

1 **A datamining approach to identifying spatial patterns of phosphorus forms**
2 **in the Stormwater Treatment Areas in the Everglades, US.**

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16

17 **Abstract**

18

19 The Everglades ecosystem in Florida, USA, is naturally phosphorus (P) limited,
20 and faces threats of ecosystem change and associated losses to habitat,
21 biodiversity, and ecosystem function if subjected to high inflows of P and other
22 nutrients. In addition to changes in historic hydropattern, upstream agriculture
23 (sugar cane, vegetable, citrus) and urbanization has placed the Everglades at risk
24 due to nutrient-rich runoff. In response to this threat, the Stormwater Treatment
25 Areas (STAs) were constructed along the northern boundary of the Everglades as

26 engineered ecological systems designed to retain P from water flowing into the
27 Everglades. This research investigated data collected over a period from 2002 to
28 2014 from the interior of the STAs using data mining and analysis techniques
29 including a) exploratory methods such as Principal Component Analysis to test
30 for patterns and groupings in the data, and b) modelling approaches to test for
31 predictive relationships between environmental variables. The purpose of this
32 research was to reveal and compare spatial trends and relationships between
33 environmental variables across the various treatment cells, flow-ways, and STAs.
34 Common spatial patterns and their drivers indicated that the flow-ways do not
35 function along simple linear gradients; instead forming zonal patterns of P
36 distribution that may increasingly align with the predominant flow path over
37 time. Findings also indicate that the primary drivers of the spatial distribution of
38 P in many of these systems relate to soil characteristics. The results suggest that
39 coupled cycles may be a key component of these systems; i.e. the movement and
40 transformation of P is coupled to that of nitrogen (N).

41

42 **Keywords:** phosphorus, data mining, stormwater treatment areas, constructed
43 wetland, Everglades, water quality

44

45 **1. Introduction**

46

47 The Stormwater Treatment Areas (STAs), located around the northern boundary
48 of the Everglades in Florida, USA, were constructed over a period from 1994 to
49 2013. As a set of engineered ecological systems, the general purpose and
50 function of the STAs is to reduce phosphorus (P) in runoff water prior to

51 discharging to the Everglades Protection Area. They consist of a series of
52 shallow, freshwater marshes divided into flow-ways and treatment cells by
53 interior levees and control structures, populated with emergent or submergent
54 aquatic vegetation (EAV and SAV, respectively) (Chen et al., 2015). The
55 Everglades as a system is naturally P limited (Entry, 2014; McCormick et al.,
56 1996), and so the water it receives must meet stringent requirements for ultra-
57 low levels of water P (Pietro and Ivanoff, 2015). Since 1995, the STAs have
58 treated approximately 16.5 billion m³ inflow volume, retained approximately
59 1,727 metric tons (mt) of total phosphorus (TP), lowering phosphorus surface
60 water concentrations from an overall annual TP of 140 micrograms per liter (μg
61 L^{-1}) to 37 $\mu\text{g L}^{-1}$ (flow weighted mean; South Florida Water Management District,
62 2015), and improving further in most recent years to exhibit outflow
63 concentrations averaging between 15-25 $\mu\text{g L}^{-1}$ (South Florida Water
64 Management District et al., 2015). STA-2 and STA-3/4 are two of the best
65 performing STAs, and have recorded reductions in surface water P from 100 and
66 87 $\mu\text{g L}^{-1}$ at inflow structures, respectively, to 23 and 18 $\mu\text{g L}^{-1}$ at outflow (Pietro
67 and Ivanoff, 2015).

68

69 The STAs are wetland systems, and the controls on the P removal process are
70 therefore set by the internal biogeochemical, ecological and physical processes
71 and conditions in each cell, in each STA (Ivanoff et al., 2013). Phosphorus
72 reduction from each STA must be maximized in order to meet stringent
73 regulatory effluent limits, which implies that these natural processes must be
74 manipulated (engineered) to maximize P retention. Phosphorus in surface water
75 can have various forms; from soluble reactive to forms of organic and particulate

76 P with varied degrees of recalcitrance (Reddy and DeLaune, 2008). The retention
77 of P in these systems needs to therefore consider these different forms.

78

79 There are abiotic processes of P retention, including P sorption to the STA soil
80 particulates (Reddy et al., 1999) and particulate (co)-precipitation with cations
81 such as calcium (Ca), magnesium (Mg), iron (Fe), and aluminium (Al) (Malecki-
82 Brown et al., 2007). Factors that influence these processes are surface flow rate
83 and path (Kadlec and Wallace, 2009) but also water and soil chemistry (e.g.
84 concentrations of Ca, Mg, Fe and Al), pH, and the oxidation reduction potential
85 (Reddy et al., 1999). Ideally this P then gets buried, or retained by the sediment
86 within the wetland, resulting in gradually lower soil P-levels as water flows from
87 the inflow point towards the outflow points (P gradient), similar to what has
88 been observed in the nearby Water Conservation Area 2A (DeBusk et al., 1994).
89 There are circumstances under which P is transported along the hydrologic
90 gradient due to sediment re-suspension, P desorption from the sediment matrix,
91 or poor vegetation condition. In properly performing STAs, these are limited and
92 water column P could be reduced further down the flow-way, reducing the slope
93 of the gradient. Uptake and retention of P by plants is generally (though not
94 exhaustively; dependent upon plant type) considered to be short-term and rapid;
95 while abiotic/physical retention processes tend to be longer term and are
96 considered to account for 50-70% of permanent storage (Richardson, 1999).

97

98 Biological cycling of P involves direct uptake of available P by plant and
99 microbial communities (Newman et al., 2001) to meet their physiological
100 requirements, action of extracellular enzymes on complex organic P to release P

101 uptake (Corstanje et al., 2007) and the release of P from the biological
102 decomposition of organic material. Under anaerobic environments,
103 decomposition of organic material is slow, resulting in formation and accretion
104 of peat; forming another sink for P as long as the peat remains intact. Biological P
105 cycling and the resulting spatial distribution of the different forms of P is highly
106 complex, as it is driven by coupled P, N and C cycles; determined by redox
107 conditions and characterized by the plant ecology (Chen et al., 2015; Orem et al.,
108 2014; Reddy et al., 2011).

109

110 Extensive sampling has been conducted over a period from 2002 to 2014, in
111 which soil, surface water and macrophytes have been sampled within the STA
112 cells, resulting in a large dataset of observations. Coupled with hyper-spectral
113 measurements made through various aerial surveys, the results comprise a fairly
114 comprehensive dataset on the spatial variation in key components of the STA
115 ecosystem. Here, we report on a broad scale analysis of these datasets, in order
116 to determine common trends across the various flow-ways in the STAs, and in
117 individual STAs. The expectation here is that common biogeochemical processes
118 will generate common multivariate patterns across STAs. We then considered,
119 given the extent and comprehensiveness of the datasets under consideration,
120 implications for future monitoring of these systems.

121

122 **2. Materials and Methods**

123 **2.1. Study Area**

124

125 The STAs, operated by the South Florida Water Management District, cover an
126 effective treatment area of circa 230 km². There are five STAs: STA-1E, STA-1W,
127 STA-2, STA-3/4, and STA-5/6 (Figure 1); STA-5/6 was formerly two separate
128 STAs until water year (WY) 2010. The STAs vary in size and location, and each is
129 constructed with sets of interconnected cells forming treatment 'flow-ways'.
130 Data from surface water (sampled along internal transects within the treatment
131 cells), floc (i.e. flocculant; loosely clumped particles either suspended in the
132 water column or resting atop the soil, analogous to litter in terrestrial systems),
133 and soil collected within the various cells were available for analysis, and have
134 been previously described and used to evaluate conditions within the STAs (e.g.
135 Pietro and Ivanoff, 2015; Reddy et al., 2009). Normalized Difference Vegetation
136 Index (NDVI) and vegetation class and habitat maps were derived from recent-
137 year hyper-spectral imagery at a resolution of approximately 1 square foot to
138 represent the approximate current state of vegetation within the cells. The
139 available datasets were diverse in spatial extents, subjects (e.g. soil samples,
140 surface water transects, vegetation coverage) and data types (e.g. categorical vs.
141 continuous), necessitating a data mining approach capable of addressing this
142 diversity. Below we describe the structure of each STA; specifics of data
143 availability are described in the sections that follow.

144

145 STA-1E began full operation in 2006-2007 and consists of three flow-ways;
146 Eastern, Western, and Central. Due to data availability only the Central Flow-way
147 was analyzed here. STA-1W's Eastern and Western flow-ways were in operation
148 from 1994 as the Everglades Nutrient Removal (ENR) project, with an additional
149 Northern flow-way constructed in 2000. All three flow-ways were analyzed. STA-

150 2 Cells 1-3, each single-cell flow-ways, were operational from 2000 onwards.
151 Additional cells, 4-8, involve multi-cell flow-ways and became operational
152 between 2008 and 2012 but were not studied here due to insufficient data
153 availability. STA-3/4 consists of three flow-ways (Flow-ways 1, 2 and 3) and
154 became operational in 2004; all were included in analysis. STA-5 originally
155 consisted of three flow-ways, denoted Flow-ways 1, 2 and 3; each consisting of a
156 combination of two cells. Flow-ways 1 and 2 became operational in 1999; Flow-
157 way 3 in 2008. Flow-ways 4 and 5 were later added, flow-capable in 2010, but
158 not studied here. Combination with STA-6 to form STA-5/6 added three
159 additional flow-ways; 6, 7 and 8, of which Flow-ways 7 and 8 are single cell flow-
160 ways (operational in 1998), and Flow-way 6 (not analyzed) couples two cells (6-
161 4, flow-capable in 2010 and 6-2, constructed in 2006).

162

163 **2.2. Data quality control**

164

165 Quality control checks were performed on all datasets at various stages of the
166 data compilation. Blank or null records were treated as no data and not zero. For
167 soil and floc data, parameter values were reported within specific ranges of the
168 profile, typically ranging from 0 to 10 cm. Some records included data on the
169 upper profile (0-10 cm), lower profile (10-30 cm), and full profile combined (0-
170 30 cm). In some cases soil nutrients within selected STA cells were measured at
171 variable depth increments (e.g. 0-2, 2-4, 4-6 cm, etc.). In such cases, all
172 parameters for relevant increments were averaged into a single 0-10 cm field for
173 analysis to ensure consistency across the dataset (including bulk density). In
174 some other cases, the sampling depth of the upper profile did not reach 10 cm,

175 but these were still marked as the upper profile. The full profile value was very
 176 rarely given, and was calculated only for the datasets that were subsequently
 177 used in the data mining analysis. In these instances, the average of the upper and
 178 lower profile was used.

179

180 **2.3. Data Analysis**

181 **2.3.1. Preparation of datasets for data mining**

182

183 The following rules were applied for inclusion of the data measured within the
 184 STAs: (1) There must be at least 10 observations for a given STA cell and year
 185 (an arbitrary cutoff point but sufficient to allow the calculation of meaningful
 186 statistics) and (2) There must be at least one instance of at least 10 observations
 187 per year within all STA cells in a flow-way. Seasonality at temporal scales finer
 188 than full years was not considered here. Additionally, any GIS data with full
 189 coverage of STA cells were considered. These included vector maps of vegetation
 190 class and habitat, NDVI rasters, and topography rasters representing the
 191 elevation differences of the STA floor at various year intervals. The resulting
 192 flow-ways included in data mining and their available data are listed in Table 1.

193 **Table 1.** List of flow-ways included in interpolation and their available data
 194 including years and number of observations (n). Surface water quality data are
 195 from transects internal to each treatment cell.

