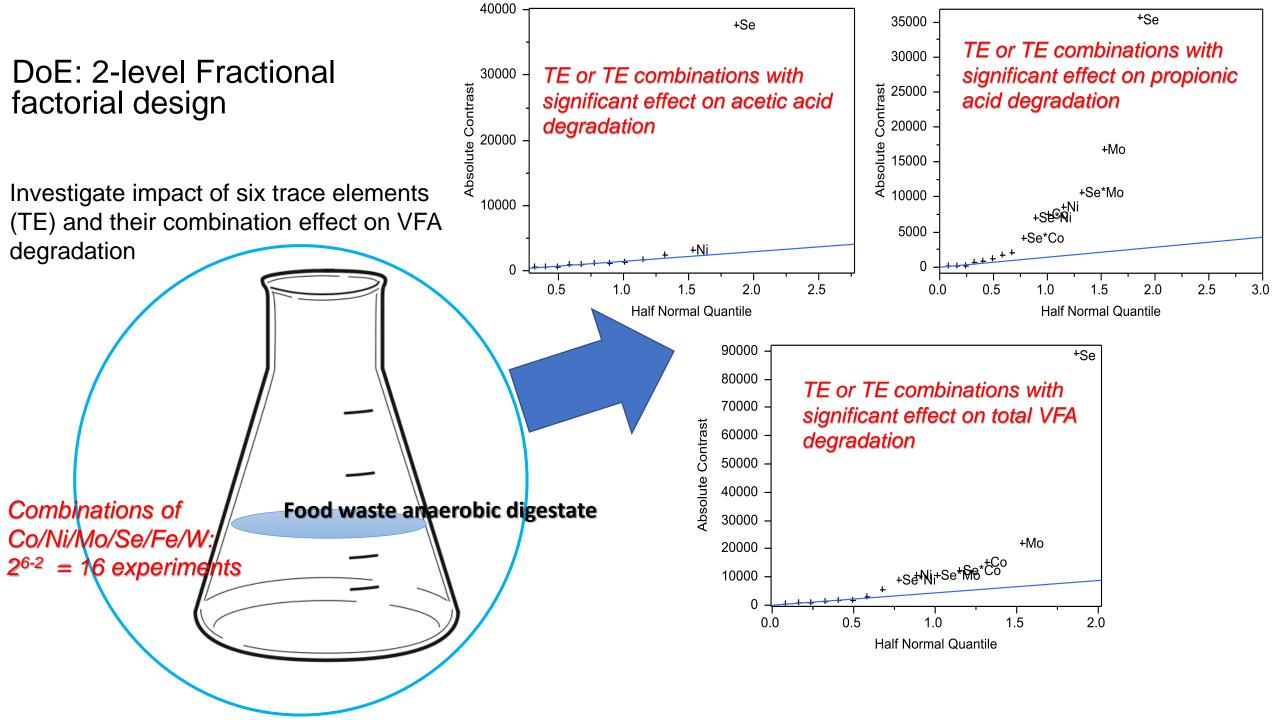
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- 1 Investigation of the Impact of Trace Elements on Anaerobic Volatile Fatty Acid
- 2 Degradation Using a Fractional Factorial Experimental Design
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10 ABSTRACT

- 11 The requirement of trace elements (TE) in anaerobic digestion process is widely
- 12 documented. However, little is understood regarding the specific requirement of
- 13 elements and their critical concentrations under different operating conditions such
- 14 as substrate characterisation and temperature.
- 15 In this study, a flask batch trial using fractional factorial design is conducted to
- 16 investigate volatile fatty acids (VFA) anaerobic degradation rate under the influence
- 17 of the individual and combined effect of six TEs (Co, Ni, Mo, Se, Fe and W). The
- 18 experiment inoculated with food waste digestate, spiked with sodium acetate and
- 19 sodium propionate both to 10 g/l. This is followed by the addition of a selection of
- 20 the six elements in accordance with a 2^{6-2} fractional factorial principle. The
- 21 experiment is conducted in duplicate and the degradation of VFA is regularly
- 22 monitored.

23	Factorial effect analysis on the experimental results reveals that within these	
24	experimental conditions, Se has a key role in promoting the degradation rates of both	
25	acetic and propionic acids; Mo and Co are found to have a modest effect on	
26	increasing propionic acid degradation rate. It is also revealed that Ni shows some	
27	inhibitory effects on VFA degradation, possibly due to its toxicity. Additionally,	
28	regression coefficients for the main and second order effects are calculated to	
29	establish regression models for VFA degradation.	
30	Key Words: Anaerobic digestion; Volatile fatty acids (VFA) degradation; Trace	
31	elements; Fractional factorial design	
32		
33	1. Introduction	
34	Anaerobic digestion (AD) of organic fraction of municipal solid waste (OFMSW) is	
35	an attractive technology for sustainable waste treatment and renewable energy	
36	production. In Europe, the installed AD capacity has increased significantly. The	
37	annual AD capacity grow from 4 million tons in 2006 (De Baere, 2006) to ~6 million	
38	tons by the end of 2010 (De Baere and Mattheeuws, 2010), reflecting the growing	
39	popularity of the technology.	
40	There is great potential for AD technology to generate both economic and	
41	environmental benefits. However, its application for certain organic waste streams has	
42	been considered intractable (Speece, 1983). This is largely due to lack of	
43	understanding of the complicated biological and biochemical reactions involved in the	
44	process, as well as the trace element (TE) requirements to maintain normal	
45	microbiological functions.	