<i>STA</i>	<i>Flow-way</i>	<i>Cells</i>	<i>STA Data Availability</i>
STA-1E	Central	3 to 4N to 4S	Soil/floc (2004, 07, 09, 10; n=97) Surface water (2013; n=16) Macrophyte nutrients (2009; n=46) Hyper-spectral imagery (2011-12)
STA-1W	Eastern Western	1A and 1B to 3 2A and 2B to 4	Soil/floc (Eastern/Western FW only: 1995-97, 99; all FW: 2003-08, 10; n=1006) Surface water (2003, 04, 09-13; n=2689)

	Northern	5A to 5B	Macrophyte nutrients (Eastern/Western FW only: 1996, 97; all FW: 2003, 04, 08-10; n=262) Hyper-spectral imagery (2011-12)
STA-2	Flow-way 1	1	Soil/floc (2003, 07, 09-11; n=830)
	Flow-way 2	2	Surface water (2003-10, 13, 14; n=1126)
	Flow-way 3	3	Macrophyte nutrients (2003, 09, 10; n=91) Hyper-spectral imagery (2011-12)
STA-3/4	Flow-way 1	1A to 1B	Soil/floc (2004, 07, 10; n=1272)
	Flow-way 2	2A to 2B	Surface water (2003-10, 13, 14; n=1134)
	Flow-way 3	3A to 3B	Macrophyte nutrients (2010-12; n=58) Hyper-spectral imagery (2011-2012)
STA-5/6	Flow-way 1	1A to 1B	Soil/floc (FW 1/2: 2002, 03, 07-11; n=617.
	Flow-way 2	2A to 2B	FW 7/8: 2003, 07-11; n=138)
	Flow-way 7	5	Surface water (FW 1/2: 2013; n=74)
	Flow-way 8	3	Macrophyte nutrients (FW 1/2: 2002, 03; n=147. FW 7/8: 2003; n=31) Hyper-spectral imagery (all FW: 2011-12)

196

197

198 2.3.2. Interpolation of flow-way data within STA cells

199

200 Interpolation was done using an Empirical Bayesian Kriging (EBK) algorithm.

201 For Bayesian geostatistical analysis, we used the Gaussian Spatial Linear Mixed

202 Model as formulated by Diggle et al. (1998) without fixed effects:

203

$$Y(s_i) = W(s_i) + \varepsilon$$

204

205 where the random variable $Y(s_i)$ is an $n \times 1$ vector of observed values at

206 locations s_1, s_1, \dots, s_i ; W represents the spatial random effect which is a Gaussian

207 process with mean of 0, variance of σ^2 (partial sill) and correlation function

208 $R(h; \varphi)$, for which we selected an exponential correlation function: $R(h; \varphi) =$

209 $\exp(-\frac{h}{\varphi})$; and ε is an $n \times 1$ vector of errors with mean of 0 and variance of

210 τ^2 (nugget variance). These semivariogram parameters were estimated using
211 restricted maximum likelihood (REML). The EBK tool produced 1137 pairs of
212 interpolated and standard error maps which, together with other spatial
213 datasets available (described above in 2.3.1), were sampled with 100 randomly
214 distributed points (separated by at least fifty feet) within each STA cell.

215

216 **2.3.3. Multivariate Analysis**

217

218 Multivariate analysis used a combination of exploratory and modeling tools to
219 identify underlying patterns in the data. Within each treatment flow-way, data
220 from all available years were pooled to facilitate a single, data-rich analysis. For
221 initial calculation of summary statistics, the record set within each cell
222 containing the greatest number of observations for each year of coverage was
223 selected, and the mean and standard deviation of TP measurements were
224 calculated across all recorded years in Microsoft Excel (Microsoft, 2003). The
225 mean and standard deviation of key soil nutrients (i.e. total phosphorus, nitrogen
226 and carbon) were calculated for entire STAs. Principal components analysis
227 (PCA) and clustering analysis (CA) were used in an exploratory mode using JMP
228 (SAS, 2013); PCA to determine the main axis of variation the datasets, and CA to
229 determine if there were any meaningful groups in the observations. The primary
230 goals were: (a) to determine if there are any consistent main drivers of variation
231 across the flow-ways (i.e. do the flow-ways and STAs behave consistently across
232 the board, or is each a unique system responding to unique operational
233 circumstances); and (b) within each flow-way, to determine if there are natural
234 groupings of multivariate data (e.g. are observations from areas around the

235 inflow sufficiently similar in floc, soil and vegetation characteristics to cluster,
236 and sufficiently distinct from other areas). We used a combination of Ward's and
237 *k*-means clustering methods (Corstanje et al., 2009). Ward's is a minimum
238 variance, hierarchical clustering method which produces a scree plot, that in turn
239 allows us to both identify the optimal number of clusters and establish the seeds
240 which are then used to run the *k*-means clustering process. This was then
241 followed by Stepwise Canonical Discriminant (SCD) analysis in JMP (SAS, 2013)
242 to help identify the primary drivers of the clusters.

243

244 Subsequently, we applied a set of non-linear, hierarchical structured models
245 using Statistica (StatSoft, 2014) to predict surface water TP concentrations
246 (Classification and Regression Trees; CART). Where no surface water TP data
247 were available (as was the case in 10 out of 24 cells: STAs 1E, 2 Cell 2 only, 3/4,
248 and 6), floc TP was substituted as the best available indicator of TP and its
249 drivers in the flowing system. The CART approach has a number of advantages;
250 the method is not sensitive to non-normal data, it accepts categorical as well as
251 continuous data (needed as soil series and soil parent material are categorical,
252 whereas soil organic matter is continuous) and it is not confounded by the
253 presence of non-linear relationships (Breiman et al., 1984; McCune and Grace,
254 2002). Bayesian Belief Networks (BBNs), having similar advantages in their
255 ability to handle non-normal and categorical data, were also created using Netica
256 (Norsys, 2014) to predict the most recently available NDVI and TP
257 (preferentially in surface water if available, otherwise in floc or soil as described
258 above) in each cell. BBNs are graphical probabilistic models; graphical in that
259 they represent the variables that affect the response of interest (e.g. floc or

260 surface water P) in the form of a network, and probabilistic in that the
261 relationships between the drivers and response are conditioned by a probability
262 (Taalab et al., 2015). Bayesian inference is thus based on a set of prior
263 probabilities that can be updated as new information becomes available. In this
264 case, some knowledge of potential drivers of P dynamics was available from the
265 CART analysis and a review of the existing P process literature; the network thus
266 consisted of those variables that the previous CART models identified as drivers.
267 For both CART and BBN approaches, model fitness and the strongest predictor
268 variables were of primary interest.

269

270 **3. Results**

271 **3.1. Summary Statistics**

272

273 Data on TP from internal surface water transects and TP, total carbon (TC) and
274 nitrogen (TN) from soil samples in all STAs and across all available years were
275 pooled and their summary statistics calculated (Tables 2 and 3), but
276 distributions were highly variable in terms of timing, data type, number of
277 observations, and data were not available or complete for all cells and flow-ways.
278 Cell 2A in STA-5/6 achieved the highest overall mean internal surface water TP
279 (0.216 mg L^{-1}) followed by STA-1W's Cell 5A (0.129 mg L^{-1}). The Cells with the
280 lowest mean internal surface water TP were STA-3/4's Cell 3B (0.012 mg L^{-1})
281 and STA-1W's Cell 4 (0.024 mg L^{-1}). Variability was present in the data, both
282 within sets of records and between different years and cells; most standard
283 deviations tended to fall proportionally between 30% and 80% of their
284 associated means. Total soluble phosphorus (TSP) and soluble reactive

285 phosphorus (SRP) in internal surface water were variable in their proportional
 286 relationship with TP (not shown); combined across all STAs, TSP averaged
 287 roughly half of TP (59.2%) with a standard deviation of 15.4%, and SRP averaged
 288 28.1% of TP with a standard deviation of 14.6%. As these statistics summarize
 289 the data for entire treatment cells they do not address spatial patterns within
 290 individual cells (this is explored below in section 3.3); however in flow-ways
 291 composed of multiple cells, an apparent trend of decreasing mean TP was visible
 292 along the length of the flow-ways from the summary statistics, evidencing the
 293 removal of phosphorus from surface water as it flows through the STAs. The
 294 greatest proportional drop was in Flow-way 2 in STA-5/6, where Cell 2A
 295 exhibited a mean TP of 0.216 mg L⁻¹ and Cell 2B a mean of 0.062 mg L⁻¹.

296

297 **Table 1:** Summary statistics for all combined data on total surface water
 298 phosphorus [mg L⁻¹] sampled within the STAs (internal surface water transect).
 299 SD = Standard Deviation, N = number of observations. Values marked 'n/a'
 300 represent cells where summary data were insufficient for calculation of
 301 summary statistics.

<i>STA</i>	<i>Flow-way</i>	<i>Cells</i>	<i>Mean</i>	<i>SD</i>	<i>N</i>
STA-1E	Central	3	n/a	n/a	0
		4N	0.108	0.017	16
		4S	n/a	n/a	0
STA-1W	Eastern	1A	0.106	0.049	8
		1B	0.065	0.044	159
		3	0.030	0.018	95
	Western	2A	0.123	0.069	77
		2B	0.047	0.022	89
		4	0.024	0.012	70
	Northern	5A	0.129	0.051	54
5B		0.071	0.079	699	
STA-2	Flow-way 1	1	0.044	0.036	197
	Flow-way 2	2	n/a	n/a	0
	Flow-way 3	3	0.034	0.024	606
STA-3/4	Flow-way 1	1A to 1B	n/a	n/a	0
	Flow-way 2	2A to 2B	n/a	n/a	0
	Flow-way 3	3A	0.037	0.005	4

		3B	0.012	0.001	42
STA-5/6	Flow-way 1	1A	0.064	0.048	12
		1B	0.031	0.023	16
	Flow-way 2	2A	0.216	0.074	12
		2B	0.062	0.045	16
	Flow-way 7	5	n/a	n/a	0
	Flow-way 8	3	n/a	n/a	0

302
303

304 **Table 3:** Summary statistics for all combined data on total soil phosphorus [TP; mg kg^{-1}], total carbon [TC; g kg^{-1}] and total nitrogen [TN; g kg^{-1}] sampled within
305 the STAs. SD = Standard Deviation, N = number of observations.
306

<i>STA</i>	<i>Soil TP (mg kg^{-1})</i>			<i>Soil TC (g kg^{-1})</i>			<i>Soil TN (g kg^{-1})</i>		
	<i>Mean</i>	<i>SD</i>	<i>N</i>	<i>Mean</i>	<i>SD</i>	<i>N</i>	<i>Mean</i>	<i>SD</i>	<i>N</i>
STA-1E	241	207	294	85.2	63.3	294	5.7	4.3	294
STA-1W	550	237	1405	432	57.6	1322	26.5	4.2	1319
STA-2	611	250	1166	392	51.2	1078	23.1	3.7	1078
STA-3/4	718	243	1858	346	74.2	1857	22.1	5.0	1857
STA-5/6	727	315	952	285	111	783	20.5	7.8	783

307

308

309 Data for TP, TC and TN in soil and floc across the STAs were analyzed at the STA
310 level. STA-5/6 exhibited the highest mean levels of soil TP (727 mg kg^{-1}), while
311 STA-1W achieved the highest values for both mean TC (432 g kg^{-1}) and mean TN
312 (26.5 g kg^{-1}). STA-1E had the lowest mean values for all three nutrients; 241 mg
313 kg^{-1} TP, 85.2 g kg^{-1} TC, and 5.7 g kg^{-1} TN. Variability was highest in STA-5/6
314 across all three nutrients; exhibiting a standard deviation of 315 mg kg^{-1} TP, 111
315 g kg^{-1} TC, and 7.8 g kg^{-1} TN. TP variability was lowest in STA-1E (standard
316 deviation of 207 mg kg^{-1}), while STA-2 displayed the lowest variability for both
317 TC (51.2 mg kg^{-1}) and TN (3.7 g kg^{-1}). Note that these statistics represent
318 averages across entire treatment cells or STAs; Table 3 reports the associated
319 variability (as standard deviations).

320

321 **3.2. Multivariate Analysis Results**

322

323 Principal Component Analysis (PCA) results are characteristically not
324 straightforward to interpret and do not involve clear cutoffs to determine
325 whether or not a component variable can be considered specifically important or
326 unimportant, so focus was placed on determining and reporting those variables
327 that were clearly the strongest drivers and/or recurred consistently across STAs.
328 Results varied by cell, but the most commonly identified variables related to soil
329 TC, soil TN, soil and floc bulk density (BD), soil and floc TP, and soil and floc ash-
330 free dry weight (AFDW) as the greatest contributors to variability in the data
331 (Table 4). Cluster analysis identified 3 or 4 clusters in most cells, with spatial
332 structure to cluster membership apparent in some but not all cells (Table 5).

333

334 **Table 4:** Summary of the main outcomes from Principal Component Analysis
335 (soil/floc/surface water parameters separated by semicolon). Abbreviations:
336 total phosphorus (TP), total carbon (TC), total nitrogen (TN), bulk density (BD),
337 sulfur (S), calcium (Ca), iron (Fe), macrophyte nutrient (macro), exchange
338 capacity (exc), surface water (sw), alkalinity (Alk) ash-free dry weight (AFDW).
339 Note that data availability was not consistent (e.g. few surface water
340 observations in STA-1E Cells 3 and 4S) so PCA may not accurately reflect the
341 importance of underrepresented variables in some cells.

<i>STA</i>	<i>Flow-way</i>	<i>Cell</i>	<i>PCA main variables</i>	<i>% var explained by PC1,..,PC3</i>
STA-1E	Central	3	Soil TC, TN, AFDW, BD, TP, Ca	80.25
		4N	Soil AFDW, BD, TC, Ca, Fe, TP	79.58
		4S	Soil AFDW, BD, TC, TN, TP, Ca, Fe	77.77
STA-1W	Northern	5A	Soil AFDW, BD, TC, TN, TP; floc AFDW; sw TP	83.23
		5B	Floc BD, TC, AFDW; sw Ca, P	68.01
	Eastern	1	n/a*	57.49
		3	Soil Al exc, Fe exc, TN, Alk, AFDW, BD, K; sw TP	76.59
	Western	2	Soil Fe, BD, TC; sw TP, Ca, AFDW	81.34

		4	Soil AFDW, Fe, TC, TN; sw Ca, TP	81.66
STA-2	1	1	Soil TC, TN, TP; floc TC; sw TP	75.09
	2	2	Floc BD, TC, TN, TP	72.19
	3	3	Soil macroDryWt, TC; sw Ca, TP	77.98
STA-3/4	1	1A	Soil BD, TN; floc BD, TP	72.54
		1B	Soil TP, TN, BD; floc TN; sw TN	71.78
	2	2A	Soil BD, TC, TN; floc BD, TC, TN	63.12
		2B	Soil TC, TP, TN, BD	71.71
	3	3A	Soil TC, TN, BD, TP; sw Ca, TP	74.21
		3B	Floc TC, TN; sw Ca, P	69.59
STA-5/6	1	1A	Soil macro TN, Fe, BD; floc DryWt; sw Ca	72.11
		1B	Soil TC, TN, S, TP; floc dryWt, AFDW; sw TP	79.28
	2	2A	Soil TC, TP, AFDW; floc BD; sw TP	77.99
		2B	Soil TN, TP, BD, TC	77.17
	7	5	Soil AFDW, TC, TN, Ca; floc moisture	77.90
	8	3	Soil AFDW, Fe, macro AFDW, TC; floc AFDW	77.87

342 * STA-1W Cell 1 PCA results consisted of similar and low average values, not
343 highlighting any particular driving variables.

344

345 **Table 5:** List of analyzed flow-ways by age, number of clusters and observed
346 spatial pattern of clusters (maps of cluster patterns available in supplementary
347 material).

STA	Flow-way	Oper. start year	Cell	No. clusters	Observed cluster pattern
STA-1E	Central	2006/7	3	4	Zonal
			4N	3	Zonal
			4S	3	Zonal
STA-1W	Eastern	1994	1	3	Zonal gradient
			3	4	Zonal gradient
	Western	1994	2	4	Zonal gradient
			4	4	Tenuous zonal gradient
	Northern	2000	5A	4	Zonal
			5B	4	Tenuous zonal gradient
STA-2	Flow-way 1	2000	1	5	Zonal gradient
	Flow-way 2	2000	2	4	Zonal
	Flow-way 3	2000	3	4	Zonal gradient
STA-3/4	Flow-way 1	2004	1A	5	Zonal
		2004	1B	4	Zonal

	Flow-way 2	2004	2A	4	Zonal
		2004	2B	3	Zonal
	Flow-way 3	2004	3A	3	Tenuous zonal gradient
		2004	3B	5	Zonal
STA-5/6	Flow-way 1	1999	1A	6	Tenuous zonal gradient
		1999	1B	3	Tenuous zonal gradient
	Flow-way 2	1999	2A	4	Zonal
		1999	2B	3	Zonal
	Flow-way 7	1998	5	3	Zonal gradient
	Flow-way 8	1998	3	5	Zonal gradient

348

349

350 CART analysis consistently found the strongest predictor variables for surface
351 water and floc TP to be other variables relating to P content (i.e. P in different
352 forms such as SRP, etc.) in soil, floc, and surface water; soil and floc BD; and soil
353 and floc TN. Measures relating to AFDW, TC and Ca also showed occasional
354 influence but were less widespread. Maps of CART model standard error by
355 location (not pictured) did not generally reveal any spatial relationships with
356 direction of flow, but did in some cases reveal zonal structures similar to the
357 cluster analysis (described below in 3.3).

358

359 Analysis with BBNs identified the strongest consistent predictors of recent year
360 NDVI to be variables relating to: vegetation type and cover, NDVI from previous
361 years, surface water TP, soil and floc TN, and soil and floc TC. BBNs predicting
362 surface water TP were most influenced by: other forms of surface water P, soil
363 BD, soil TN, soil TC, and soil TP.

364

365 **3.3. Spatial Trends**

366

367 Spatial patterns varied to a degree among treatment flow-ways. For instance, floc
368 and macrophyte characteristics dominated the models which predicted surface
369 water TP in STA-5/6; soil physical properties (e.g. bulk density) described many
370 of the spatial patterns in the treatment flow-ways of STA-3/4, etc.
371 Notwithstanding this, some general observations can be made regarding all
372 treatment flow-ways: (1) there are clear zonal patterns consistently present in
373 these systems that are, in many cases, independent of the direction of flow and
374 do not exhibit a simple linear gradient (Figure 2 shows STA-3/4 Flow-way 3 as
375 an example of purely zonal pattern; other examples include Flow-ways 1 and 2 in
376 the same STA and STA-1E's Central Flow-way, shown in supplementary
377 material); however these zonal patterns appear to align along the direction of
378 flow in the case of some older STAs and flow-ways (Figure 3 shows STA-1W's
379 Eastern flow-way as an example of zone-based gradient pattern; other examples
380 include STA-1W's Western Flow-way, STA-2's Flow-ways 1 and 3, and STA-5/6's
381 Flow-ways 7 and 8, shown in supplementary material and summarized in Table
382 5); (2) There is some consistency in the spatial arrangement of these zones over
383 the treatment flow-ways, such as surface water TP concentration being highest
384 close to the inflow structures and there closely associated with a zone of higher
385 floc and soil TP concentrations. Following these points, there is rarely any
386 further consistency in the spatial organization of zones, or in their
387 characterization, across flow-ways; but 3) soil TN often becomes an important
388 factor characterizing the zone around the outflow (e.g. STA-1W, STA-3/4).

389

390 **4. Discussion and Conclusions**

391 **4.1. Summary Statistics**

392

393 Two results stood out from the cell-wide summary statistics that were consistent
394 with expectations. Firstly, the lowest mean values of internal surface water TP
395 were found in flow-ways present in STAs 2 and 3/4, which have been previously
396 cited as being two of the best-performing STAs for P removal (Pietro and Ivanoff,
397 2015). Secondly, all flow-ways consisting of multiple cells exhibited a trend of
398 decreasing TP along the length of the flow-way (cell-wide summary statistics did
399 not consider internal spatial patterns of single-cell flow-ways; these are
400 discussed below), demonstrating the effects of P removal by the system at the
401 STA scale. Taken broadly, this is consistent with the expectation that wetlands
402 experiencing a uniform sheet flow should exhibit P decreases along a
403 longitudinal flow-based gradient (Walker and Kadlec, 2011).

404

405 **4.2. Multivariate Analysis**

406

407 In considering the outputs from the data-mining analysis for the flow-ways; PCA
408 is a general dimension reduction technique in which the underlying variation is
409 maintained. It was used here because it is one of the primary steps in any
410 multivariate data analysis as well as an effective way to represent variation in
411 the data. Generally the PCA was successful, with an average of 75% of the
412 variation explained. The most common variables identified as influential in the
413 PC loadings were soil TC, soil TN, soil and floc BD, soil and floc TP and soil and
414 floc AFDW. It should be noted that this particular analysis does not take into
415 account non-continuous data (e.g. categorical variables such as soil series and
416 parent material). In essence, the outcome from this analysis is an effective

417 summarization of the data but with little further insight into drivers, mainly
418 highlighting that most of the within cell/within flow-way variation is driven by
419 sediment nutrient concentrations and, to a lesser degree, floc TC and nutrient
420 content.

421

422 Cluster analysis resulted in cluster memberships that could be assigned to the
423 original data, revealing spatial patterns and structure in the data. Of interest here
424 were two points; do the data resolve clearly in clusters, and if so, how many (i.e.
425 how many classes of data are there in an STA flow-way), and are these classes
426 meaningful in any way? In general, most cells could be described by 3 to 5
427 clusters and only in one case (STA-5/6 Cell 1A; 6 clusters) were more clusters
428 needed (see Table 5). Clusters consistently grouped spatially into zone features
429 which did not appear to be tied to cell location within the flow path in many
430 cases; however in some cells these zonal features were observed to align along
431 the direction of flow. While not an unequivocal relationship, these 'zone-based
432 gradient' patterns appeared more likely to occur in older STAs and flow-ways
433 (Table 5). Patterns seemed only tenuously related to flow path at best in STAs-1E
434 and -3/4 (completed in 2007 and 2004, respectively), and generally more
435 obviously following the flow gradient in STA-1W (completed in 1994-2000),
436 STA-2 (completed in 2000), and STA-5/6 (completed in 1998/9).

437

438 The CART and BBN analyses both revealed similar relationships and driving
439 variables in the data. Surface water TP was found to share consistently strong
440 linkages with other forms of phosphorus in surface water (e.g. SRP and TSP) as
441 well as in floc and soil. Nitrogen, carbon, and bulk density in soil and floc also

442 factored in frequently; this highlights the potential importance of soil properties
443 to P dynamics in the STAs, as well as the possibility of coupled cycles wherein P,
444 N, and possibly C dynamics share co-dependencies and interrelationships.

445

446 **4.3. Observed Relationships and Drivers of P Dynamics**

447

448 It is evident from studies in the Everglades and elsewhere (Bayley and Mewhort,
449 2004; Bostic and White, 2007; Gu and Dreschel, 2008; Riggsbee et al., 2012), that
450 plant communities actively regulate P dynamics in wetlands. In the STAs, low
451 levels of water column P are achieved using strategic combinations of SAV and
452 EAV to address P in different forms and in different stages of the flow-ways
453 (Chen et al., 2015). In projecting this fact on the data mining exercise, one would
454 expect the spatial patterns of soil P to reflect plant community composition, and
455 plant communities would be expected to be a strong determinant in any
456 predictive model for soil or flocculent P. In our analysis this was only rarely the case;
457 however these effects may be obscured by the fact that much of the available
458 data on vegetation composition were categorical (e.g. vegetation class and
459 habitat type; NDVI being the notable exception as a continuous variable), and
460 thereby only possible to include in CART and BBN analyses. Both CARTs and
461 BBNs modeling surface water TP did not commonly reveal vegetation-related
462 measures as key predictors, but BBNs predicting NDVI frequently did highlight
463 surface water TP as an important driver (i.e. TP did not appear driven by
464 vegetation, but vegetation appeared driven by TP). Linkages between TP and
465 vegetation therefore may not be direct or omnipresent, but our analysis shows
466 support for some relationships.

467

468 Where P is limiting, or effectually buried, and therefore not available for the
469 plant communities, this may be reflected as plant stress (i.e. P limitation), which
470 can be remotely determined using NDVI (Henrik, 2012). The hypothesis is that
471 the indication of effective functioning of an STA is that, in the lower reaches of a
472 flow path, the vegetation may become P-limited. As a first instance, predictive
473 modeling of NDVI should indicate whether this is responsive to floc and soil
474 nutrient status. For BBNs predicting NDVI this indeed was the case; the strongest
475 predictors consistently included floc and soil nutrients, along with surface water
476 TP and other measures of vegetation health and composition. Note however that
477 prolonged exposure to low P concentrations may trigger a shift in plant
478 community composition to species that are more adapted to the low levels; such
479 a shift would be reflected in categorical habitat variables but not necessarily by a
480 decrease in NDVI. This highlights the importance of vegetation-related measures
481 beyond NDVI, and in turn the importance of methods such as BBNs that can
482 consider categorical expressions of vegetation community.

483

484 **4.4. Spatial Patterns of P and their Implications**

485

486 The observation that consistent spatial patterns appear zonal rather than based
487 on simple gradients is probably the most significant finding of the data mining, in
488 that the processes controlling P in these systems operate in zones in the
489 treatment flow-way, rather than along a smooth linear gradient as would be the
490 expectation (see Table 5). These zones are observed repeatedly across STAs and
491 flow-ways, and are consistently present as modeling outcomes (e.g. cluster

492 analysis and CART outputs) and as such are unlikely to be a modeling artifact.
493 There are a number of implications from approaching the STA flow-ways as
494 zones rather than a simple gradient. From a research perspective, the relative
495 importance of different factors, transformation and transport pathways of P
496 occurs in spatial patterns, and the form and shape of these patterns indicates the
497 relative importance of particular pathways. Likewise, this affects the
498 experimental sampling design, as these would then target zones rather than
499 seeking to measure along a gradient (biased sampling). From a management
500 perspective, this could simplify management options in that the operation and
501 management strategies can be directed at particular zones within a treatment
502 flow-way rather than an entire cell or the full flow-way, particularly once the
503 drivers of these zones are better understood. Nevertheless, in older STAs (e.g.
504 STA-1W, -2, and -5/6) these zonal patterns appeared to align more frequently
505 and obviously with the direction of flow, suggesting that P dynamics may
506 function largely in zonal patterns but slowly shift toward a zone-based gradient
507 pattern over the operational time of an STA. Of particular note, STA-2 flow-way 3
508 exhibited a strong gradient pattern in the cluster analysis result and has been
509 previously studied as one of the longest-running and best-performing treatment
510 flow-ways (Juston and Debusk, 2011; Juston et al., 2013).