46	Whilst TEs are required only at low concentration compared to macronutrients such
47	as C, N and P, they have a critical role in methanogenesis enzymatic activity (Demirel
48	and Scherer, 2011; Kida et al., 2001; Osuna et al., 2003) and microbial respiration
49	processes (Zandvoort et al., 2006). According to a number of studies and reviews
50	(Diekert et al., 1980; Fermoso et al., 2009; Gonzalez-Gil et al., 1999; Speece et al.,
51	1983; Zhang et al., 2003), cobalt, iron and nickel are required for both acetoclastic
52	and hydrogenotrophic methanogenesis pathways. Cobalt is the major constituent of
53	corrinoids which act as key cofactors in both methanogenesis pathways (Muller, 2003;
54	Murakami and Ragsdale, 2000; Thauer et al., 2008). Iron and nickel form the centre
55	parts of the Ni-Fe-S cluster and Fe-S cluster. These are important subunits of enzymes
56	such as hydrogenase and acetyl-CoA synthase involved in methanogenesis pathways
57	(Lindahl and Chang, 2001; Thauer et al., 2010). Nickel also forms the central site of
58	Methyl-Coenzyme M reductase – the key enzyme in the final step of all the
59	methanogenesis (Ermler et al., 1997).
60	Selenium, molybdenum and tungsten are also reported to be required in
61	
	methanogenesis (Fermoso et al., 2009). Based on an extensive literature survey of
62	methanogenesis (Fermoso et al., 2009). Based on an extensive literature survey of biochemical reactions involved in methanogenesis pathways, there is no suggestion of
62 63	
	biochemical reactions involved in methanogenesis pathways, there is no suggestion of
63	biochemical reactions involved in methanogenesis pathways, there is no suggestion of a requirement for selenium in the acetoclastic methanogenesis pathway.
63 64	biochemical reactions involved in methanogenesis pathways, there is no suggestion of a requirement for selenium in the acetoclastic methanogenesis pathway.(Deppenmeier, 2002; Ferry, 1999, 1992a,; Hille et al., 2014; Kryukov and Gladyshev,
63 64 65	 biochemical reactions involved in methanogenesis pathways, there is no suggestion of a requirement for selenium in the acetoclastic methanogenesis pathway. (Deppenmeier, 2002; Ferry, 1999, 1992a,; Hille et al., 2014; Kryukov and Gladyshev, 2004; Müller et al., 2010; Muller, 2003; Murakami and Ragsdale, 2000; Rother and
63 64 65 66	 biochemical reactions involved in methanogenesis pathways, there is no suggestion of a requirement for selenium in the acetoclastic methanogenesis pathway. (Deppenmeier, 2002; Ferry, 1999, 1992a,; Hille et al., 2014; Kryukov and Gladyshev, 2004; Müller et al., 2010; Muller, 2003; Murakami and Ragsdale, 2000; Rother and Krzycki, 2010; Stock and Rother, 2009; Thauer et al., 2008). Whilst selenium is
6364656667	 biochemical reactions involved in methanogenesis pathways, there is no suggestion of a requirement for selenium in the acetoclastic methanogenesis pathway. (Deppenmeier, 2002; Ferry, 1999, 1992a,; Hille et al., 2014; Kryukov and Gladyshev, 2004; Müller et al., 2010; Muller, 2003; Murakami and Ragsdale, 2000; Rother and Krzycki, 2010; Stock and Rother, 2009; Thauer et al., 2008). Whilst selenium is understood to be involved extensively in the hydrogenotrophic methanogenic pathway

dominant microbial population in digesters treating high ammonia feedstocks, such as
food waste (Angelidaki and Ahring, 1993; Banks et al., 2012;; Schnürer and Nordberg,
2008). This can result an increased dependence on hydrogenotrophic methanogenic
route in these digesters, hence the increased requirements of Se to form key enzymes.
Banks et al. (2012) observed a significant improvement of volatile fatty acid
degradation when Se was added to digesters treating food waste with an elevated
ammoniacal nitrogen concentration.
Molybdenum and tungsten are closely involved in formate dehydrogenase (in
hydrogenotrophic methanogens and syntrophic oxidising bacteria) and formyl-
methanofuran dehydrogenase (in hydrogenotrophic methanogens) (Hille et al., 2014;
Johnson et al., 1996; Romão, 2009). Although these two elements are considered to
be chemically analogous in enzyme formation (Kletzin and Adams, 1996), their
requirements for each methanogen species cannot be replaced by each other. The
stimulatory effect of W on the growth of methanogens Methanococcus vannielii was
first reported in the late 1970s (Jones and Stadtman, 1977). When formate was used as
the carbon source, the growth of <i>M. vannielii</i> was significantly enhanced by the
addition of W, but not by Mo. However, the growth stimulation by W was not
observed when the organism was fed with H_2 and CO_2 , which suggested the
involvement of a W-containing formate dehydrogenase (FDH). In a lab scale
digestion trial, an improvement of performance was reported by Jiang et al. (2012)
following the addition of W to anaerobic digesters treating vegetable waste.
Combinations of key TEs were reported to have synergistic or antagonistic effects to
the methanogenesis. Patidar and Tare (2006) report that addition of Fe/Co, Fe/Ni/Zn
and Ni/Zn/Co combinations lead to maximum total methanogenic activity in a batch