511

512 The finding of zonal patterns of P concentrations in the STAs (whether forming
513 zone-based flow gradients or not), rather than simple uniform gradients
514 decreasing along the axis of water flow, differs from previous findings and the
515 usual expectation of P dynamics in wetlands (e.g. Kadlec, 1999; Walker and
516 Kadlec, 2011). One possible explanation for this difference is that the treatment

517 cells may be wide enough to allow partial mixing of water rather than a relatively
518 uniform sheet flow; this would account for more complex patterns (Walker and
519 Kadlec, 2011). If true, this would have implications for the assumptions made in
520 future flow modeling efforts in the STAs, and require a more complex
521 interpretation of the system than a one-dimensional sheet flow. Chen et al.
522 (2015) cautioned that analyses focused solely on inflow and outflow P
523 concentrations, while useful, do not consider P removal processes internal to the
524 treatment cells, as well as recommending that future studies consider
525 multivariate relationships. Doing so here has enabled additional findings, such as
526 the potential importance of relationships between P and soil factors, and the
527 possibility of P-N coupled cycles impacting dynamics. This latter result, while not
528 widely explored previously, is consistent with previous findings in Water
529 Conservation Area 2A (WCA 2a) on P and N functional linkages (White and
530 Reddy, 2003). Corstanje et al. (2009, 2007) found evidence that areas enriched
531 with P in WC-2a are mediated by N related parameters, such as potentially
532 mineralizable N and related microbial extracellular enzymatic activities. In STA
533 areas closest to the inflow, as P is relatively plentiful, the cycling P is likely to be
534 co-mediated by N and its dynamics.

535

536 **4.5. Data-Mining Advantages and Future Research**

537

538 Previous studies have examined the extensive data now available for P dynamics
539 in the STAs (e.g. Chen et al., 2015; Juston et al., 2013; Pietro and Ivanoff, 2015),
540 but this is one of the first known studies to comprehensively make use of the
541 diverse data collected in the interior treatment cells and flow-ways (e.g., soils,

542 vegetation, internal water quality) and the first to do so at such a broad scale
543 through a data mining approach. Doing so has facilitated new findings and
544 understanding around the functional P dynamics of the STA systems. Approaches
545 making use of these techniques are valuable for identifying biogeochemical
546 relationships, and should be considered and further employed in future studies
547 of the STAs as well as other engineered wetlands where sufficient data are
548 available.

549

550 In addition, there remain a number of further considerations moving forward.
551 First, many links between plant community composition and P dynamics remain
552 unclear beyond known differences between EAV and SAV in P removal (e.g.
553 Dierberg et al., 2002; Juston and DeBusk, 2006). In particular, we suspect there is
554 an element of scale effect; where these processes occur and are important at
555 scales finer than we considered in this study. Second, the approach used here
556 focused on data mining techniques, and while effective for exploring patterns in
557 the data it lacks a detailed process understanding of P biogeochemistry. The
558 incorporation of process understanding and process models (e.g. first order
559 equations) into the more stochastic modeling environment considered in this
560 study could produce a set of hybrid models which would both reflect process
561 knowledge and understanding but also, critically, allow for scaling and mapping.
562 Such an approach could better explore the process-based reasons for the zonal
563 patterns observed here and their potential relationships with flow-way age.
564 Finally, future research should seek to effectively consider the interaction
565 between different datasets available from the STAs in order to rigorously
566 consider time series analysis and pulsed events. A future study which initiates

567 with a thorough decomposition of the STA inflow and outflow data (volume and
568 concentrations), considers the stochasticity of this data and then moves to
569 incorporate it in the models of flow-way behavior should generate significant
570 insights in the STA dynamics, and to what degree performance is related to
571 stochastic events (e.g. storms or droughts) vs. deterministic processes (e.g. P
572 biogeochemistry, SAV, periphyton). Eventually this will relate to a measure of the
573 resilience of these systems; expressed as their capacity to withstand pressures
574 and maintain long term performance.

575

576 **4.6. Conclusions**

577

578 In conclusion, the use of data mining approaches on STA treatment cell and flow-
579 way data has identified, in a very general sense, spatial patterns in these systems.
580 These patterns are consistently zone-based across all flow-ways, which suggests
581 that the flow-ways function first as zonal systems rather than simple linear
582 gradient systems. Our analysis suggests that the primary drivers of the spatial
583 distribution of P in many of these systems are related to soil characteristics, and
584 that the zonal patterns of P distribution may begin to follow the predominant
585 flow path over time. The data further suggest the importance of coupled cycles in
586 these systems; in other words, the movement and transformation of P is coupled
587 to that of N.

588

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591 (SFWMD) in support of the Restoration Strategies Science Plan (SFWMD 2013).

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593

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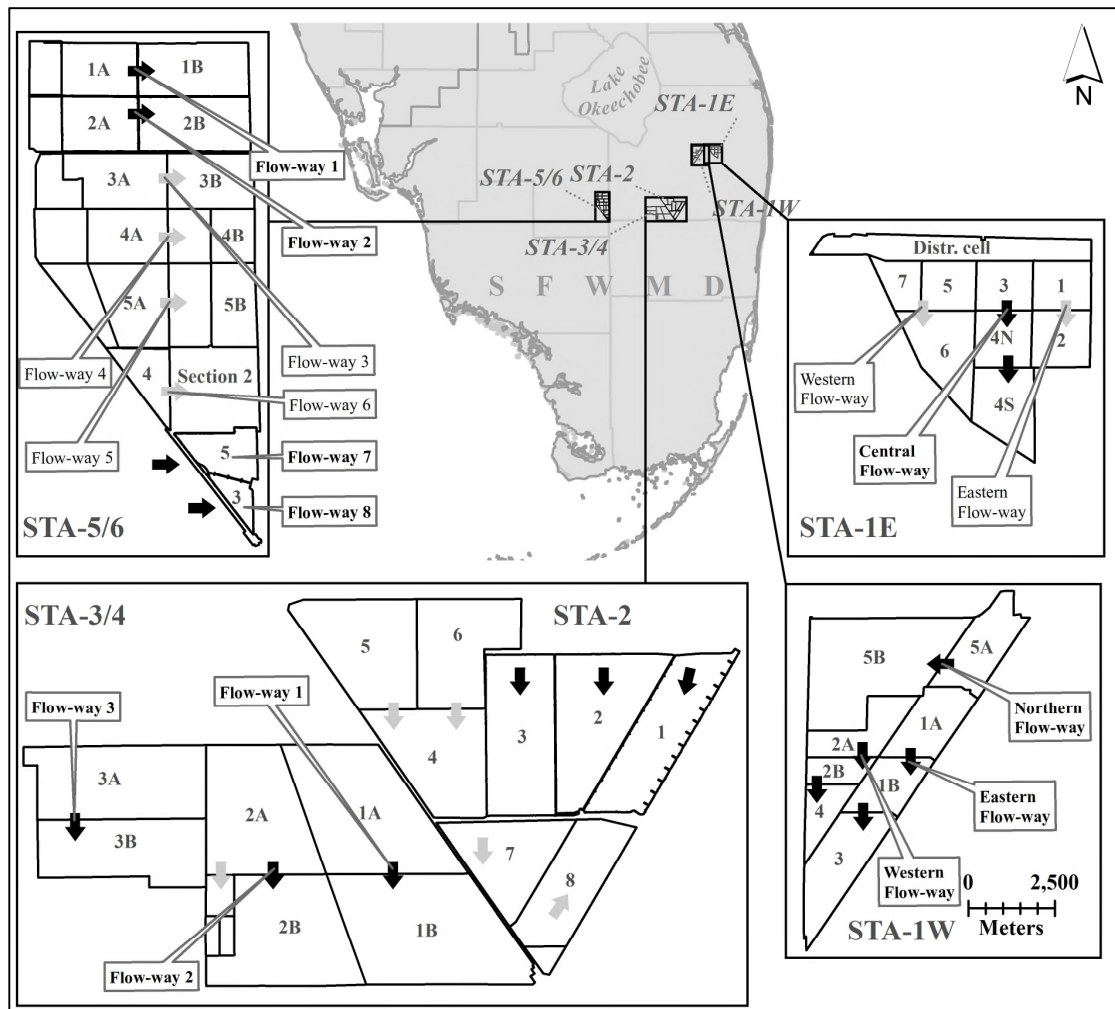
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711

712 **Figures**



713

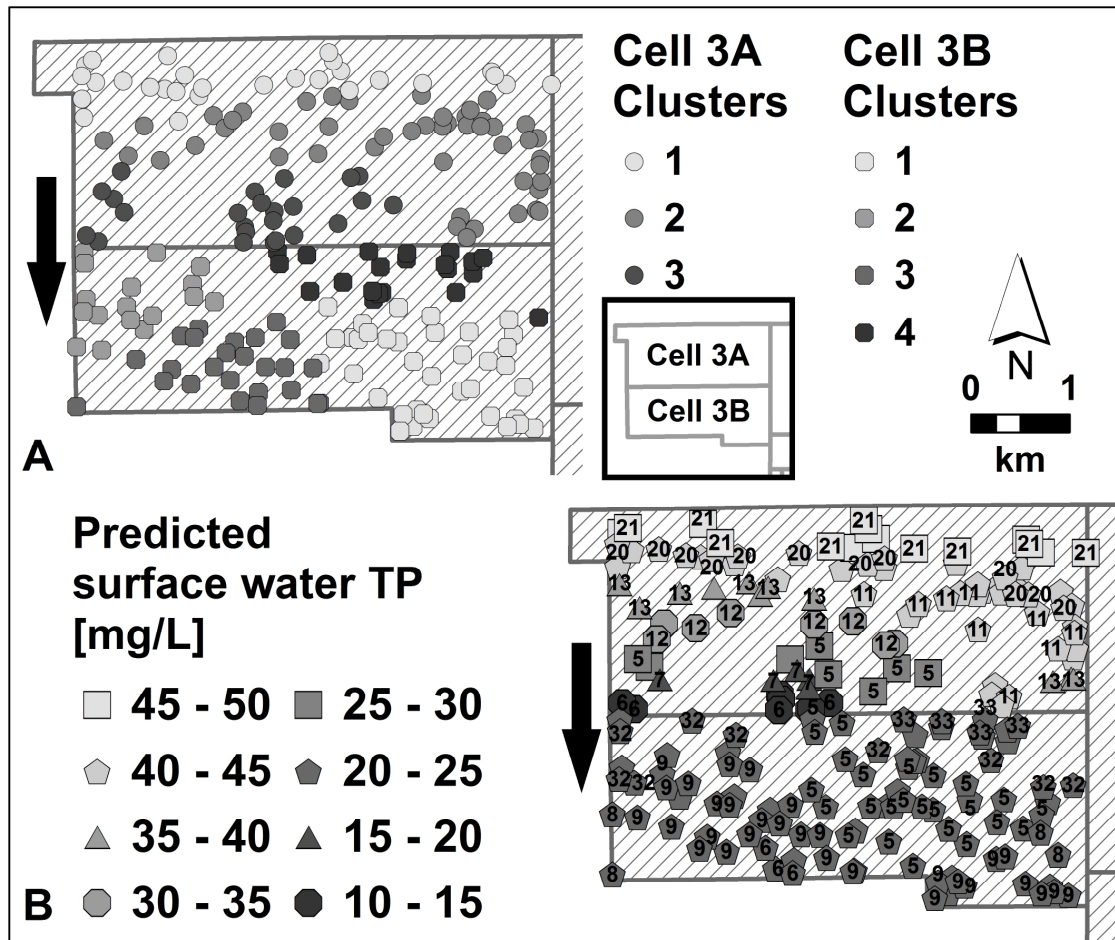
714 Figure 1: Locations of the Stormwater Treatment Areas in south Florida, USA,

715 indicating individual treatment cells and direction of flow. Bolded flow-way

716 names and darkened arrows denote flow-ways included in analysis.