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95	study; however addition of Ni affects Co and Zn uptake in methanogens, due to
96	antagonistic effect of metal ions. Feng et al. (2010) investigated the effects of
97	additions of Co, Ni/Mo/B and Se/W on the biogas production and the associated
98	anaerobic microbial community. The study revealed that the highest methane
99	production occurs at high Se/W concentrations in combination with a low level of Co.
100	However, the influence of the trace metal additions on the microbial community
101	composition was not significant.
102	It is clear that TE addition can benefit the operation of many lab and commercial scale
103	anaerobic digesters with TE deficiency. However, baseline TE concentration
104	feedstocks as well as microbial community structure can significantly influence the
105	TE requirement in anaerobic digesters. To date, it remains a challenge to determine
106	whether a digester is deficient of TE particularly to identify exact TE species required.
107	The currently practice of TE supplementation in many commercial AD plants tends to
108	be a full addition of all trace elements, many of which are not required and even
109	potentially harmful to their AD processes. It is therefore important for the operators to
110	understand the specific requirement of each digester, in order to supplement only
111	those elements required for improved performance and reduction of operational cost.
112	Significant work has been carried out to investigate the effects of TE and their
113	combinations on VFA degradation (Feng et al., 2010; Worm et al., 2009). Feng et al.
114	(2010) demonstrated advanced statistical methods can be a useful tool to interpret the
115	impact of TE on anaerobic digestion process efficiency. However, to the authors' best

- 116 knowledge, to date no study has applied factorial design method to investigate the
- 117 effects of multiple elements simultaneously. Factorial design is widely applied in
- scientific investigations to identify the key factors and/or combinations of factors

influencing the process. In this study, the effects of six TEs (Co, Ni, Se, Mo, Fe and
W) on the volatile fatty acids degradation rate were studied. In order to differentiate
the effects of each element and/or element combination, a design of experiment (DoE)
approach of 2-level fractional factorial design was adopted. Regression model has
also been developed to correlate the response with each factor.

124 2. MATERIALS AND METHODS

125 **2.1 Food waste digestate inoculum**

126 Inoculum used in this study was acquired from a 5-litre lab scale CSTR mesophilic

127 (37 °C) anaerobic digester fed with food waste. The digester had been in operation for

128 over a year and reached its stable state, i.e. stable VFA concentration, ammonia

129 nitrogen level and daily biogas production in the past 2 months. During the entire

130 period of its operation, no TE supplementation regime was adopted. Before starting

131 this study, acetate and propionate concentrations in digestate inoculum were both

132 increased to 10 g/l by addition of sodium acetate and sodium propionate (Reagent

133 grade, Fisher Chemical, UK). D-glucose and starch were added as substrates, both at

134 concentrations of 4 g/l in the digestate. The background ammonia nitrogen

135 concentration in the inoculum was high at 6.5 g/l.

136 2.2 Analytical methods

137 The baseline TEs concentrations in digestate inoculum were determined using ICP-

138 MS or ICP-OES at a UKAS accredited commercial laboratory (Severn Trent Services,

139 Coventry, UK) after in-house hydrochloric – nitric acid digestion in accordance with

140 US EPA standard method 3010A. Baseline TEs concentrations in the inoculum have

141 been determined and the results are shown in Table 1.

142	Volatile fatty acid (VFA) concentrations were quantified in a Shimazdu 2010 gas		
143	chromatograph (Shimadzu, Milton Keynes, UK) using a method described in Jiang et		
144	al. (2012). Samples were acidified using formic acid. Three standard solutions		
145	containing 50, 250 and 500 mg l ⁻¹ of acetic, propionic, iso-butyric, n-butyric, iso-		
146	valeric, valeric, hexanoic and heptanoic acids were used for VFA calibration.		
147	2.3 Trace elements supplementation and VFAs degradation flask trial		
148	The batch experiment was set up in conical flasks each with 250 ml capacity. Each		
149	flask was inoculated with 200 ml digestate. Working solutions of six trace elements		
150	(Co, Ni, Mo, Se, Fe and W) were prepared using CoCl ₂ , NiCl ₂ , (NH ₄) ₆ Mo ₇ O ₂₄ ,		
151	Na ₂ SeO ₃ , FeCl ₂ and Na ₂ WO ₄ (Analytical grade, Fisher Chemical, UK). Addition of		
152	these trace element solutions into each flask followed the pattern as shown in Table 2,		
153	with strings of plus (+) and minus (-) signs denoting with and without addition of a		
154	particular element. There are in total 16 TE addition combinations including a control		
155	where no TE was added. When TE solutions are added, the concentrations of		
156	elements increases by 1 mg/l, 1 mg/l, 0.2 mg/l, 0.2 mg/l, 5 mg/l and 0.2 mg/l for Co,		
157	Ni, Mo, Se, Fe and W, respectively. To ensure data quality, this experiment was		
158	conducted in duplicate, i.e. 32 total flasks.		
159	Flask headspaces were flushed with carbon dioxide and nitrogen (20:80) mixture gas		

Flask headspaces were flushed with carbon dioxide and nitrogen (20:80) mixture gas (BOC, UK). Each flask was then sealed with rubber bungs with an outlet connected to a 1-litre Tedlar bag to maintain an ambient pressure in the headspace. All flasks were randomised and placed in an orbital incubator (Weiss-Gallenkamp, UK) set at 36°C and 60RPM.