717

718



719

720

Figure 2: Spatial patterns detected by cluster (A) and CART (B) analyses – an

721

example for STA-3/4 flow-way 3. Image B represents the distribution of

722

CART nodes (symbol numbers represent the number of nodes in the CART

723

model) corresponding to the prediction of surface water total P

724

(concentration denoted by symbol color). Note that patterns are

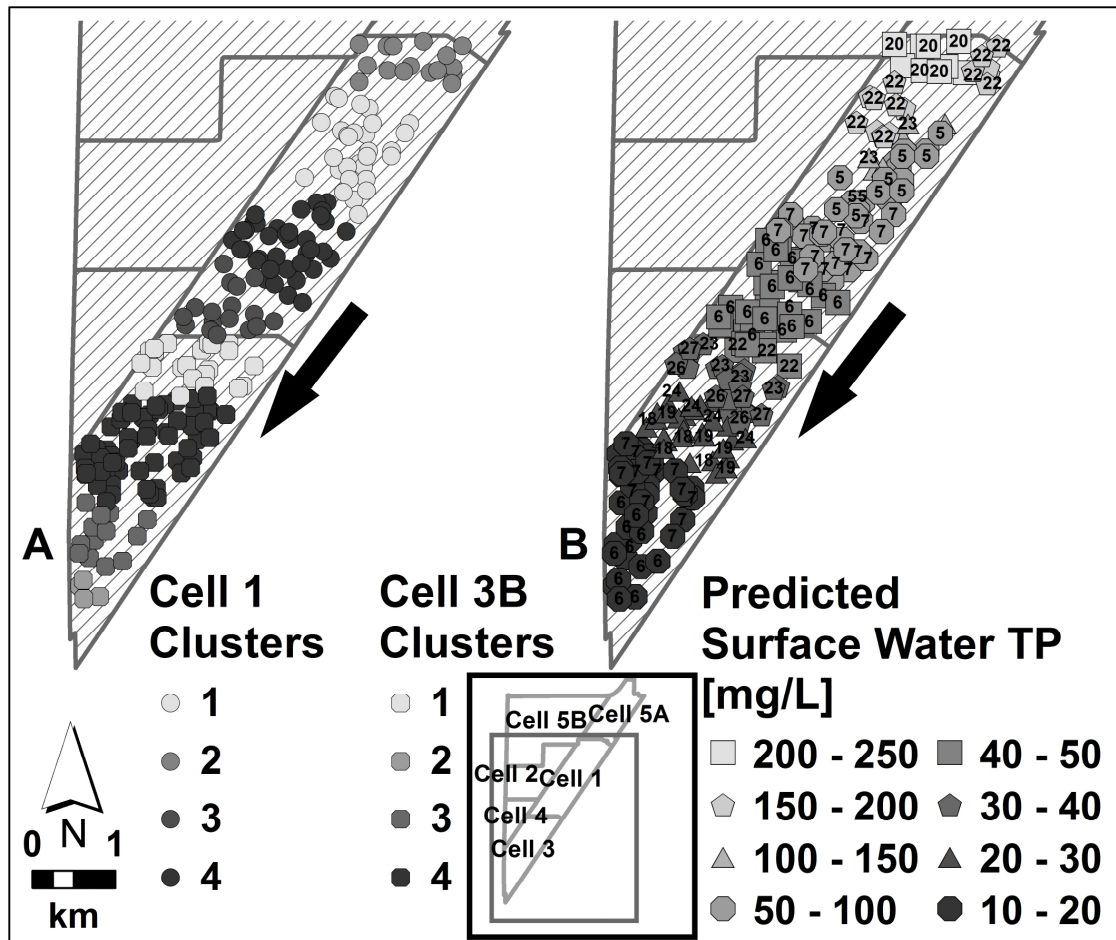
725

predominantly zonal and only tenuously aligned with flow direction.

726

727

728



729

730 Figure 3: Spatial patterns detected by cluster (A) and CART (B) analyses – an
 731 example for STA-1W Eastern flow-way. Image B represents the distribution
 732 of CART nodes (symbol numbers represent the number of nodes in the
 733 CART model) corresponding to the prediction of surface water total P
 734 (concentration denoted by symbol color). Note that zonal patterns appear
 735 largely aligned with flow direction, indicating a gradient-based behavior to
 736 the individual zones.

737

A datamining approach to identifying spatial patterns of P forms in the Stormwater Treatment Areas in the Everglades, US

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Supplementary Materials

These maps show the K-mean clusters and tree nodes resulting from Classification and Regression Trees (CARTs) analysis performed within particular flow-ways of the Stormwater Treatment Areas (STAs) that had sufficient data to do so. Please note that, in the case of CARTs, the results are only shown for the flow-ways with availability of data on total surface water phosphorus.

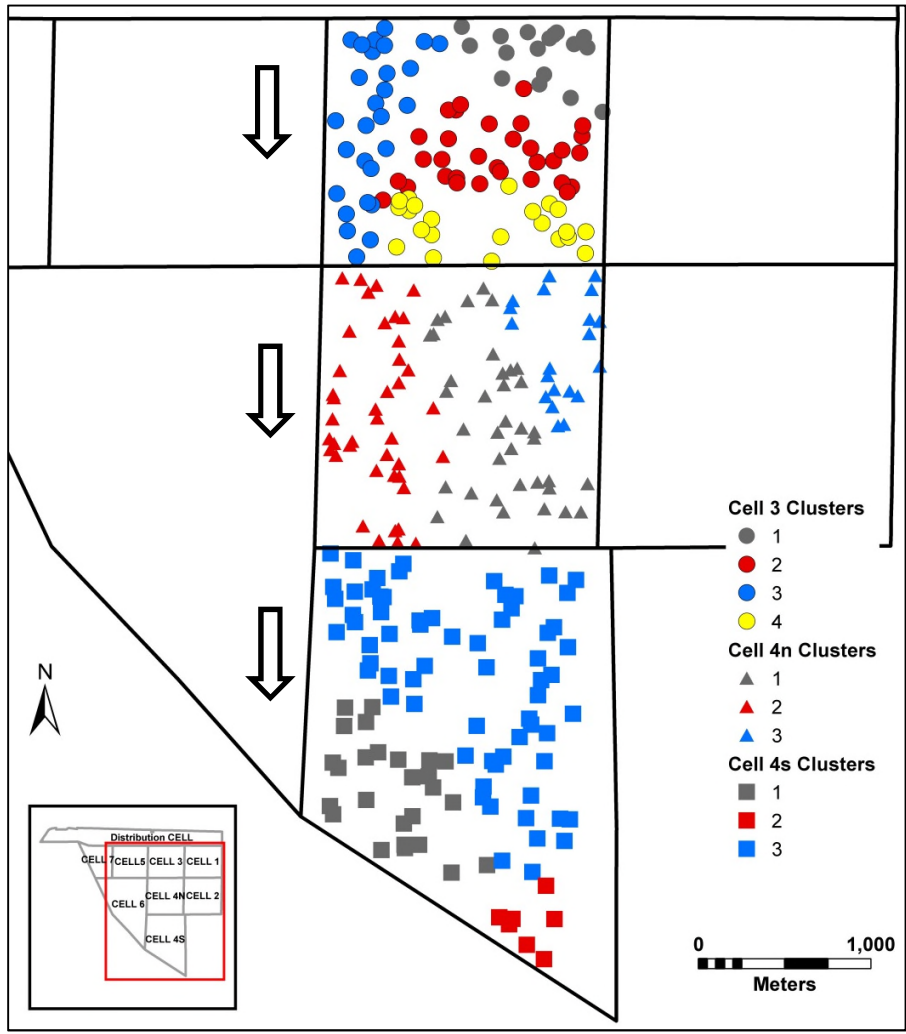
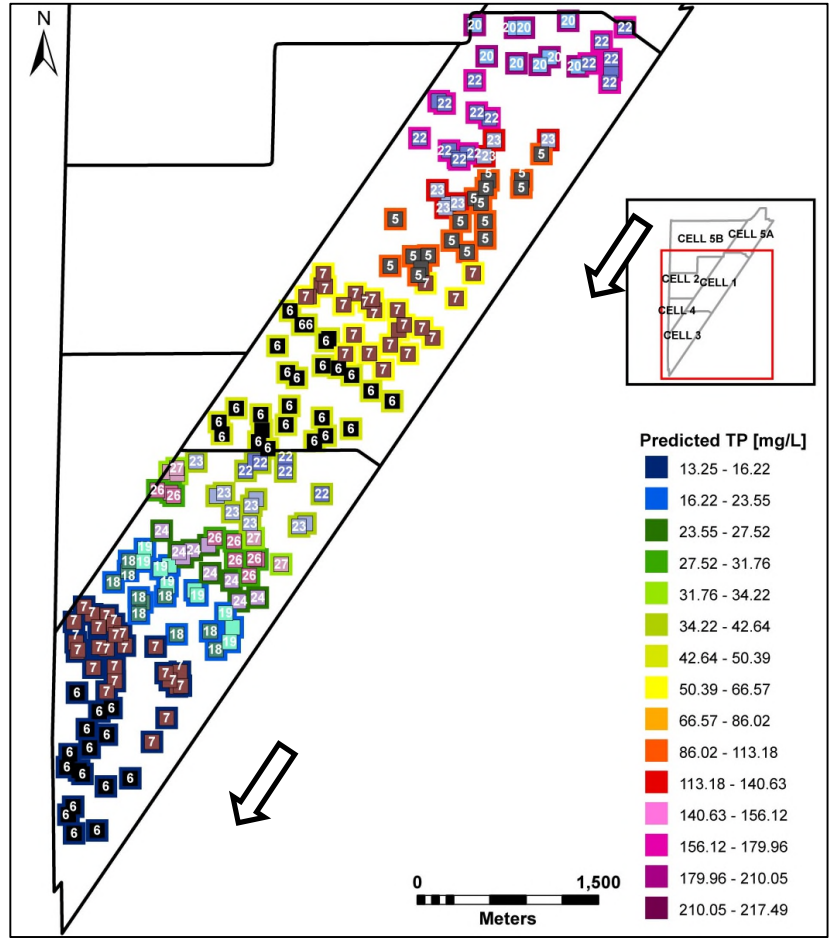
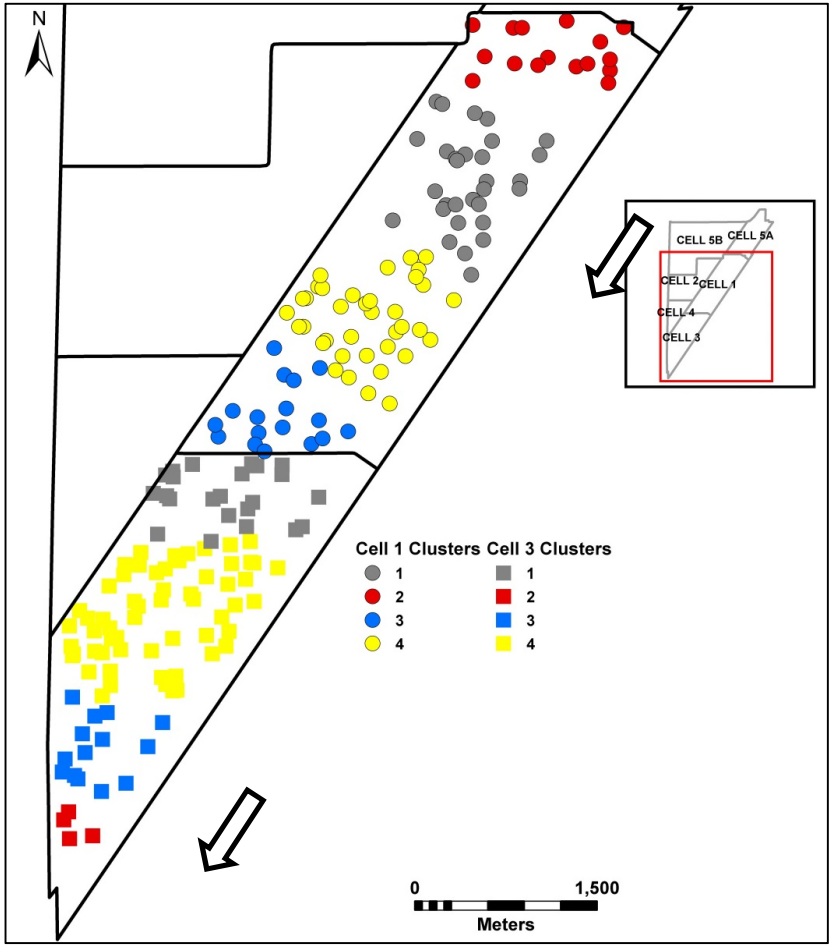