164 Digestate samples in flasks were routinely sampled and analysed for VFA

165 concentrations in order to plot VFA degradation curves over the duration of the

- 166 experiment. 1 ml of the digestate was withdrawn from each flask using hand pipette
- 167 with cut-off pipette tip. After sampling, the headspace of each flask was flushed with
- 168 CO_2/N_2 (20:80) and then sealed again before returning to the incubator. The flask
- 169 experiment has continued for ~90 days until all VFAs in the flasks were depleted.
- 170 Similar to the approach adopted by Olaisen et al. (2002), the numerical integrals of
- 171 each VFAs degradation curves, assigned as the Degradation Index (DI), are calculated
- 172 according to the trapezoidal rule (Equation 1).

173 $DI = \int_0^\infty f(t) dt \approx \sum_{i=0}^n (t_{i+1} - t_i) \times (C_{i+1} + C_i)/2$ (Equation 1)

- 174 Where: *t* is the time in days; *C* is the measured value of VFA concentration.
- 175 DI enables the quantification of VFA degradation efficiency based on the degradation
- 176 curve; a smaller *DI* number indicates a more efficient degradation. The calculated *DI*
- 177 numbers were used as experimental responses in the following statistical and
- 178 modelling analysis.

179 **2.4 Statistical method**

180 2.4.1Fractional factorial design and data analysis

181 The experiment was designed to expose the key TEs which are the most effective on

182 improving anaerobic degradation of VFAs, whilst understanding possible synergistic

183 effects from combinations of elements. Six trace elements were selected as factors to

184 be investigated in this 2-level factorial experiment design. Due to practical difficulties,

- 185 it is not feasible to run a 2-level full factorial experiment, as this requires $64 (=2^6)$
- 186 flasks to be prepared for a single experiment. As a result, a 2^{6-2} fractional factorial

187 design, which is ¹/₄ of the full factorial design with a resolution of IV was adopted.

- 188 The chosen design reveals all the main effects. However it cannot distinguish fully
- 189 the 2-factor interaction due to effect aliasing (Wu and Hamada, 2009). The

- 190 experimental arrangement of this 2^{6-2} subset (fraction) of a full factorial design was
- 191 generated using the SAS JMP statistical suite (SAS Institute Inc. Cary, North Carolina,
- 192 USA).
- 193 The average independent effect of each TE (factor), known as the main effect (ME) is
- 194 calculated as follows:
- 195 For any factor (A), the main effect $ME(A) = \overline{z}(A+) \overline{z}(A-)$ (Equation 2)
- 196 Where $\bar{z}(A+)$ is the average value of all observations (VFA concentrations) when
- 197 factor (*A*) is present (+ level).
- 198 Similarly, \bar{z} (A-) is the average value of all observation when factor (A) is not present
- 199 (level) of factor (*A*), i.e. without addition of the particular TE. Higher order factorial
- 200 effects can be calculated in the in a similar way.
- 201 2.4.2 Regression Model
- 202 Factorial effects in a 2^k factorial design can be estimated using a multiple regression
- 203 model (Wu and Hamada, 2009). For the current experiment design which contains16
- 204 observations, the model can be expressed as:

205
$$z_i = \beta_0 + \sum_{i=1}^{16} \beta_i x_i + \sum_{j < i} \beta_{ij} x_i x_j$$
 (Equation 3)

- 206 Where: β_0 , β_i and β_{ij} are the regression coefficients which are estimated using Least 207 squares method. z_i is the experiment response (*DI*). x_i and x_j are the variables (i.e. the
- 208 factors in the experiment, coded with +/-).

209 3. RESULTS AND DISCUSSION

210 **3.4 Trace elements flask trial**

211 A reasonable repeatability of the experiment was achieved as indicated by the error

bar shown the range of the result from the duplicated experiment (Figure 1). Clear

- 213 divergence in degradation rate was observed amongst the sample groups where
- 214 different TE supplementation regimes were adopted. Under each TE addition

215	combinations, the VFAs degradation indexes were calculated based on VFA
216	concentration results using Equation 1. The <i>DI</i> results are summarised in Table 3.
217	Notably, in samples where selenium was present, the acetate and propionate
218	degradations were significantly stimulated in comparison with selenium deprived
219	groups.
220	On day 28, acetic acid concentration in all samples was observed to be degraded
221	below 100 mg/l. However, this is followed by an increase of acetic acid
222	concentrations in some samples as indicated by the bumps on the curves in Figure 1a
223	& 1b around day 30. This increase of acetic acid is caused by acetic acid produced
224	from propionate degradation as the timing coincided with an increased propionate
225	degradation rate in those samples during this period as shown in Figure 1c & 1d. The
226	fractional factorial design significantly reduced the experimental efforts normally
227	required to understand complex TE requirement in AD process. Due to the simplicity
228	of the experimental and analytical requirement, this method can therefore be adopted
229	by AD operators to assess the deficiency of TE in digesters and provide critical
230	information on specific TE requirement. This enables a more targeted element
231	supplementation regime to avoid a full TE addition, therefore significantly reduce
232	operational costs.

233 **3.5 Analysis of factorial effects**

The factorial effects of each individual element (Se, Mo, Co, Ni, W, Fe) and 2-factor combinations in contributing to VFA degradation rate is calculated according to Equation 2 and summarised in Table 4. The '-' signage indicates the effect contribute negatively to the *DI*. In other words, when a factorial effect is negative, that factor results a decrease in *DI* (an improved VFA degradation).

239 A high resolution factorial design (Resolution IV) was adopted in this study. This 240 design ensures main effect of each individual factor is clear from confounding. 241 However, higher order effects are affected by confounding and less likely to be 242 important according to effect hierarchy principle (Wu and Hamada, 2009). For these 243 reasons, this study only considers effects of individual elements and 2-factor 244 combinations. 245 The factorial effects are illustrated in main effect plots and half- normal probability plots in Figure 2. In main effect plots, each factor is represented by a line connecting 246 247 the average values of all observations at the high (+) level and low (-) level; the vertical height of each line indicates the main effect. In half- normal probability plots, 248 factors of more significance are found at the upper right corner away from the 'near-249 250 zero' line and the significance is quantified as Lenth's t-ratio shown in Table 4. 251 The factorial effect of Se suggests it significantly improves the degradation for both acetic and propionic acids. Mo and Co show a moderate effect on improving 252 253 propionic degradation rate; whereas Ni shows some negative effects on both acetic 254 and propionic acids degradation as also indicated by its factorial effects (Table 4). The inhibitory effect of Ni during anaerobic digestion processes were widely reported 255 256 in previous literatures (Ahring and Westermann, 1983; Lin, 1993, 1992). Lin (1992) 257 observed a moderate inhibitory effect from Ni in a batch experiment under mesophilic 258 condition. The concentrations of Ni to inhibit 50% of the acetic and propionic acid 259 degradation are reported to be 89 and 226 mg/l, respectively. In current study, the 260 concentration of Ni at 1.3 mg/l is considerably lower than the reported values which 261 are considered to have significant inhibitory effects. However, it is possible that the

Ni inhibition starts at low concentration and only can be revealed with statisticalanalysis.

264 Combination effects of two elements show no significant effect on acetic acid 265 degradation as indicated by p-value (Table 4). For propionic acid, however, Se/Mo combination shows significant effect on improving the degradation of the acid, 266 whereas Mo/W combination exhibits an inhibitory effect. However, the effect of 267 268 Se/Mo is confounded with Co/W and Ni/Fe due to alias effect caused by the intrinsic 269 inadequacy of a fractional design. Therefore, the improved propionic acid degradation 270 can equally be caused by Co/W and Ni/Fe combinations. Further experiment is 271 needed to clarify this ambiguity. For the same reason, the inhibitory effect of Mo/W is confounded with Se/Co. However, in a previous semi-continuous food waste 272 273 digestion experiment which provided the inoculum for the study, it was confirmed 274 that Co was beneficial to the degradation of VFA including propionic acid at the 275 presence of selenium (Banks et al., 2012). This clearly suggests the confounding 276 factorial effect of Se/Co can be dismissed in the current study, thus leaving the W/Mo 277 effect as evident. Indeed, the antagonistic effects between W and Mo for some 278 methanogens were reported in a number of studies (Kletzin and Adams, 1996; May et 279 al., 1988; Zellner and Winter, 1987). The studies suggest that W replaces Mo as the 280 centre of formate dehydrogenase (FDH), a key enzyme involved in propionic acid 281 degradation, and consequently deactivates the enzyme (May et al., 1988). 282 The factorial analysis reveals a strong influence of Se on the VFA degradation, 283 suggesting a specific requirement for Se to improve VFA, particularly acetate and 284 propionate degradation in this type of digestate. Previous study (Banks et al., 2012) 285 confirmed the foodwaste digestate used as inoculum in this study adopt

295	3.6 Regression models for VFA degradation indexes (DI)
294	the factorial analysis.
293	improved propionate degradation overserved in the presence of these two elements in
292	dehydrogenases and hydrogenases required for propionate degradation, hence the
291	reported by Worm et al. (2011, 2009), Se and Mo are involved in formate
290	syntrophic methanogenic pathway, but not in the acetoclastic pathway. In addition, as
289	1990; Jones et al., 1979; Vorholt et al., 1997) that Se involves significantly in the
288	this study, as it is understood based on wide range of literature evidence (Ferry, 1999,
287	elevated ammoniacal nitrogen level. This explains the significant effect of Se found in
286	hydrogenotrophic methanogenesis as the principle methanogenic route due to the