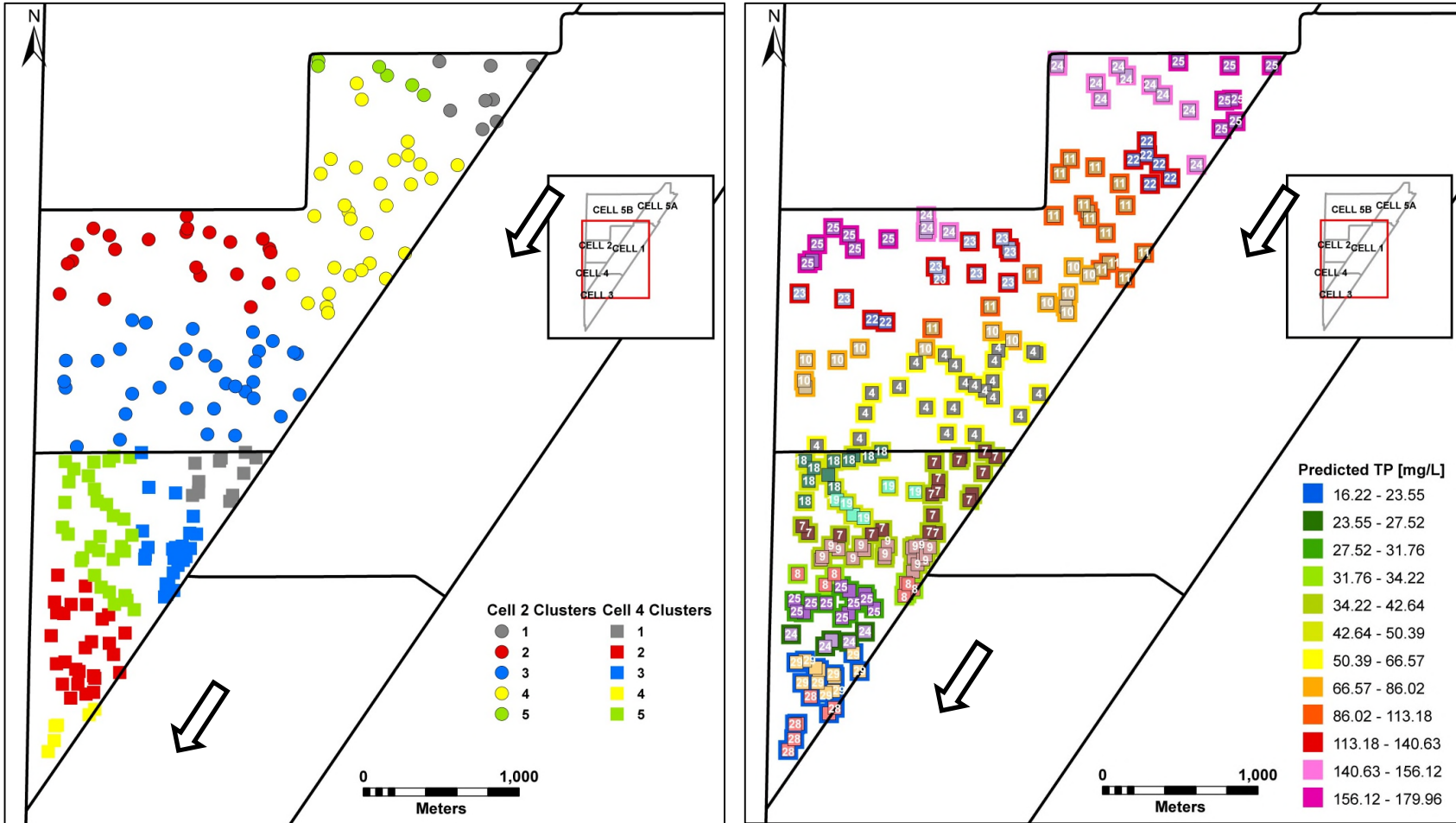
Figure S1: Cluster analysis for STA-1E Central flow-way. Arrows indicate the direction of water flow through the flow-way.



A

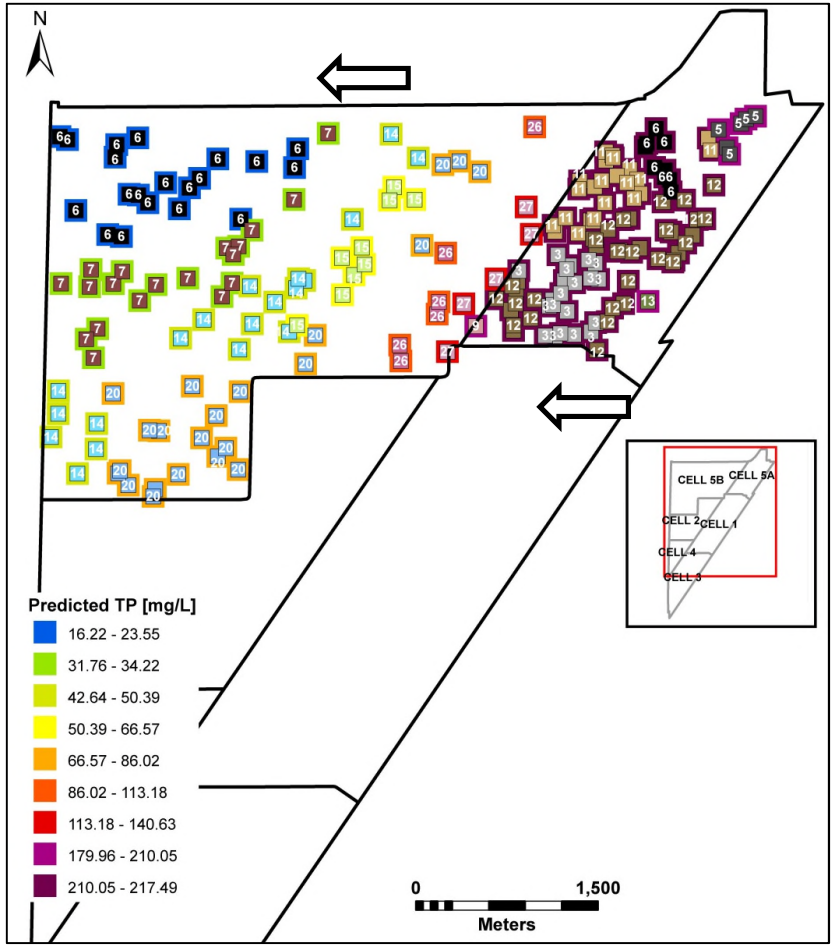
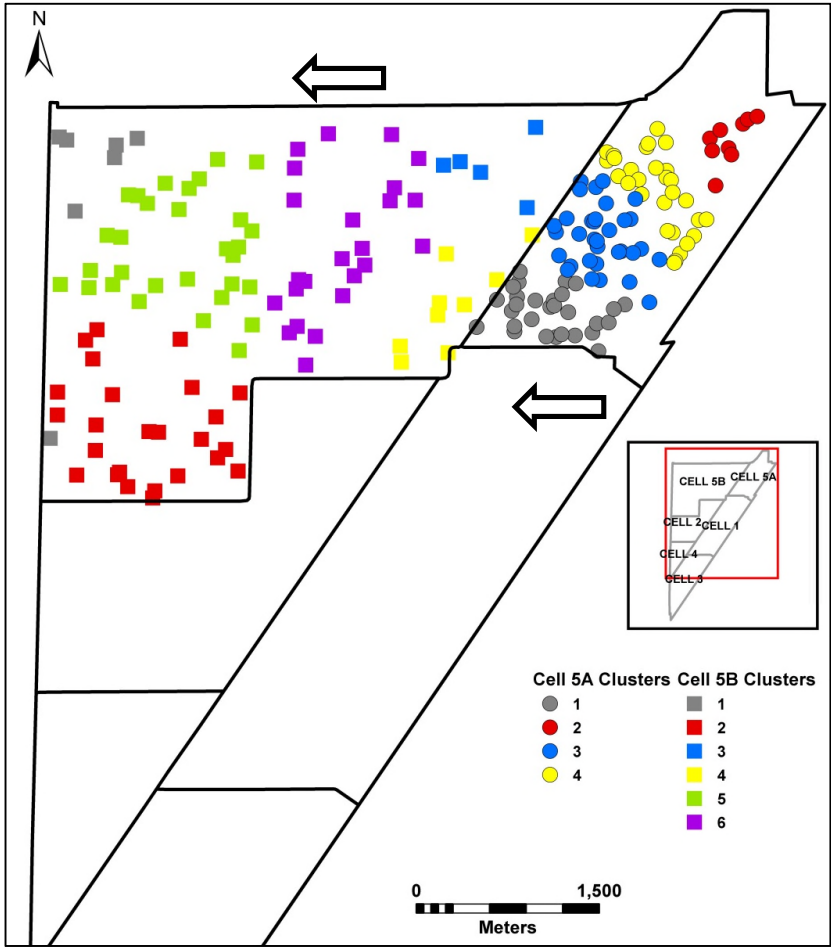
B

Figure S2: Results of A – cluster analysis, and B – CARTs analysis for STA-1W Eastern flow-way. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.



A **B**

Figure S2: Results of A - cluster analysis, and B - CARTs analysis for STA-1W Western flow-way. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.



A

B

Figure S3: Results of A – cluster analysis, and B – CARTs analysis for STA-1W Northern flow-way. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.

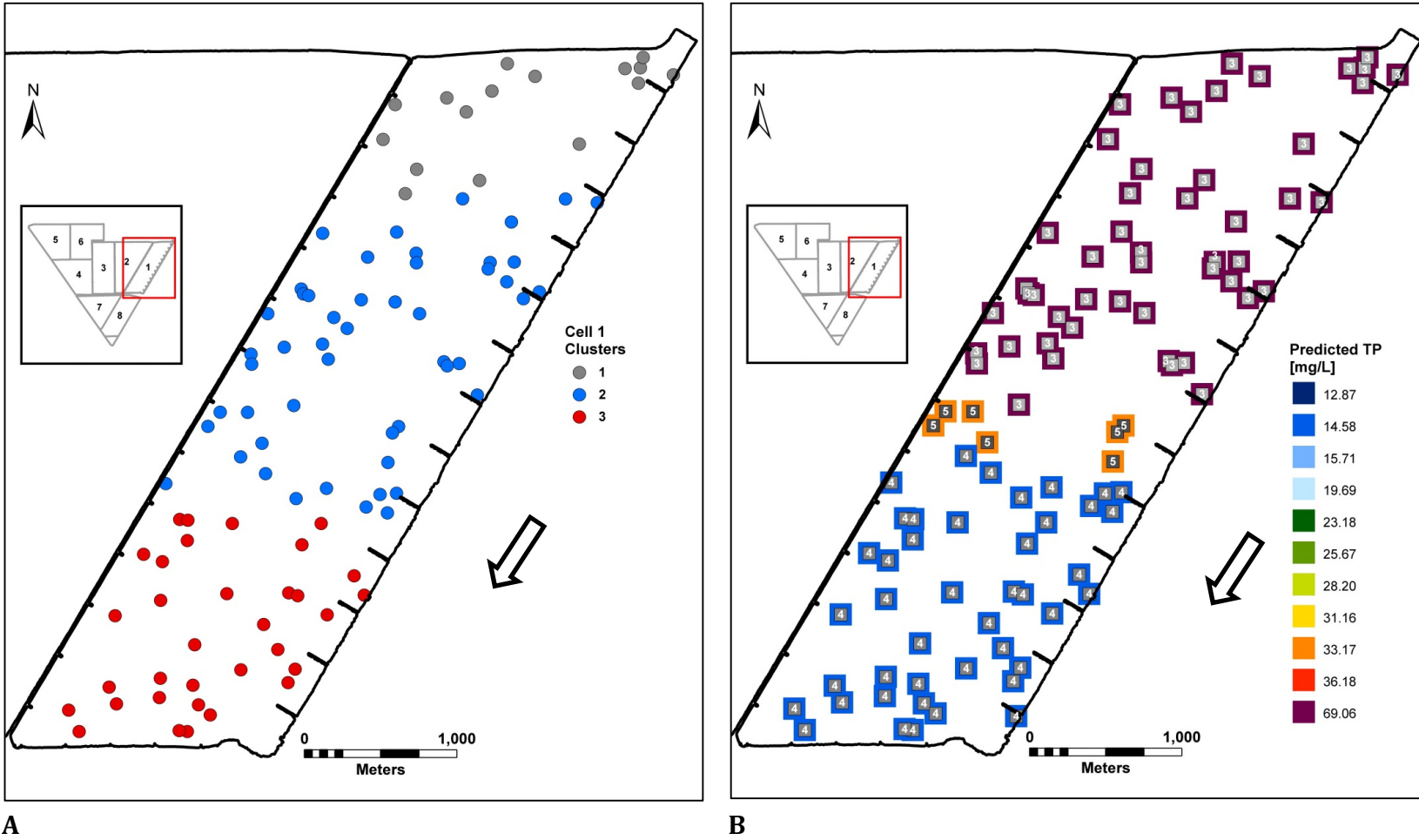


Figure S4: Results of A – cluster analysis, and B – CARTs analysis for STA-2 flow-way 1. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.

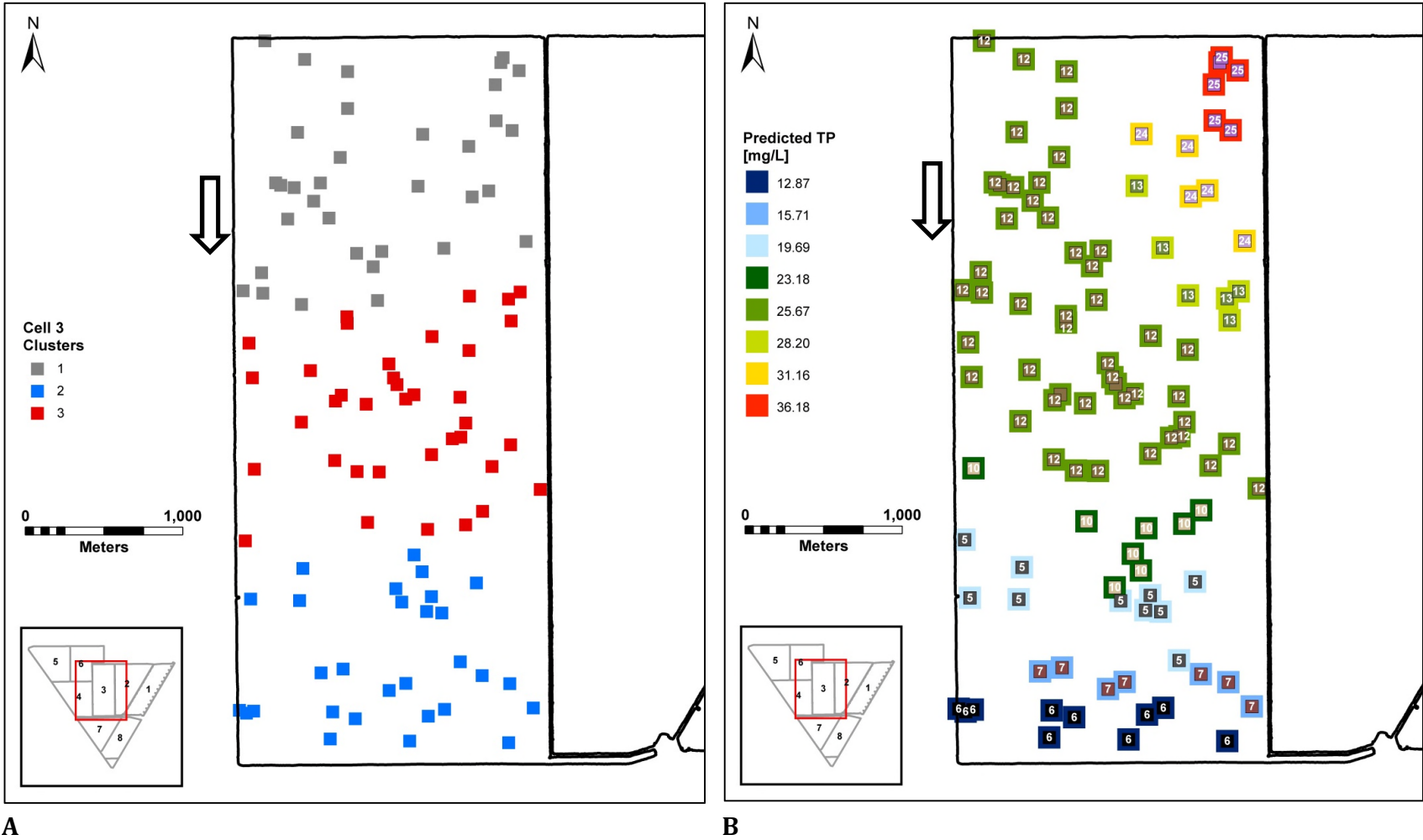


Figure S5: Results of A – cluster analysis, and B – CARTs analysis for STA-2 flow-way 3. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.

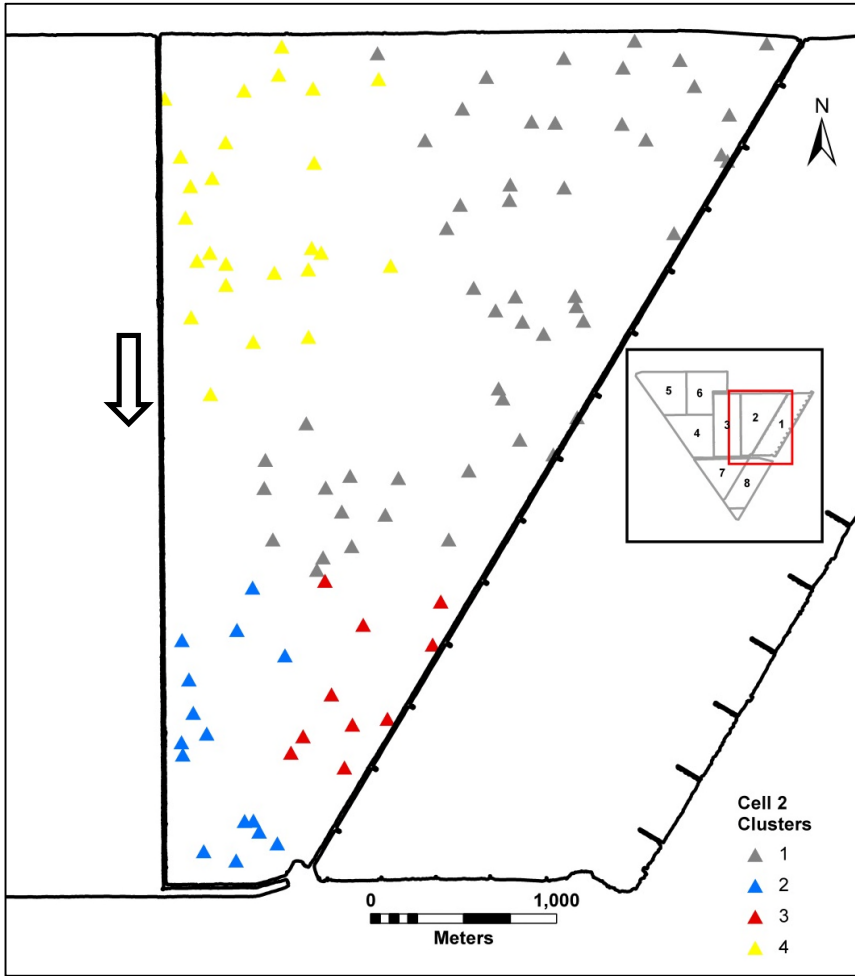
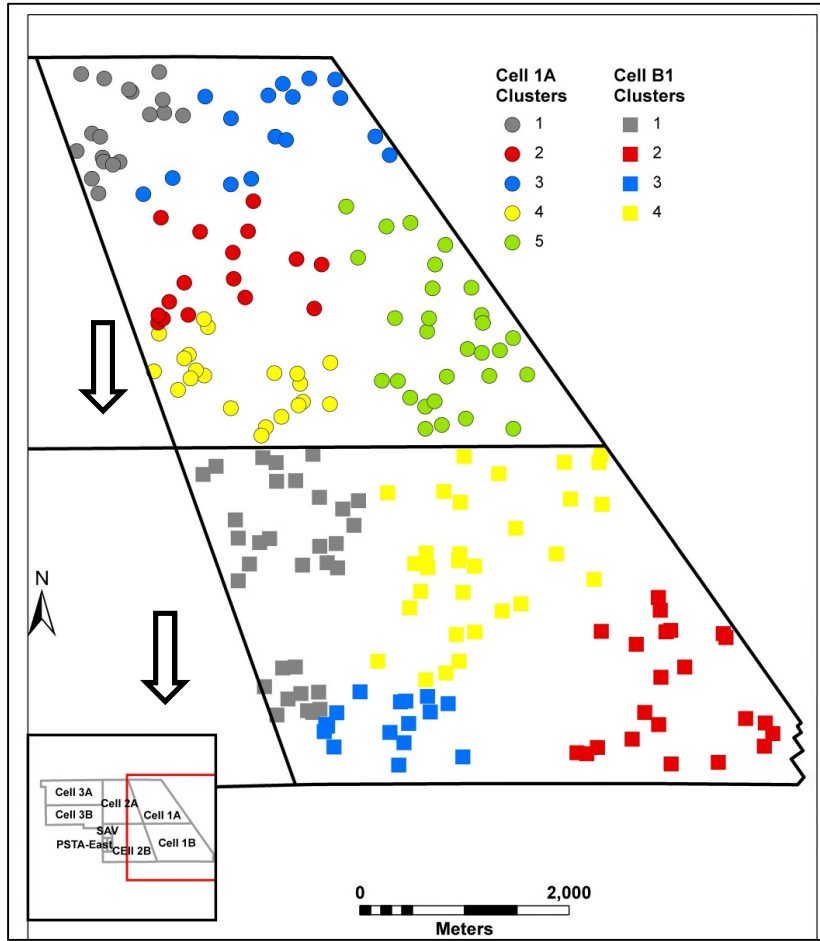
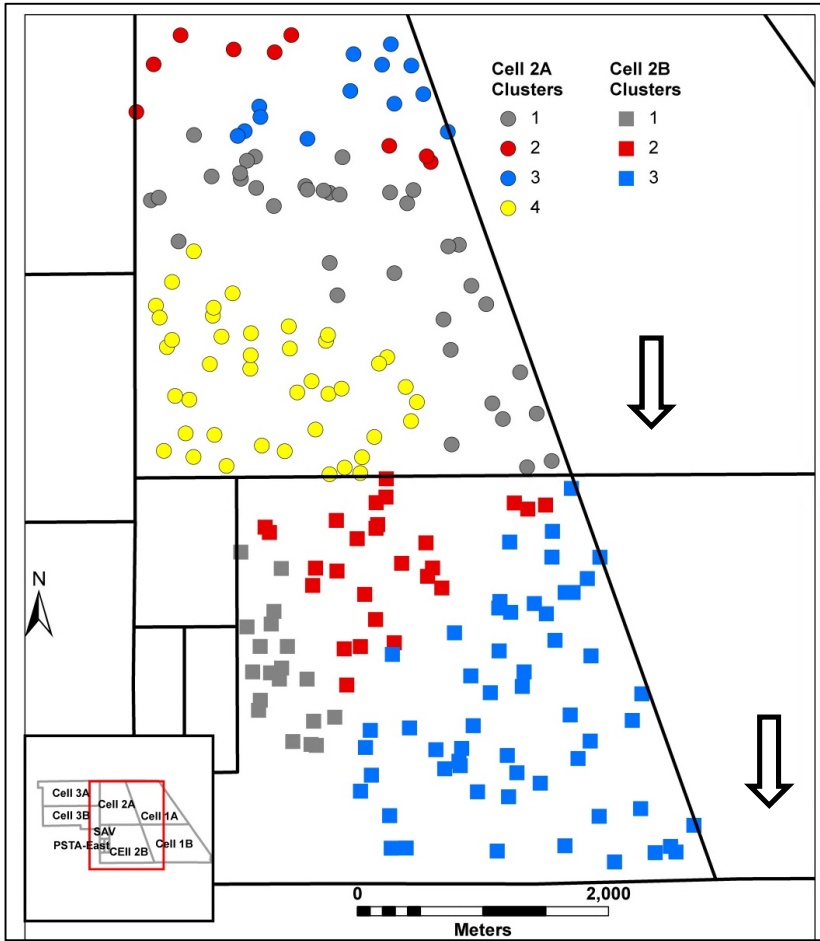