296 Regression coefficients for factorial effects of the individual elements and 2-factor 297 combinations were calculated using the least squares method and summarised in 298 Table 5. Using these coefficients, regression models correlates factorial effects with degradation index for acetic acid (DI_{HAc}) , propionic acid (DI_{HPr}) and total VFA 299 300 (DI_{Total}) can be established according to Equation 3 as follows: $DI_{HAc} = 125728.25 - 33687.5 X_{Se} - 1642.9 X_{Mo} - 9.4 X_{Co} + 5027.2 X_{Ni} - 2940.6 X_W$ 301 302 $+3383X_{Fe}+1008.8X_{Se}X_{Mo}$ -1685.4X $_WX_{Mo}$ -3156.7X $_{Mo}X_{Co}$ +2849.6X $_{Se}X_{Ni}$ +3379.1X MoXNi -1991.9X CoXNi -3692.1 X NiXW 303 (Equation 4) 304 $DI_{HPr} = 203360.9 - 40051.1X_{Se} - 18208.0X_{Mo} - 11276.6X_{Co} + 9196.4X_{Ni} - 1808.9X_W - 180$ 305 $1580.0X_{Fe} - 9316.2X_{Se}X_{Mo} + 7821.2 X_{W}X_{Mo} - 863.7X_{Mo}X_{Co} + 6341.9X_{Se}X_{Ni} - 6341.9X_{Se}X_$

306
$$1000.2X_{Mo}X_{Ni} + 550.1X_{Co}X_{Ni} - 1504.0X_{Ni}X_W$$
 (Equation 5)

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307	$DI_{Total} = 392720.7 - 89005.3X_{Se} - 21657.1X_{Mo} - 15040.2X_{Co} + 10573.0X_{Ni} - 5443.6X_{W} - 6443.6X_{W} - 6443.6X_{W$
308	2883.7 X_{Fe} -10644.4 $X_{Se}X_{Mo}$ +12260.4 $X_{W}X_{Mo}$ -893.9 $X_{Mo}X_{Co}$ +8807.8 $X_{Se}X_{Ni}$
309	$+1258.5X_{Mo}X_{Ni} + 927.7X_{Co}X_{Ni} - 1540.0X_{Ni}X_{W} $ (Equation 6)
310	Insignificant factorial effects at less than 5% significance level, i.e. adjusted p-
311	value >0.05 (Table 4) can be ignored to simplify the above models, where the
312	equations then become:
313	$DI_{HAc} = 125728.25-33687.5X_{Se}$ (Equation 7)
314	$DI_{HPr} = 203360.9 - 40051.1X_{Se} - 18208.0X_{Mo} - 11276.6X_{Co} + 9196.4X_{Ni} - 9316.2X_{Se}X_{Mo}$
315	+7821.2 $X_{W}X_{Mo}$ (Equation 8)
316	$DI_{Total} = 392720.7 - 89005.3X_{Se} - 21657.1X_{Mo}$ (Equation 9)
317	Analysis of variance (ANOVA) is then used to evaluate the adequacy of the empirical
318	models. The results show coefficients of determination (\mathbf{R}^2) for acetic, propionic and
319	total VFA model are 0.9976, 0.9998 and 0.9998, respectively. There are significantly
320	high F-ratios for acetic, propionic and total VFA models (65.27, 949.25 and 614.63,
321	respectively) indicating a large values of model sum of squares (SS). All three models
322	have sufficiently low (<5%) p-values indicating small probability of error for SS.
323	Both F-ratios and p-values strongly support the robustness of the regression models
324	developed under the experimental conditions in this study.
325	4. Conclusion

326 This study applied a fractional factorial design in batch experiments to explore the

impact of six trace elements (Co, Ni, Mo, Se, Fe and W) on VFA degradation rates in 327

foodwaste digestate. The factorial analysis of the results shows a significant influence 328

- 329 from Se in improving the VFA degradation. This is in agreement with previous
- 330 knowledge of the involvement of Se in hydrogenotrophic methanogenesis which is
- 331 typically the main methane forming route in foodwaste digesters.
- 332 The factorial results also reveal Ni shows slight inhibitory effect to VFA degradation
- 333 and W/Mo combination inhibits propionic acid degradation, likely due the
- antagonistic effect of the two elements.
- 335 The fractional factorial method has introduced an efficient experimental approach to
- identifying the deficient elements in anaerobic digesters, whilst offering guidance on a
- tailored TE supplementation recipe for digesters operating under different conditions.
- 338

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Element	Baseline concentration (mg/	
Fe	103	-/
Со	0.035	
Mn	14.2	
A1	42.8	
Zn	7.94	
Mo	0.13	
Cu	2.23	
Ni	0.304	
Se	0.061	
W	<0.1	

Table 1. Baseline	e TE concen	trations in	the inoculum
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Flasks	Co	Ni	Мо	Se	Fe	W
1, 2	-	-	-	-	-	-
3,4	-	-	-	+	+	+
5,6	-	-	+	-	+	+
7,8	-	-	+	+	-	-
9,10	-	+	-	-	+	-
11, 12	-	+	-	+	-	+
13, 14	-	+	+	-	-	+
15, 16	-	+	+	+	+	-
17, 18	+	-	-	-	-	+
19, 20	+	-	-	+	+	-
21, 22	+	-	+	-	+	-
23, 24	+	-	+	+	-	+
25, 26	+	+	-	-	+	+
27, 28	+	+	-	+	-	-
29, 30	+	+	+	-	-	-
31, 32	+	+	+	+	+	+