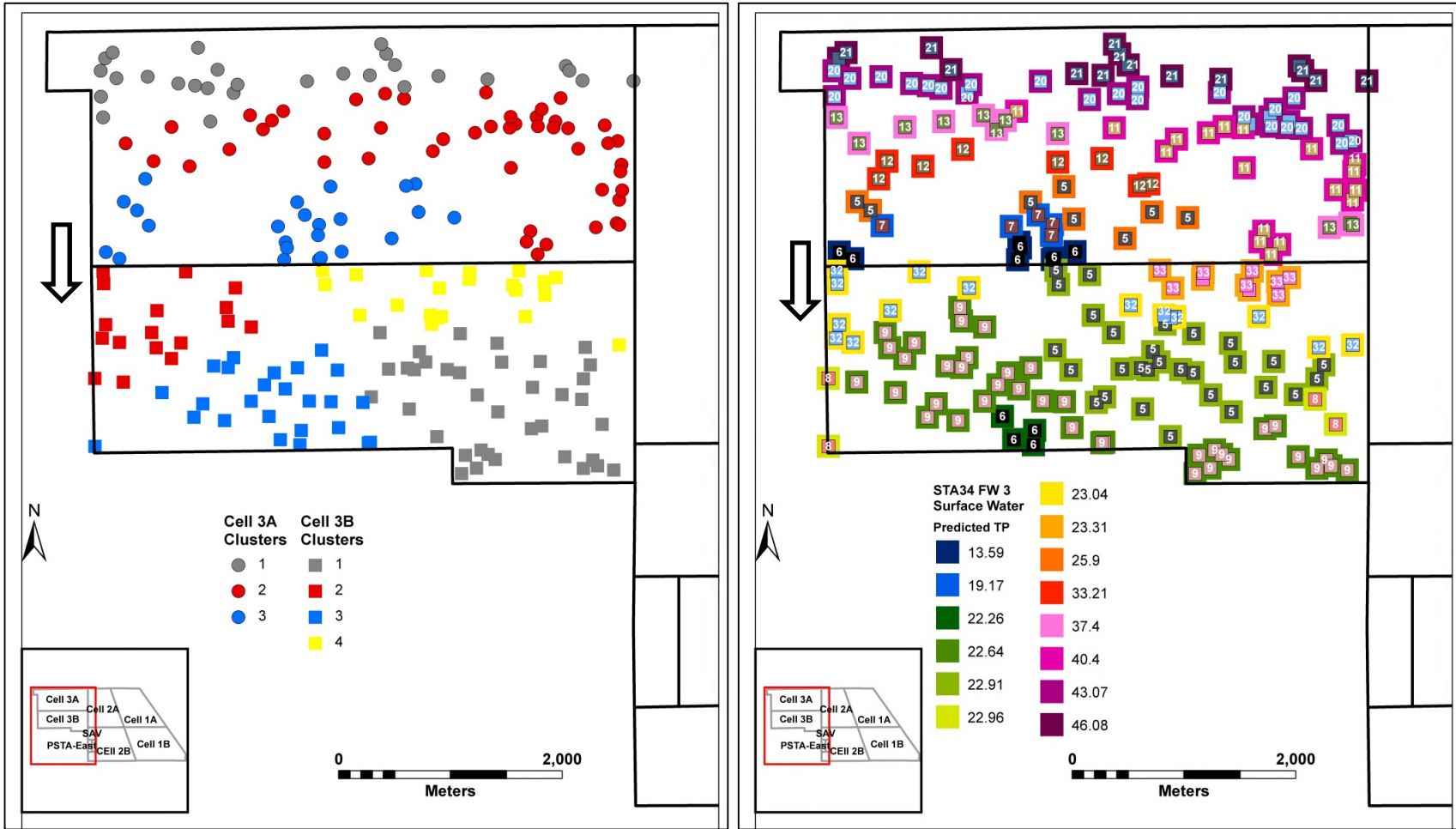
Figure S6: Cluster analysis for STA-2 flow-way 2. Arrows indicate the direction of water flow through the flow-way.



A

B

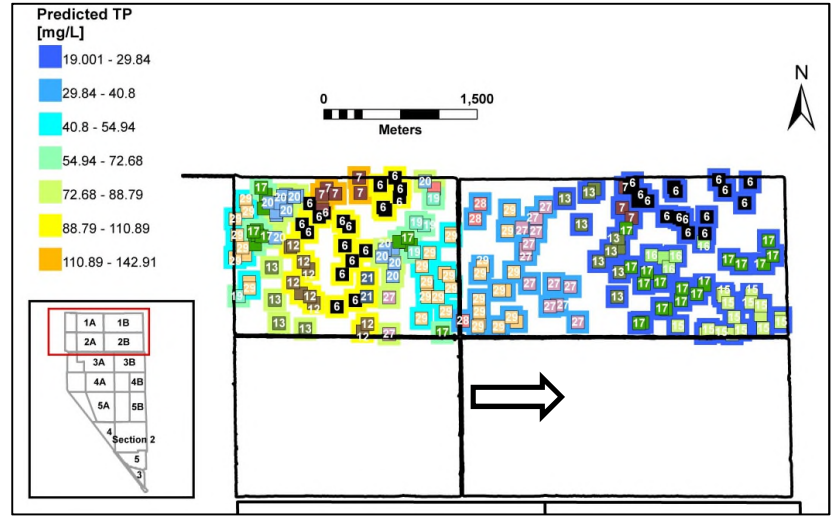
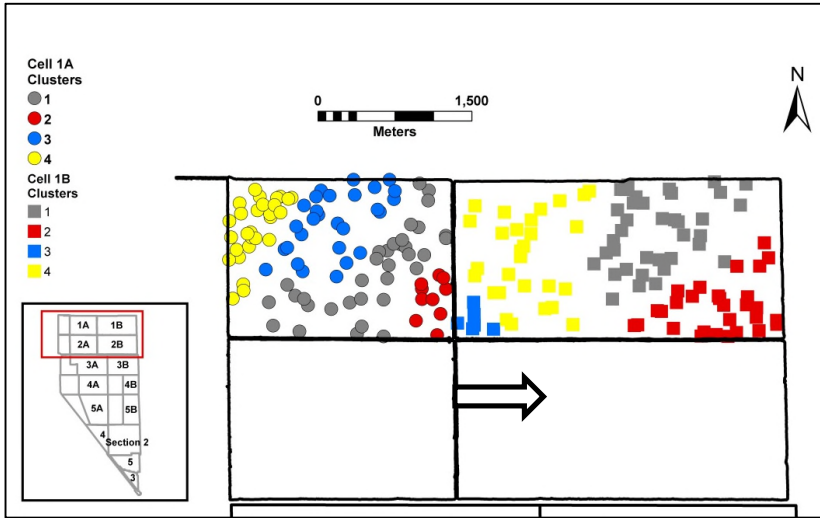
Figure S7: Results of cluster analysis for A – STA-3/4 flow-way 2 and B – STA-3/4 flow-way 1. Arrows indicate the direction of water flow through the flow-way.



A

B

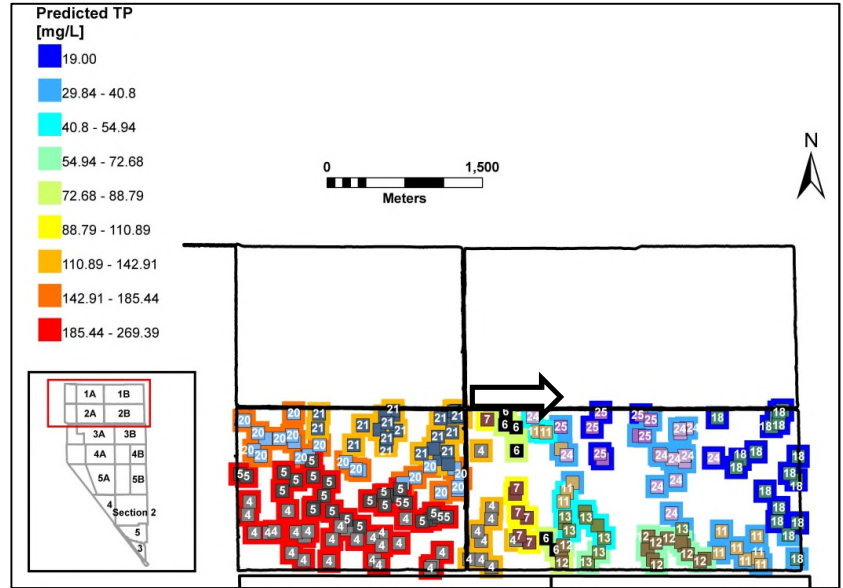
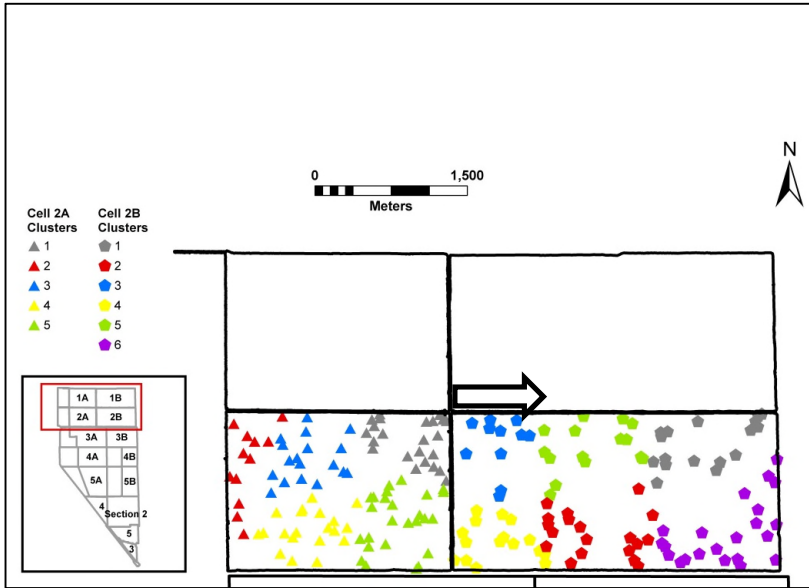
Figure S8: Results of A – cluster analysis, and B – CART's analysis for STA-3/4 flow-way 3. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.



A

B

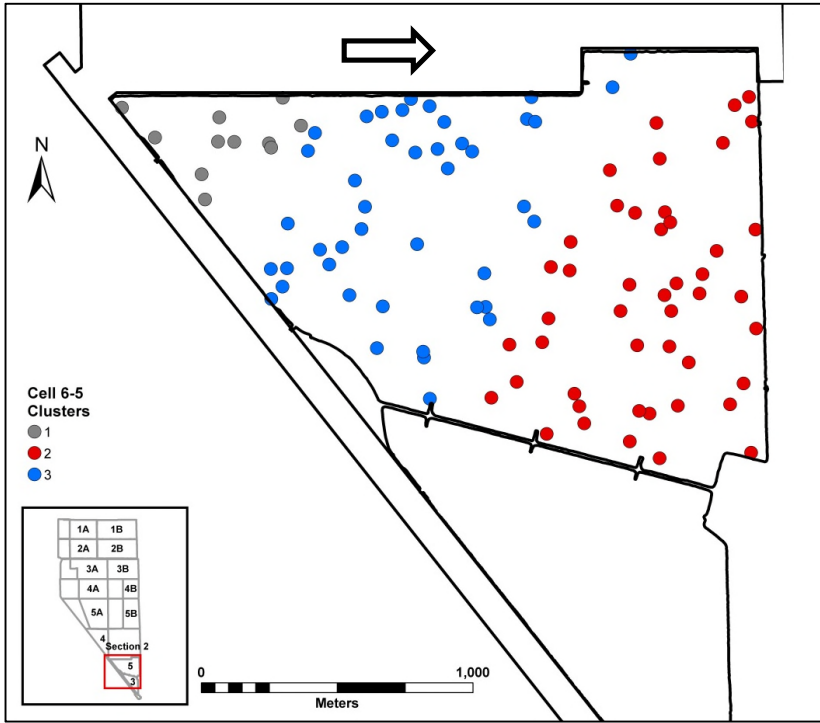
Figure S9: Results of A – cluster analysis, and B – CART's analysis for STA-5/6 flow-way 1. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.



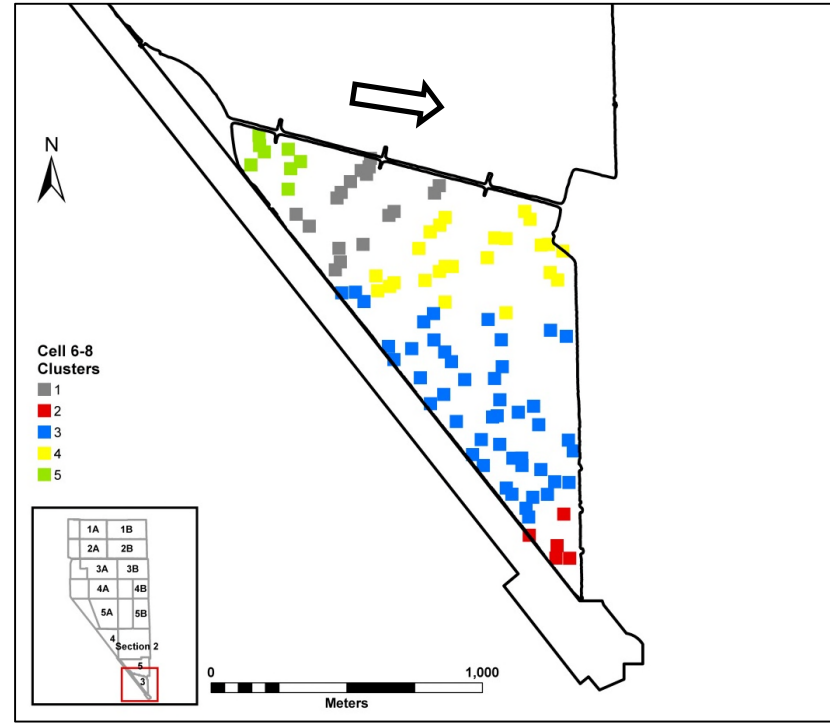
A

B

Figure S11: Results of A – cluster analysis, and B – CARTs analysis for STA-5/6 flow-way 2. Arrows indicate the direction of water flow through the flow-way. Numbers in CART results indicate the number of nodes in the CART model.



A



B

Figure S10: Results of cluster analysis for A – STA-5/6 flow-way 7 and B – STA-5/6 flow-way 8. Arrows indicate the direction of water flow through the flow-way.