Table 2. Experimental design matrix

s TE Combinations			
Comoniations		DI(Ave)	
	Acetic	Propionic	Total
Control	153,014	267,184	527,562
Se/Fe/W	84,583	174,707	315,682
Mo/Fe/W	155,723	249,938	488,264
Mo/Se	81,555	127,735	260,954
) Ni/Fe	160,765	277,510	521,616
2 Ni/Se/W	85,390	208,527	349,237
4 Ni/Mo/W	161,457	255,408	498,665
6 Ni /Mo/Se/Fe	123,414	156,092	300,108
8 Co/W	163,569	232,874	462,669
0 Co/Se/Fe	88,583	170,724	314,340
2 Co/Mo/Fe	156,647	212,234	441,349
4 Co/Mo/Se/W	81,935	117,920	246,362
6 Co/Ni/Fe/Se	157,778	231,647	459,108
8 Co/Ni/Se	99,000	209,378	364,809
0 Co/Ni/Mo	166,373	220,501	454,576
2 Co/Ni/Mo/Se/Fe/W	91,865	141,395	278,231
	R		

 Table 4. Summary of factorial effects

		Acetic acid					Propionic acid				Total VFA		
Term	Aliases	Factorial effect	Lenth t-Ratio	Individual p-Value	Adjusted p-Value	Factorial effect	Lenth t- Ratio	Individual p-Value	Adjusted p-Value	Factorial effect	Lenth t-Ratio	Individual p-Value	Adjusted p-Value
Se	Mo/Co/W, Mo/Ni/Fe	-67375	-12.19	<0.0001*	<0.0008*	-80102	-26.69	<0.0001*	<0.0001*	-178011	-20.58	0.0001*	0.0002*
Mo	Se/Co/W, Se/Ni/Fe	-3286	0.59	0.59	1.00	-36416	-12.14	0.0002*	0.0008*	-43314	-5.01	0.0035*	0.02*
Co	Se/Mo/W, Ni/W/Fe	-19	-0.00	1.00	1.00	-22553	-7.52	0.0005*	0.0052*	-30080	-3.48	0.01*	0.09
Ni	Se/Mo/Fe, Co/W/Fe	10054	1.82	0.09	0.59	18393	6.13	0.0019*	0.01*	21146	2.44	0.03*	0.26
W	Se/Mo/Co, Co/Ni/Fe	-5881	-1.06	0.28	0.99	-3618	-1.21	0.22	0.95	-10887	-1.26	0.20	0.93
Fe	Se/Mo/Ni, Co/Ni/W	3383	0.61	0.58	1.00	-3160	-1.05	0.27	0.99	-5767	-0.67	0.50	1.00
Se/Mo	Co/W, Ni/Fe	2018	0.37	0.74	1.00	-18632	-6.21	0.0019*	0.01*	-21289	-2.46	0.03*	0.26
Mo/W	Se/Co	-3371	-0.61	0.58	1.00	15642	5.21	0.0028*	0.02*	24521	2.83	0.02*	0.20
Mo/Co	Se/W	-6313	-1.14	0.25	0.97	-1727	-0.58	0.60	1.00	-1788	-0.21	0.85	1.00
Se/Ni	Mo/Fe	5699	1.03	0.29	0.99	12684	4.23	0.0064*	0.05	17616	2.04	0.06	0.45
Mo/Ni	Se/Fe	6758	1.22	0.21	0.95	-2000	-0.67	0.50	1.00	2517	0.29	0.79	1.00
Co/Ni	W/Fe	-3984	-0.72	0.45	1.00	1100	-0.37	0.74	1.00	1855	0.21	0.84	1.00
Ni/W	Co/Fe	-7384	-1.34	0.18	0.90	-3008	-1.00	0.30	0.99	-3080	-0.36	0.74	1.00

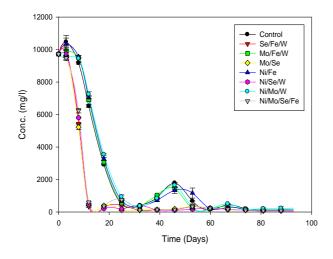
Note: p-Values marked with '*' denote the factor being significant (<0.05)

being significant (volue)

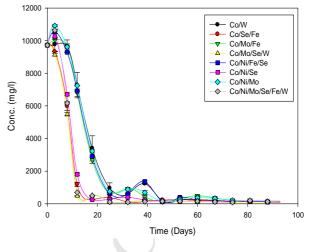
	Acetic Acid		Propionic Acid						
Term	Coefficient	t Ratio	Prob> t	Coefficient	t Ratio	Prob> t	Coefficient	t Ratio	Prob> t
Constant	125728.25	104.63	< 0.0001	203360.9	466.45	< 0.0001	392720.7	367.71	< 0.0001
Se	-33687.5	28.03	0.0013	-40051.1	91.86	0.0001	-89005.3	83.34	0.0001
Mo	-1642.9	-1.37	0.3049	-18208.0	41.76	0.0006	-21657.1	20.28	0.0024
Co	-9.4	0.01	0.9945	-11276.6	25.87	0.0015	-15040.2	14.08	0.0050
Ni	5027.2	-4.18	0.0527	9196.4	-21.09	0.0022	10573.0	-9.90	0.0100
W	-2940.6	2.45	0.1342	-1808.9	4.15	0.0535	-5443.6	5.10	0.0364
Fe	1691.5	-1.41	0.2945	-1580.0	3.62	0.0684	-2883.7	2.70	0.1142
Se/Mo	1008.8	0.84	0.4895	-9316.2	-21.37	0.0022	-10644.4	-9.97	0.0099
Se/Co	-1685.4	-1.40	0.2958	7821.2	17.94	0.0031	12260.4	11.48	0.0075
Mo/Co	-3156.7	-2.63	0.1195	-863.7	-1.98	0.1861	-893.9	-0.84	0.4907
Se/Ni	2849.6	2.37	0.1411	6341.9	14.55	0.0047	8807.8	8.25	0.0144
Mo/Ni	3379.1	2.81	0.1066	-1000.2	-2.29	0.1487	1258.5	1.18	0.3599
Co/Ni	-1991.9	-1.66	0.2392	550.1	-1.26	0.3343	927.7	0.87	0.4766
Ni/W	-3692.1	-3.07	0.0916	-1504.0	-3.45	0.0747	-1540.0	-1.44	0.2861

Table 5. Coefficients of the empirical regression models for acetic, propionic and total VFA degradation index (DI)

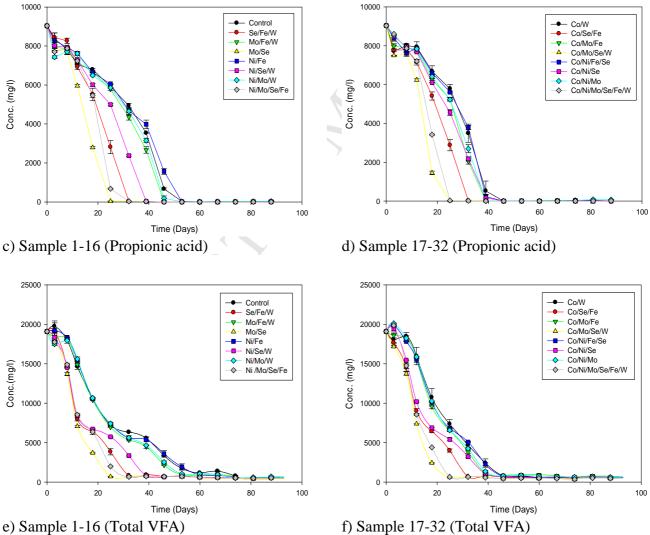
Note: Prob>|t|: Probability value of getting an even greater t-statistic (in absolute value), less than 0.05 are considered as significant evidence that the parameter is not zero

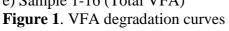


a) Sample 1-16 (Acetic acid)

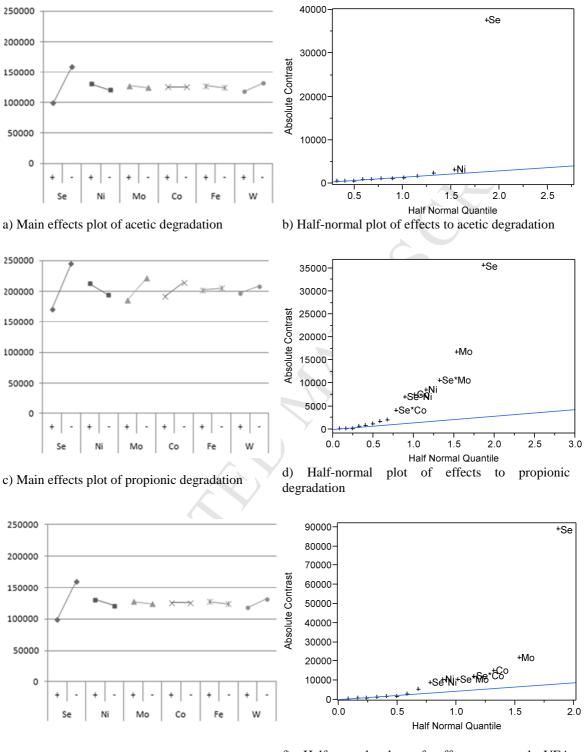


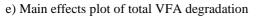
b) Sample 17-32 (Acetic acid)





Note: Error bar indicate the range of results from duplicated experiment





f) Half-normal plot of effects to total VFA degradation

Figure 2. Main effects plot and Half normal probablilty plot of standardised effects for VFA degradation

Highlights

- A fractional factorial design of experiment to study effects of TE on VFA degradation
- Method enables to determine specific TEs requirement in a given AD plant
- Se has the most significant effect on improving VFA degradation in foodwaste digestate
- Using numerical integral to quantify VFA degradation performance over a given time